



Review

Challenges of using blooms of *Microcystis* spp. in animal feeds: A comprehensive review of nutritional, toxicological and microbial health evaluation

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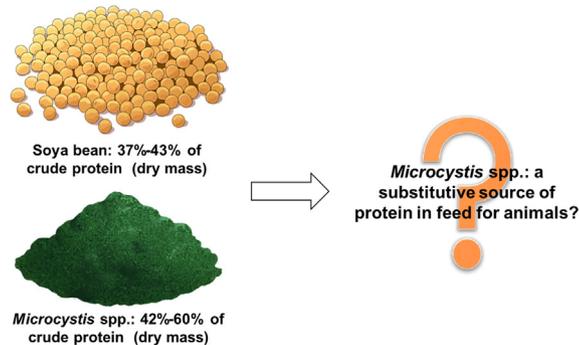
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HIGHLIGHTS

- Challenges of using blooms of *Microcystis* spp. in animal feeds were reviewed.
- *Microcystis* causes toxicity to mollusks, crustaceans, fish, amphibians, mammals and birds.
- *Microcystis* induces toxicity in liver, kidney, intestine, spleen and other organs.
- Fish fed *Microcystis* may be not safe for consumption for humans.
- Microbial pathogens may be present in cyanobacterial blooms.

GRAPHICAL ABSTRACT



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ABSTRACT

Microcystis spp., are Gram-negative, oxygenic, photosynthetic prokaryotes which use solar energy to convert carbon dioxide (CO₂) and minerals into organic compounds and biomass. Eutrophication, rising CO₂ concentrations and global warming are increasing *Microcystis* blooms globally. Due to its high availability and protein content, *Microcystis* biomass has been suggested as a protein source for animal feeds. This would reduce dependency on soybean and other agricultural crops and could make use of “waste” biomass when *Microcystis* scums and blooms are harvested. Besides proteins, *Microcystis* contain further nutrients including lipids, carbohydrates, vitamins and minerals. However, *Microcystis* produce cyanobacterial toxins, including microcystins (MCs) and

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other bioactive metabolites, which present health hazards. In this review, challenges of using *Microcystis* blooms in feeds are identified. First, nutritional and toxicological (nutri-toxicological) data, including toxicity of *Microcystis* to mollusks, crustaceans, fish, amphibians, mammals and birds, is reviewed. Inclusion of *Microcystis* in diets caused greater mortality, lesser growth, cachexia, histopathological changes and oxidative stress in liver, kidney, gill, intestine and spleen of several fish species. Estimated daily intake (EDI) of MCs in muscle of fish fed *Microcystis* might exceed the provisional tolerable daily intake (TDI) for humans, 0.04 µg/kg body mass (bm)/day, as established by the World Health Organization (WHO), and is thus not safe. Muscle of fish fed *M. aeruginosa* is of low nutritional value and exhibits poor palatability/taste. *Microcystis* also causes hepatotoxicity, reproductive toxicity, cardiotoxicity, neurotoxicity and immunotoxicity to mollusks, crustaceans, amphibians, mammals and birds. Microbial pathogens can also occur in blooms of *Microcystis*. Thus, cyanotoxins/xenobiotics/pathogens in *Microcystis* biomass should be removed/degraded/inactivated sufficiently to assure safety for use of the biomass as a primary/main/supplemental ingredient in animal feed. As an ameliorative measure, antidotes/detoxinants can be used to avoid/reduce the toxic effects. Before using *Microcystis* in feed ingredients/supplements, further screening for health protection and cost control is required.

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1. Introduction

Cyanobacteria, commonly referred to as blue-green algae, are Gram-negative, oxygenic, photosynthetic prokaryotes which use energy from sunlight to convert carbon dioxide (CO₂) and nutrients into organic carbon products and biomass (Fogg et al., 1973; Whitton, 2012; Huisman et al., 2018). In recent decades, eutrophication, rising concentrations of CO₂ in the atmosphere and in surface waters, and global warming, have been associated with global increases in the frequency, intensity, duration and geographical spread of blooms of cyanobacteria in susceptible aquatic ecosystems. Cyanobacterial blooms can adversely affect water quality, including increased turbidity, hypoxia and anoxia, and the production of unpleasant taste and odor compounds (Deng et al., 2020). Of particular concern is the production of cyanobacterial toxins (cyanotoxins) by these organisms (van Apeldoorn et al., 2007; Metcalf and Codd, 2012; Chen et al., 2016a; Codd et al., 2017; Janssen, 2019; Svirčev et al., 2019).

To avoid or reduce risks associated with the occurrence of blooms of cyanobacteria, many measures, including mechanical collection of biomass, physical adsorption and removal, use of chemical algacides and biological control, have been-, and continue to be implemented in lakes (Chen et al., 2012, 2017b). Among these measures, in China, mechanical harvesting of cyanobacterial biomass is widely used as an urgent strategy during bloom seasons. After the drinking water crisis in Lake Tai (Ch: *Taihu*) in 2007 in China, over 1000 tons of fresh cyanobacterial biomass is collected daily from May to August every year from Bays of Meiliang Gonghu, by the Wuxi Water Authorities.

Table 1
Chemical composition and nutritional values of *Microcystis* sp. (de la Fuente et al., 1977).

Component	% of <i>Microcystis</i> sp. (dry mass)
Moisture	10.4
Protein (Nitrogen × 6.25)	55.6
Ether extract	1.2
Crude fiber	1.6
Ash	5.2
Nitrogen-free extract	26.0
Ribonucleic acid	7.4
Deoxyribonucleic acid	0.3
Caloric value (kcal)	540.3
Calcium (mg)	1169.1
Phosphorus (mg)	633.4

Table 2
General nutritional composition of conventional feed ingredients, microalgae and cyanobacteria including *Microcystis* spp. (% of dry mass).

Source	Proteins (%)	Carbohydrates (%)	Lipids (%)	References
Soybean	37	30	20	Lum et al., 2013
Corn	10	85	4	Lum et al., 2013
Wheat	14	84	2	Lum et al., 2013
<i>Anabaena cylindrica</i>	43–56	25–30	4–7	Lum et al., 2013
<i>Arthrospira maxima</i>	60–71	13–16	6–7	Lum et al., 2013
<i>Chlorella vulgaris</i>	51–58	12–17	14–22	Lum et al., 2013
<i>Spirogyra</i> sp.	6–20	33–64	11–21	Lum et al., 2013
<i>Synechococcus</i> sp.	73	15	11	Lum et al., 2013
<i>Microcystis aeruginosa</i>	42.14	NA	NA	Boyd, 1973
<i>Microcystis</i> sp.	55.6	NA	NA	de la Fuente et al., 1977
<i>M. aeruginosa</i>	23.5–50.1	NA	13.7–23.4	Piorreck et al., 1984
<i>M. aeruginosa</i>	51–60	NA	NA	de Moor and Scott, 1985
<i>M. aeruginosa</i> NIVA-CYA 228/1	35–72	5–60	NA	Bickel et al., 2000
<i>M. aeruginosa</i> FACHB 41	NA	26–52	NA	Jin et al., 2005
<i>M. aeruginosa</i>	NA	NA	18.48–30.32	Sharathchandra and Rajashekhar, 2011
<i>M. aeruginosa</i> NPCD-1	NA	NA	13.19–28.10	Da Rós et al., 2012
<i>M. aeruginosa</i> 3123	51.3–69.5	NA	NA	Servaites et al., 2012
<i>M. aeruginosa</i> FACHB 469	22–55	NA	NA	Li et al., 2014
<i>M. aeruginosa</i> FACHB-912	NA	38.9–60.5	16–18	Zuo et al., 2018

NA: not analyzed.

However, because water contents of harvested biomass are typically >90%, dewatering, concentration, transportation, and further utilization of blooms are not cost-effective or energy-efficient (Chen et al., 2012).

Among freshwater cyanobacteria, *Microcystis* is one of the most commonly observed genera. While *Microcystis aeruginosa* has been the most commonly reported and investigated species, a range of further morpho-species of *Microcystis* species are recognized, including *M. wesenbergii*, *M. ichthyoblabe*, *M. flos-aquae*, *M. viridis*, *M. botrys*, *M. panniformis*, *M. firma*, *M. natans*, *M. novacekii*, *M. smithii*, *M. bengalensis*, *M. ramosa*, and *M. pseudofilamentosa* (Šejnohová and Maršálek, 2012; Harke et al., 2016; Bernard et al., 2017). Similar to other cyanobacteria, such as *Spirulina platensis* and some microalgae, including *Chlorella vulgaris*, *Microcystis* spp. contain 42%–60% crude protein (dry mass basis), which is also similar to the protein content of soybean (37%–43%) (Boyd, 1973; de la Fuente et al., 1977; de Moor and Scott, 1985; Qiao et al., 2013; Tables 1 and 2, Fig. 1). Therefore, *Microcystis* spp. bloom biomass have been suggested as a source for protein in animal feeds (Fig. 2). This could potentially reduce use of soybean and other agricultural crops and could make use of “waste” biomass when *Microcystis* scums and blooms are harvested (Fig. 3). *Microcystis* spp. contain further substances of potential nutritional value in addition to proteins, including lipids and poly-unsaturated fatty acids (PUFAs), carbohydrates, pigments, antioxidants, vitamins and minerals.

With the ever-increasing growth of human population and great demand for proteins and lipids, sustainable, alternative sources of these nutrients have become a priority due to asymptotic limitations of production of traditional crops under the burden of climate change and energy resource limits (Vanthoor-Koopmans et al., 2014; Colla et al., 2020). Thus, the potential for use of non-toxic *Microcystis* blooms should be highlighted. However, several *Microcystis* species including *M. aeruginosa* are known to produce a diverse family of cyclic heptapeptides, microcystins (MCs) (Harke et al., 2016; Bernard et al., 2017; Svirčev et al., 2019), and MC- and non-MC-producing species/strains of *Microcystis* can occur simultaneously throughout blooms (Rinta-Kanto et al., 2005; Ha et al., 2009; Baxa et al., 2010; Wood et al., 2011; Harke et al., 2016). The general structure of MCs is: cyclo (-D-Ala1-L-X2-D-erythro-β-methylAsp(iso-linkage)3-L-Z4-Adda5-D-Glu(iso-linkage)6-N-methyldehydro-Ala7) (Catherine et al., 2017). More than 279 congeners of MCs have been described (Spoof and Catherine, 2017; Bouaïcha et al., 2019; Du et al., 2019). The L-amino acid residues 2 (X) and 4 (Z) contribute to the major varieties of MCs, and also determine the suffix in the nomenclature of these toxins. For instance, microcystin-LR (MC-LR) includes leucine (L) and arginine

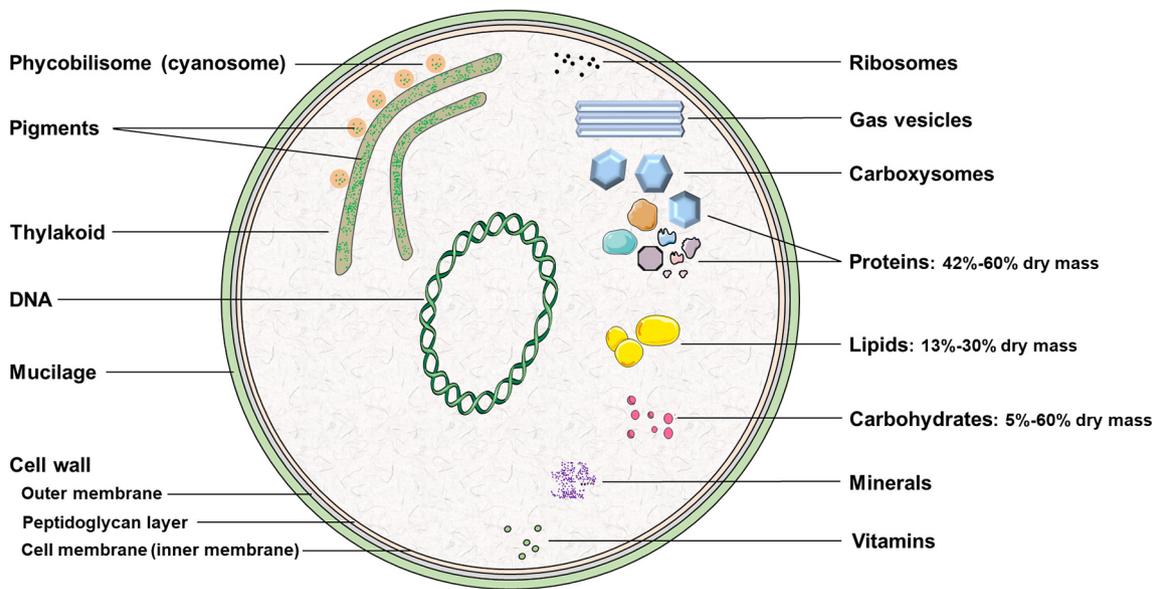


Fig. 1. Schematic figure of *Microcystis* spp. cell illustrating chemical composition and nutritional factors. The chemical composition and the nutrient value of *Microcystis* spp. vary considerably based on species, strains, environmental and growth conditions, the stage of cyanobacterial life cycle, geographic location and season, but *Microcystis* spp. are generally mostly composed of proteins, lipids and carbohydrates based on dry mass.

(R) at positions 2 and 4, respectively. MC-LR is one of the most common and toxic congeners and is also the most widely studied of the MCs (Metcalf and Codd, 2012; Chen and Xie, 2016; Svirčev et al., 2017). MCs have been found to accumulate in liver, kidneys, brain, testes and ovaries of vertebrates, where they can result in adverse outcomes (Chen et al., 2016a; Buratti et al., 2017; McLellan and Manderville, 2017). Based on doses, routes and duration of exposure, toxic effects including hepatotoxicity, nephrotoxicity, neurotoxicity, cardiovascular disease, immunomodulation, endocrine disruption, reproductive and developmental toxicity, and death can occur concomitantly or sequentially in animals exposed to MCs (Metcalf and Codd, 2012; Adamovsky

et al., 2015; Chen et al., 2013, 2014, 2016a,b, 2017a, 2018; Chen and Xie, 2016; Hu et al., 2016; Buratti et al., 2017; Svirčev et al., 2017). In 1998, the World Health Organization (WHO) established a provisional guideline value of 1 µg MC-LR/L in drinking water to protect human health against potential life-time exposure (WHO, 1998; Chorus and Bartram, 1999). In 2010, the International Agency for Research on Cancer (IARC) classified MC-LR as a Group 2B carcinogen, which is “possibly carcinogenic to humans” (IARC, 2010). In addition to MCs, *Microcystis* spp. can also produce other secondary metabolites which cause undesirable tastes and odors (odorous metabolites) or are biochemically active (bioactive metabolites) including cyanotoxins and protease

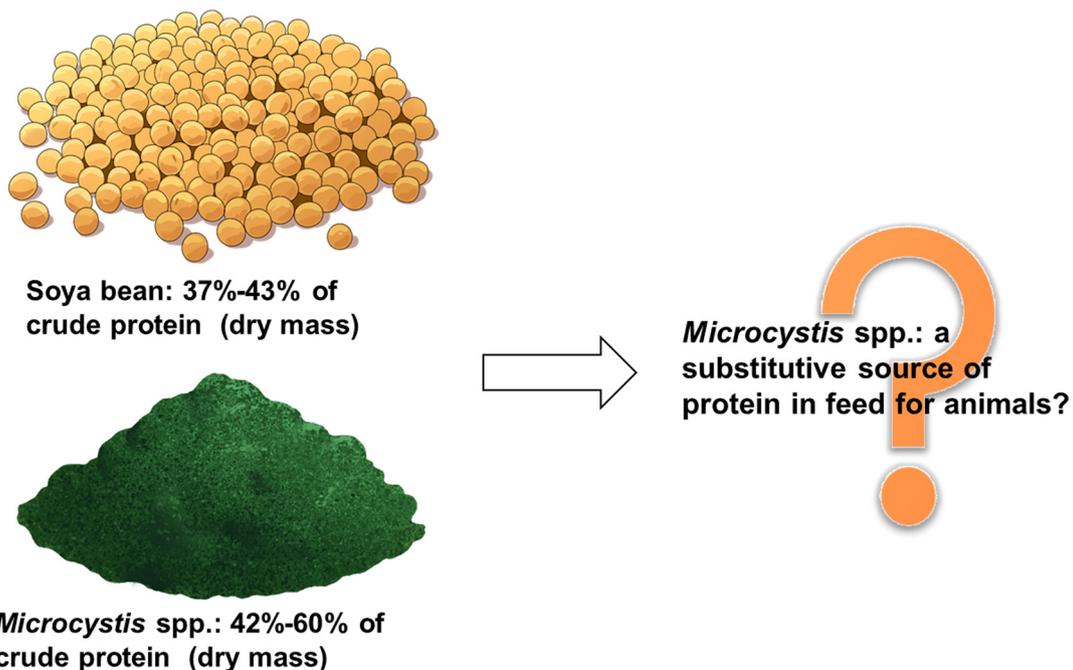


Fig. 2. Potential use of *Microcystis* spp. as a substitutive source of protein in feed for animals. *Microcystis* spp. have been reported to contain 42%–60% crude protein (dry mass), which is close to the protein content of soybean (43%). Therefore, *Microcystis* spp. have been proposed to be used as protein sources for animal feeds.

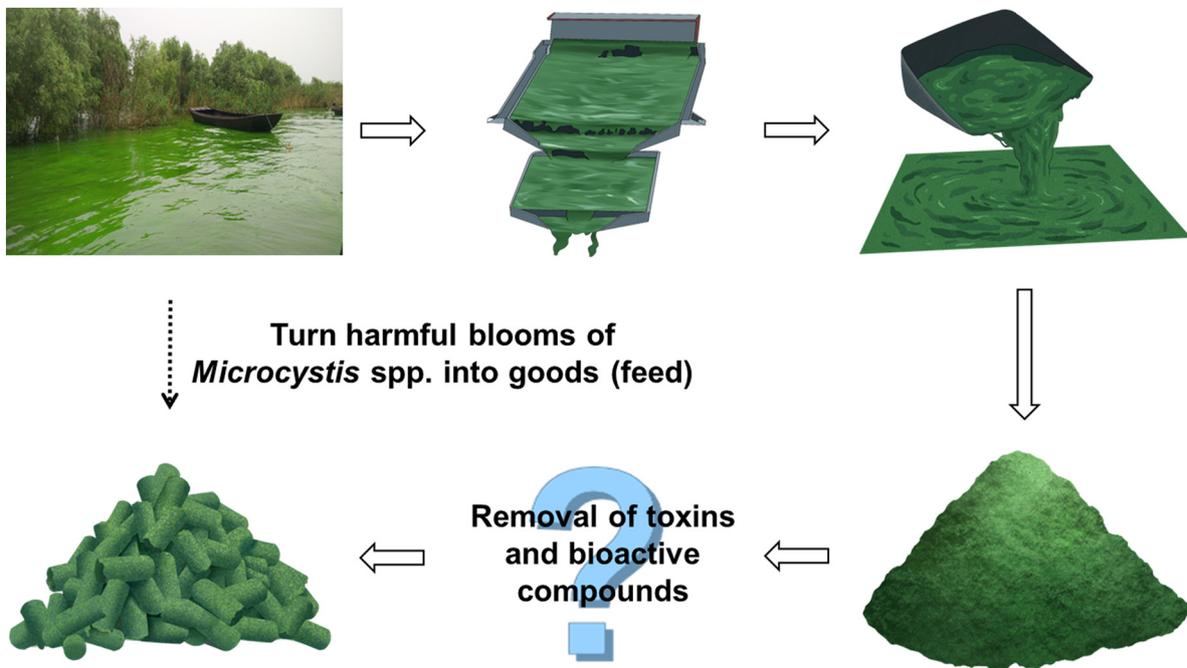


Fig. 3. Schematic procedure to transform biomass of *Microcystis* spp. from blooms to substitutive source of protein in feed for animals, which turn harmful blooms into goods (feed). *Microcystis* biomass is harvested from the environment, dried and powdered. Toxins, bioactive compounds and other harmful ingredients should be removed before the dried biomass is processed into feed.

inhibitors (Smith et al., 2008; Metcalf and Codd, 2012). Therefore, unlike *C. vulgaris*, the presence of MCs and other bioactive products potentially restricts the use of harvested *Microcystis* spp. (Gantar and Svirčev, 2008).

In addition to the presence of MCs, blooms of cyanobacteria and their mucilaginous layers potentially contain microbial pathogens which, if not removed or killed, may limit their use as a source of protein in feed (Deen et al., 2019; Vadde et al., 2019). In this review, challenges of using blooms of *Microcystis* spp. in animal feeds are identified.

2. Chemical composition and nutritional values of *Microcystis*

For *Microcystis* spp. to be considered as a potential food/feed ingredient, one crucial factor is their composition and nutritional content (Torres-Tiji et al., 2020). Chemical compositions vary significantly among *Microcystis* species/strains, and even within the same strains, nutritional content can vary tremendously according to environmental and growth conditions, the stage of cyanobacterial life cycle, geographic location and season (Mišurcová et al., 2014). Important nutrients and components to consider are protein, lipid and carbohydrate content, as well as vitamin, mineral, and pigment content, all of which can positively affect animal health.

2.1. Proteins and amino acids

Protein is an essential macronutrient for animals and is regarded as the limiting nutrient for animal growth. It also accounts for major costs in the production of animal feed stuffs (Kong and Adeola, 2014). The relatively great protein contents of microalgae and cyanobacteria, including *Microcystis* spp., are principal reasons for interest in the use of these organisms as non-conventional sources of protein. Interest in the potential use of microalgae and cyanobacteria as sources of single cell protein (SCP) began as early as the 1950s (Gouveia et al., 2008; Christaki et al., 2011). In general, the crude protein contents of *Microcystis* spp. (for example, about 56% of dry mass) compare favorably with those of some agricultural crops, such as soybean, corn and

wheat, as well as microalgae (Table 2, de la Fuente et al., 1977; Lum et al., 2013).

Nutritional qualities of protein are determined by contents, proportions and availability of their constituent amino acids (Becker, 2007). Based on the Food and Agriculture Organization (FAO) protein reference pattern, the protein score of *Microcystis* sp. was calculated to be 42 (Table 3, de la Fuente et al., 1977). The total proportion of sulfur-containing amino acids is the most significant limiting factor in evaluating the nutritional value of *Microcystis* protein, which is similar to that of other potential sources of SCP, such as microalgae, yeasts and molds. The next most limiting amino acid is lysine (Lys), with a protein score of 81, with 70% of the total lysine content being chemically available to support growth (de la Fuente et al., 1977).

Protein quality is an important criterion used to evaluate the nutritional values of novel ingredients in feeds and this depends on the profile and compositions of the relative amounts of amino acids, their solubilities, susceptibility to chemical and enzymatic hydrolysis in the digestive system, and their physiological utilization after absorption in the gastrointestinal tract (Tibbetts et al., 2015). All of the above factors can be affected by the source of feed ingredients, processing treatments, interactions with other dietary components, presence of anti-nutritional and toxic factors and the feeding habits of target animals. The most accurate method to evaluate quality of proteins is to perform *in vivo* biological assays and calculate parameters including protein efficiency ratio (PER), biological value (BV), digestibility coefficient (DC) and net protein utilization (NPU) (Becker, 2007; Villarruel-López et al., 2017). PER is expressed as gain of body mass per unit protein ingested by animals during short-term feeding experiments. BV is based on the nitrogen (N) retained for growth or maintenance. DC is the proportion of protein digested by animals, compared to that hydrolyzed by chromic oxide (Cr_2O_3), an indigestible marker. Net protein utilization ($\text{NPU} = \text{DC} \times \text{BV}$) is a measure of both digestibility of the protein consumed in food and the biological value of the amino acids. When *Microcystis* sp. was fed to Wistar rats (10.4% of protein content) as the only protein source, little was consumed and mortality was greater than that of the controls, whether or not the diet was supplemented with 0.4% DL-methionine (de la Fuente et al., 1977). However, when *Microcystis* sp.

Table 3

Composition of amino acids of conventional protein sources, microalgae and cyanobacteria including *Microcystis* spp., and the Food and Agriculture Organization (FAO) protein reference pattern (g/100 g protein) (Boyd, 1973; de la Fuente et al., 1977; Becker, 2007).

	WHO/FAO	Egg	Soybean	<i>Chlorella vulgaris</i>	<i>Dunaliella bardawil</i>	<i>Scenedesmus obliquus</i>	<i>Arthrospira maxima</i>	<i>Spirulina platensis</i>	<i>Aphanizomenon</i> sp.	<i>Microcystis aeruginosa</i> (Boyd, 1973)	<i>Microcystis</i> sp. (de la Fuente et al., 1977)
Ile	4.00	6.6	5.3	3.8	4.2	3.6	6	6.7	2.9	5.57	7.55
Leu	7.04	8.8	7.7	8.8	11	7.3	8	9.8	5.2	8.75	8.88
Val	4.96	7.2	5.3	5.5	5.8	6	6.5	7.1	3.2	6.20	7.40
Lys	5.44	5.3	6.4	8.4	7	5.6	4.6	4.8	3.5	5.23	5.41
Phe	6.08 ^a	5.8	5	5	5.8	4.8	4.9	5.3	2.5	4.23	4.19
Tyr		4.2	3.7	3.4	3.7	3.2	3.9	5.3	–	4.00	3.22
Met	3.52 ^b	3.2	1.3	2.2	2.3	1.5	1.4	2.5	0.7	1.73	0.90
Cys		2.3	1.9	1.4	1.2	0.6	0.4	0.9	0.2	0.28	0.92
Try	0.96	1.7	1.4	2.1	0.7	0.3	1.4	0.3	0.7	–	1.51
Thr	4.00	5	4	4.8	5.4	5.1	4.6	6.2	3.3	5.68	4.31
Ala		–	5	7.9	7.3	9	6.8	9.5	4.7	7.88	9.13
Arg		6.2	7.4	6.4	7.3	7.1	6.5	7.3	3.8	9.71	10.06
Asp		11	1.3	9	10.4	8.4	8.6	11.8	4.7	12.95	14.93
Glu		12.6	19	11.6	12.7	10.7	12.6	10.3	7.8	12.81	13.74
Gly		4.2	4.5	5.8	5.5	7.1	4.8	5.7	2.9	4.79	5.25
His		2.4	2.6	2	1.8	2.1	1.8	2.2	0.9	1.40	1.49
Pro		4.2	5.3	4.8	3.3	3.9	3.9	4.2	2.9	3.55	1.71
Ser		6.9	5.8	4.1	4.6	3.8	4.2	5.1	2.9	5.21	2.64

^a Sum of phenylalanine (Phe) and tyrosine (Tyr), according to the FAO protein reference pattern.

^b Sum of methionine (Met) and cysteine (Cys).

was supplied as 25% of the total protein of a corn-*Microcystis* diet (2.05% of protein contributed by *Microcystis* sp.), the PER was significantly improved. These results suggested a possible use of *Microcystis* spp. as a protein supplement to corn, or cereal-based diets for mono-gastric animals.

Using *in vivo* assays with animals to estimate quality of protein in diets is time-consuming and costly, whereas *in vitro* methods, which simulate digestion of protein by including proteolytic enzymes, provide alternative approaches (Tibbetts et al., 2015). Although not fully definitive, *in vitro* assays are widely considered as effective tools to predict potential quality of protein for use in industry and as a nutrient. This is because *in vitro* assays are relatively cheap, they do not use live experimental animals and results are quickly obtained by use of small numbers of samples. They can also complement data on chemical composition and amino acid profiles. *In vitro* digestibility of protein from *Microcystis* sp. was 69.5% (de la Fuente et al., 1977), which is less than that of other cyanobacteria and microalgae (78–94%, Tibbetts et al., 2015). Therefore, efforts should be made to devise practical techniques for improving the digestibility of *Microcystis* spp. protein.

2.2. Lipids and fatty acids

Lipids are essential components in animals and are used as feed additives. Cyanobacteria and microalgae produce various lipids, including glycolipids, phospholipids (polar lipids), glycerolipids, neutral storage lipids and free fatty acids (FFAs) (Villarruel-López et al., 2017). Contents of lipids of *M. aeruginosa* ranged from 13–30% of dry mass (dm), according to different strains and growth conditions (Piorreck et al., 1984; Sharathchandra and Rajashekhar, 2011; Da Rós et al., 2012; Zuo et al., 2018; Table 2).

Among lipids, fatty acids can constitute major constituents of cyanobacterial and microalgal biomass and occur in various forms. They occur as: phospholipids which form a phosphate group at position sn-3; as glycerolipids based on addition of glycerol esterified at positions sn-1 and sn-2; and as non-polar glycerolipids, which can be esterified at 1 (monoacyl-glycerol), 2 (diacyl-glycerol), or 3 positions (triacyl-glycerol, TAG) (Villarruel-López et al., 2017). *M. aeruginosa* contains saturated fatty acids (SFAs), mono-unsaturated fatty acids (MUFAs) and poly-unsaturated fatty acids (PUFAs). Most of the fatty acids are saturated SFAs (49%–81%). Hexadecanoic acids (palmitic

acids, C16:0) account for the highest proportions (24%–76%) (Piorreck et al., 1984; Walsh et al., 1997; Sharathchandra and Rajashekhar, 2011; Da Rós et al., 2012). *Microcystis* spp. also have a relatively high content of PUFAs (6%–44%), including α -linolenic acid (C18:3) (ALA) and γ -linolenic acid (C18:3) (GLA) (Piorreck et al., 1984; Krüger et al., 1995; Walsh et al., 1997; Sharathchandra and Rajashekhar, 2011; Da Rós et al., 2012).

2.3. Carbohydrates

Carbohydrates are the most important sources of energy for cells, tissues and organs. In cyanobacteria and microalgae, carbohydrates are present as mono-saccharides or polymers (Villarruel-López et al., 2017). The most abundant mono-saccharides include glucose, rhamnose, xylose and mannose, and the polymers vary in size (di-, oligo-, and polysaccharides) and composition. In contrast to higher plants, cyanobacteria and microalgae do not contain hemicellulose or lignin. Compositions of carbohydrates vary among strains and versus growth conditions and can range from about 5% to 61% of dm (Bickel et al., 2000; Jin et al., 2005; Zuo et al., 2018; Table 2). For example, in *M. aeruginosa*, compared to normal conditions of nitrate supply (e.g. 17.6 mM NaNO₃), the contents of soluble carbohydrates, mainly mono-saccharides and some oligosaccharides, were significantly greater under non-N conditions (0 mM NaNO₃) (Zuo et al., 2018). The contents of insoluble carbohydrates, mainly structural or stored poly-saccharides, were greater under both low-N (8.8 mM NaNO₃) and non-N conditions (Zuo et al., 2018).

2.4. Pigments

One of the most obvious characteristics of cyanobacteria and microalgae is their pigmentation and color (Gouveia et al., 2008). In general, each genus or species has its own particular combination of pigments and characteristic color. In addition to chlorophylls, which are the primary photosynthetic pigments and impart green coloration, microalgae and cyanobacteria also contain various accessory or secondary pigments, including carotenoids (orange coloration), xanthophylls (yellowish shade) and phycobilins (red or blue coloration) (Villarruel-López et al., 2017). These natural pigments aid in absorption of sunlight, increase efficacy of utilization of energy from light and protect cyanobacteria and algae from the harmful effects of sunrays (Christaki

et al., 2011). Moreover, their function as anti-oxidants in photosynthetic cells including those of higher plants, and the algae and cyanobacteria, exhibits interesting parallels with their protective roles as anti-oxidants in animals and humans (Gouveia et al., 2008).

2.4.1. Chlorophylls

The proportion of chlorophylls in *M. aeruginosa* is 0.21%–0.95% of dm (Piorreck et al., 1984). Chlorophyll-a is the only form of chlorophyll in most cyanobacteria, including *Microcystis* spp. (Gouveia et al., 2008). In addition to their use as colorants for food and pharmaceuticals, chlorophyll derivatives have health-promoting activities (Gouveia et al., 2008). These compounds are used in medicine as anti-inflammatory-, wound healing-, and calcium (Ca) oxalate-controlling agents and for internal deodorization. Chlorophylls can also decrease the risk of colon cancer due to the reduction of toxic and hyper-proliferative effects of heme in dietary red meat (de Vogel et al., 2005; Balder et al., 2006).

2.4.2. Carotenoids

Carotenoids are lipophilic pigments derived from 5-C (carbon) isoprene units which are enzymatically polymerized to form highly conjugated 40-C structures (with up to 15 conjugated double bonds) (Cardozo et al., 2007; Sathasivam et al., 2019). One or both ends of the carbon skeleton can undergo cyclization to form ring β -ionone end groups, which can be substituted at various positions by oxo, hydroxy or epoxy groups, to form various xanthophylls. There are >600 carotenoids in nature and about 50 have pro-vitamin A activity, including α -carotene, β -carotene and β -cryptoxanthin (Gouveia et al., 2008). Carotenoids are contained in cyanobacteria and microalgae at 0.1–0.7% of dm (Christaki et al., 2011; Polyak et al., 2013). Animals and humans are incapable of synthesizing carotenoids and thus must obtain these compounds from the diet (Cardozo et al., 2007). Results of epidemiological studies have demonstrated that diets rich in carotenoids could decrease risks of several diseases, especially those in which free radicals are involved in initiation, including arteriosclerosis, cataracts, multiple sclerosis and cancer (Gouveia et al., 2008).

2.4.3. Phycobiliproteins

Phycobiliproteins are deep-colored, water-soluble accessory pigments, including phycoerythrins (red), phycocyanins (blue) and allophycocyanin (light-blue) (Gouveia et al., 2008). They are formed with proteins linked covalently to tetra-pyrrolic chromophoric prosthetic groups, i.e. phycobilins, including phycourobilin (PUB, yellow), phycoerythrobilin (PEB, red), phycocyanobilin (PCB, blue) and phycobiliviolin (PXB, purple). These pigments are used as highly sensitive fluorescence markers in clinical diagnosis, and as natural colorants in food products and cosmetics. Phycobiliproteins also exhibit many pharmacological properties, such as anti-oxidant, anti-inflammatory, neuro-protective and hepato-protective effects (Gouveia et al., 2008).

2.5. Vitamins

Microalgae and cyanobacteria can also produce a broad spectrum of vitamins, including A, B1, B2, B3 (niacin and niacinamide), B5 (pantothenic acid), B6, B7 (H, or biotin), B9 (M, or folic acid), B12, C and E (Gouveia et al., 2008). Nevertheless, contents of vitamins in *Microcystis* spp. are unknown.

2.6. Minerals

Microalgae and cyanobacteria can also be a source of macrominerals, including sodium (Na), potassium (K), calcium (Ca) and magnesium (Mg), and microminerals, such as iron (Fe), copper (Cu), zinc (Zn) and manganese (Mn) (Christaki et al., 2011). Nevertheless, the mineral content *Microcystis* spp. in a potential nutritional context is unknown.

3. Challenges of using blooms of *Microcystis* spp. in animal feeds (Fig. 4)

3.1. Different sources of *Microcystis* biomass

It should be recognized that the chemical composition, including the nutrient value and MC content of *Microcystis* spp. vary considerably

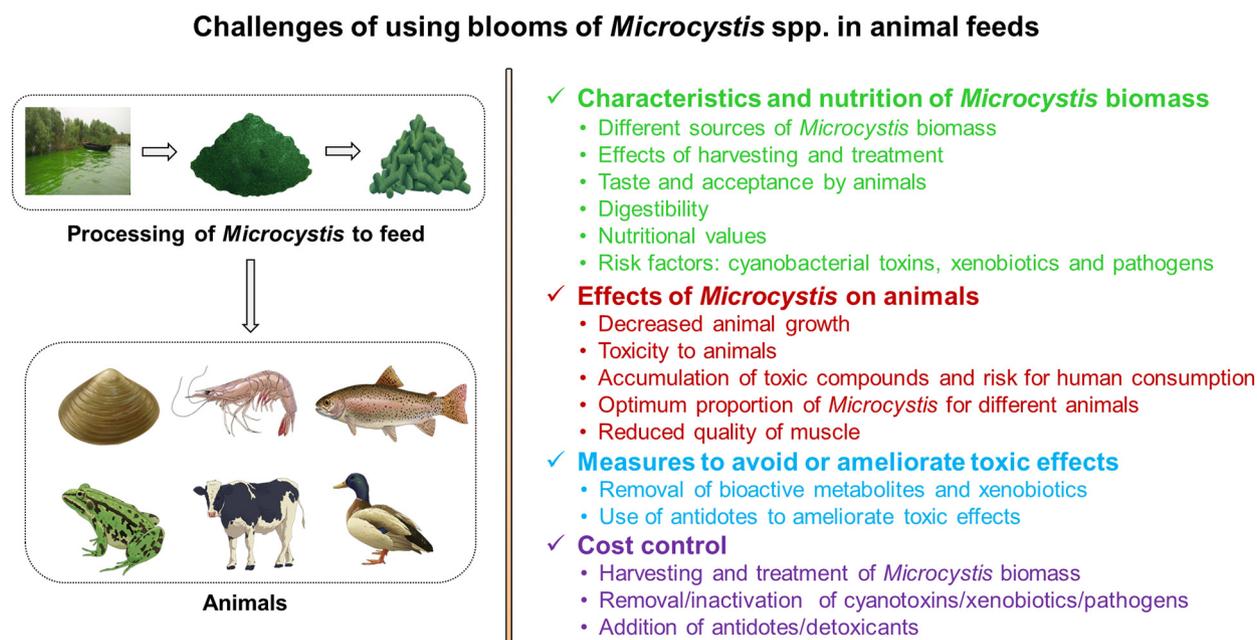


Fig. 4. Overview of the challenges regarding the use of *Microcystis* biomass as a component of animal feed. The potential pitfalls include the variable taxonomical and chemical composition of cyanobacterial biomasses, as well as documented adverse effects on a broad spectrum of animals. The cost-benefit estimation of the use *Microcystis* as primary/main/supplemental ingredient in animal feed is also dependent on a number of expensive actions to harvest and process the *Microcystis* biomass, and ameliorate toxic effects of harmful compounds in the biomass.

based on species, strains, environmental and growth conditions, the stage of cyanobacterial life cycle, geographic location and season (Mišurcová et al., 2014). Within a species, the MC-synthesizing genes may not be always expressed, and both MC-producing and non-MC-producing strains can co-exist (Janssen, 2019). Therefore, methods and/or processes necessary to treat *Microcystis* biomass from different sources may be different. Also, it is necessary to evaluate the optimum ratios of addition of *Microcystis* biomass for animal consumption.

3.2. Harvesting and treatment of *Microcystis* biomass

Research is in progress on the harvesting of *M. flos-aquae* biomass by multi-stage filtration methods (Zang et al., 2020). This is primarily in relation to nutrient removal and bloom control in lakes. The harvesting and processing of *Microcystis* biomass with the aim of producing feed-stuffs, have not been studied. However, the harvesting and processing may affect the chemical compositions of the final products (Mišurcová et al., 2014). The finding of different protein efficiency ratios (PER) for several cyanobacterial and algal species, versus different post-harvest treatments, has revealed the important role of process control of the harvested biomass (Becker, 2007).

3.3. Taste and acceptance of *Microcystis*-containing diets by animals

Microcystis spp. can produce secondary metabolites which cause undesirable tastes and odors (odorous metabolites) (Smith et al., 2008). Palatability of *Microcystis* biomass for terrestrial animals is not known. However, a progressive decrease in cyanobacterial grazing rate by tilapia (*Oreochromis niloticus*) and of opercular beat rates occurred when the proportion of toxic *M. aeruginosa* PCC7820 in their food was increased (Keshavanath et al., 1994).

3.4. Digestibility of *Microcystis*

Poor digestibility of *Microcystis* spp. has been observed in several studies. The in vitro digestibility of protein from *Microcystis* sp. was 69.5% (de la Fuente et al., 1977), which is less than that of other cyanobacteria and microalgae (78–94%, Tibbetts et al., 2015). The mean assimilation efficiency of *Oreochromis mossambicus* fed harvested *M. aeruginosa* bloom material was 51% for total organic matters, 64% for proteins and 76% for phosphorus (P), respectively (de Moor and Scott, 1985). Examination of feces of fishes by transmission electron microscopy (TEM) showed that most *Microcystis* cell walls had become permeable allowing cellular contents to be released. Further digestion caused cell wall degradation. However, up to 25% of the cells appeared to be intact even after passing through the gut. Weak assimilation of radiolabeled *Microcystis* by roach (*Rutilus rutilus*) was detectable, and assimilation rates decreased with increasing proportion of *Microcystis* in a mixture with *Aphanizomenon* (Kamjunke et al., 2002a). Therefore, efforts should be made to improve the digestibility of *Microcystis* spp. protein.

3.5. Nutritional values of *Microcystis*

Use of *Microcystis* biomass in animal nutrition has not yet been discussed in terms of the nutritional value of biomass in the daily diet. Depending on the unknown cyanobacterial biomass vitamin and mineral content, the intake of an uncontrolled amount of *Microcystis* biomass could lead to the detrimental effect of some biomass components due to hypervitaminosis or the accumulation of some minerals and other nutritional ingredients.

In the case of use of *Microcystis* biomass, very specific analyses and precise calculations would have to be performed with respect to all limiting and other significant elements, as a precaution against potential accumulation and overdose. In further calculations of the daily intake it is necessary to take into account how many other important elements

would have been ingested via 100 g of *Microcystis* biomass (proteins, vitamins, minerals, etc.) and how much more would be needed through the traditional diet.

3.6. Effects of *Microcystis* on animal growth

Due to the possibility of poor digestibility of *Microcystis* spp. and possible low/imbalanced nutritional values discussed above, animal growth might be retarded or inhibited (Table 4). Also, *Microcystis* spp. can produce protease inhibitors, including aeruginopeptins, cyanopeptolins, micropeptins, microviridins, aeruginosins and microcins, which can inhibit the digestive proteases, trypsin and/or chymotrypsin (Smith et al., 2008). This would result in incomplete digestion of proteins, decreased nutritional intake and reduced growth rates of target species (Smith et al., 2008). Animal growth may also be decreased as a response to the re-allocation of energy to detoxification of MCs and other cyanotoxins and/or repair processes following intoxication (Smith et al., 2008; Ziková et al., 2010).

3.6.1. Mollusks

Bioenergetics analysis showed that *M. aeruginosa* (CCAP 1450/10 and CCAP 1450/06) resulted in lesser rates of clearance, filtration, ingestion, absorption, and production of feces by zebra mussels, but resulted in a greater rate of production of pseudo-feces and pseudo-diarrhea (Juhel et al., 2006). Net energy balance (NEB), representing the scope for growth (SFG) of mussels feeding on *M. aeruginosa* CCAP 1450/10, was less than that of mussels ingesting non-toxic diets. These results reveal sub-lethal, stressful effects of MCs (particularly MC-LF) on feeding behavior and energy balance of zebra mussels.

3.6.2. Crustaceans

Both *M. aeruginosa* strain IZANCYA2, which produces mostly MC-LR, and strain IZANCYA6, which does not produce MCs, caused a decrease in the growth of juvenile crayfish (Vasconcelos et al., 2001). Acute exposure of post-larvae of white shrimp (*Litopenaeus vannamei*) to *M. aeruginosa* resulted in increased mortalities, stress, lethargy, decreased feeding behavior, soft shells, empty gastrointestinal tract, and whitish stomach and hepatopancreas (Morales-Covarrubias et al., 2016). The moults disappeared and the specific growth rate decreased in freshwater crayfish (*Paranephrops planifrons*) fed with diets containing *Microcystis* sp. CYN06 for 27 days (Clearwater et al., 2014).

3.6.3. Fishes

Oral exposure to *M. aeruginosa* resulted in lesser body mass of *Oreochromis mossambicus* after 15 days, but after 21 days a slight gain in body mass was observed (de Moor and Scott, 1985). 1+ year roach (*Rutilus rutilus*) fed with *Microcystis* showed lesser growth rate, liver glycogen and muscle protein (Kamjunke et al., 2002a,b). Oral exposure to *M. aeruginosa* scum at a dosage of 50 µg/MCs/kg bm/day for 28 days inhibited growth of common carp (Li et al., 2004). Inclusion of natural bloom biomass of cyanobacteria (mainly *M. aeruginosa*) in fish diets, also resulted in greater mortality and feed rate (FR), as well as reduced growth and condition factor (CF) in gibel carp (*Carassius auratus gibelio*) (Zhao et al., 2006a), hybrid tilapia (*Oreochromis niloticus* × *O. aureus*) (Dong et al., 2009), hybrid sturgeon (*Acipenser baeri* × *A. gueldenstaedtii*) (Dong et al., 2011), and yellow catfish (*Pelteobagrus fulvidraco*) (Dong et al., 2012). Greater feed rate (FR), but lesser feed conversion ratio (FCR) or feed conversion efficiency (FCE), specific growth rate (SGR) apparent digestibility coefficient for dry matter/protein/energy (ADCd/ADCp/ADCe), protein retention efficiency (PRE) and energy retention efficiency (ERE) were also observed. However, there were no significant changes in mortality, body mass or total length in Sacramento splittail (Acuña et al., 2012b), and threadfin shad (Acuña et al., 2012a). The lack of such responses in the latter two studies may have been due to the small ratio of added cyanobacteria in the diet

Table 4Effects of oral exposure to *Microcystis* spp. on animal growth and accumulation of microcystins (MCs) in muscle.

Animals	<i>Microcystis</i> spp. characteristics and MC concentrations	Dose	Duration	Effects on mortality and growth	MC Concentrations in muscle (ng/g dm)	References	
Mollusks	Zebra mussel	<i>M. aeruginosa</i> CCAP 1450/10, <i>M. aeruginosa</i> CCAP 1450/06	<i>M. aeruginosa</i> CCAP 1450/10: 2.7×10^6 cell/mL, <i>M. aeruginosa</i> CCAP 1450/06: 1.6×10^6 cell/mL	2 h	Clearance rate↓, filtration rate↓, ingestion rate↓, pseudofaecal production rate↑, absorption rate↓, and faecal production rate↓, net energy balance↓	NA	Juhel et al., 2006
Crustaceans	Crayfish <i>Procambarus clarkii</i> juvenile	<i>M. aeruginosa</i> IZANCYA2, 11.3 µg MC/mg dm; <i>M. aeruginosa</i> IZANCYA6	4% cyanobacteria in diets	8 weeks	Growth↓	NA	Vasconcelos et al., 2001
Fish	<i>Oreochromis mossambicus</i>	<i>M. aeruginosa</i> (95%), collected from Hartbeespoort Dam	–	15, 21 days	bm↓ on day 15, bm↑ on day 21	NA	de Moor and Scott, 1985
	Roach (<i>Rutilus rutilus</i>)	<i>Microcystis</i> sp.	20% fish wm	10 days	Growth rate↓	NA	Kamjunke et al., 2002a,b
	Common carp	<i>M. aeruginosa</i> (91.3%), collected from Naktong River, Korea, total MC (-RR, -LR, -YR, mainly -RR, 73.7%) 357.3 ± 26.8 µg MC/g dm	50 µg MC/kg bm/day	4 weeks	Growth rate↓muscle	38.3 ± 12.3 ng MC-LReq/g wm	Li et al., 2004
	Gibel carp	<i>M. aeruginosa</i> (90%), collected from Lake Dianchi, China; 0.011% MC dm, 110 µg MC/g dm	15.15%, 29.79%, 44.69%, 59.58%, 74.48% cyanobacteria in diets; 39.12, 124.14, 174.5, 203.03, 228.92 ng MC/g diets	12 weeks	Mortality↑, final bm↓, SGR↓, FCE↓, FR↑, ADCd↓, ADCp↓, ADCe↓	0.019, 0.03, 0.062, 0.147, 0.171	Zhao et al., 2006a
	Nile tilapia	<i>M. aeruginosa</i> (90%), collected from Lake Dianchi, China; 0.14% MC dm, 1400 µg MC/g dm	1.19%, 2.34%, 3.51%, 4.68%, 5.85% cyanobacteria in diets; 1253.52, 2210.07, 3363.02, 4440.64, 5460.06 ng MC/g diets	12 weeks	Final bm↑, FR↑	0.77, 1.69, 2.59, 6.47, 14.62	Zhao et al., 2006b
	Hybrid tilapia	50% <i>M. aeruginosa</i> , 50% <i>M. wesenbergii</i> , collected from Lake Taihu, China	43.6% cyanobacteria in diets; 80 µg MC/g diets	30, 60 days	Final bm↓, FR↑, SGR↓, FCE↓, PRE↓, ERE↓, ADCd↓, ADCp↓, ADCe↓	>600	Dong et al., 2009
	Hybrid tilapia	<i>M. aeruginosa</i> (95%), collected from Lake Dianchi, China	50.8% cyanobacteria in diets; 410 µg MC/g diets	30, 60 days	Final bm↓, FR↑, SGR↓, FCE↓, PRE↓, ERE↓, ADCd↓, ADCp↓, ADCe↓	>600	Dong et al., 2009
	Female Nile tilapia	<i>M. aeruginosa</i> (90%)	5%, 20% cyanobacteria in diets; 4.92, 19.54 µg MC-LR/g diets	1, 7, 28 days	SGR↓, FCR↑	<Detection limit of 3 ng/g fresh mass	Ziková et al., 2010
	Hybrid sturgeon	<i>M. aeruginosa</i> (90%), collected from Lake Dianchi, China; 0.81 mg MC/g dm	3%, 10%, 25% cyanobacteria in diets; 26.6, 78.82, 201.03 µg MC/g diets	24, 47 days	Final bm↓, CF↓, FR↓, SGR↓, FCE↓, PRE↑↓, ADCd↓, ADCp↓	>200	Dong et al., 2011
	Yellow catfish	<i>M. aeruginosa</i> (90%), collected from Lake Taihu, China	6.16%, 18.48% cyanobacteria in diets; 32.3, 71.96 µg MC/g diets	30, 60 days	Final bm↓, FR↓, SGR↓	>100	Dong et al., 2012
	Sacramento splittail	<i>Microcystis</i> , collected from San Francisco Estuary, USA, 872 mg MC-LR/kg dm	0.6%, 1.2%, 2.3% cyanobacteria in diets; 3.55, 9.14, 17.13 µg MC-LR/g diets	28 days	No significant changes of mortality, body mass, total length, CF	NA	Acuña et al., 2012b
	Female and male threadfin shad	<i>Microcystis</i> , collected from San Francisco Estuary, USA, 872 mg MC-LR/kg dm	0.573%, 1.15% cyanobacteria in diets; 4.4, 10 µg MC/g diets	57 days	No significant changes of mortality, body mass, total length, SGR, CF↓	NA	Acuña et al., 2012a
	Indian major carp	<i>M. aeruginosa</i> (Kütz)	0.05%, 0.1%, 0.5% cyanobacteria in diets	90 days	SGR↑, FCR↓	NA	Das et al., 2013
	Female goldfish	<i>M. aeruginosa</i> (90%), collected from Lake Dianchi, China, 1.41 mg MC/g dm (-RR, -LR, -YR, 0.84, 0.50, 0.07 mg/g)	10, 20, 30, 40% cyanobacteria in diets	4, 8, 12, 16 weeks	Final bm↑↓, final body length↑, final total length↑↓	> 40	Liang et al., 2015
	Common carp larvae	<i>M. aeruginosa</i> , collected from Lalla Takerkoust reservoir, 976 µg MC-LR/g dm; <i>M. aeruginosa</i> , collected from Lalla Takerkoust reservoir, 968 µg MC-LR/g dm	0.1 g <i>M. aeruginosa</i> /kg dm, 60 ng MC/g dm, 0.06 ng MC/larva/day	12 days	larval mass↓, standard length↓, specific growth rate↓	NA	Ghazali et al., 2010
Amphibians	Bullfrogs larval	<i>Microcystis incerta</i>	–	8 days	dm↓, snout-vent length (SVL)↓, body width↓	NA	Pryor, 2003
	<i>Xenopus laevis</i> tadpoles	Cyanobacterial biomass consisting mainly of <i>M. aeruginosa</i>	10%, 50% <i>M. aeruginosa</i> in diets; 42.8, 187 µg MC-LR/g diets	1, 3, 7, 21 days	bm↓	NA	Ziková et al., 2013

↑, increased; ↓, decreased; bm, body mass; dm, dry mass; wm, wet mass; FR, feed rate; FCR, feed conversion ratio; FCE, feed conversion efficiency; SGR, specific growth rate; CF, condition factor; ADCd/ADCp/ADCE, apparent digestibility coefficient for dry matter/protein/energy; PRE, protein retention efficiency; ERE, energy retention efficiency. NA, not analyzed.

and correspondingly small doses of MCs used (Acuña et al., 2012a,b). Surprisingly, the dietary intake of cyanobacteria promoted growth of Nile tilapia, and the authors inferred that this species may be more tolerant to MCs than other fish species (Zhao et al., 2006b). Increased specific growth rate (SGR) and decreased feed conversion ratio (FCR) were also observed in Indian major carp (*Labeo rohita* Ham.) fed with *M. aeruginosa* bloom material (Kütz) for 90 days (Das et al., 2013). In another study, female Nile tilapia were fed with diets containing *Microcystis* spp. biomass (5%, 20%), and *Arthrospira* sp. (20%) for 28 days (Ziková et al., 2010). For the last week, the fish exposed to diets containing 20% of *Microcystis* spp. showed a slightly reduced rate of growth, with the FCR ranking as: control < *Microcystis*-5% < *Arthrospira*-20% < *Microcystis*-20%. These results suggested that diet highly supplemented with *Microcystis* spp. might inhibit growth by excessive use of energy to enhance hepatic metabolism and detoxification of MC-LR and further metabolites (Ziková et al., 2010). Lesser doses of *M. aeruginosa* promoted the growth of goldfish (*Carassius auratus*), but greater doses inhibited growth (Liang et al., 2015).

Larvae of fishes were also affected by *M. aeruginosa* (Ghazali et al., 2010). A 12-day experiment was performed to compare the effects of two bloom samples of *M. aeruginosa* from Lalla Takerkoust reservoir with different profiles of MCs on the growth of larval common carp. Larvae were fed with diets containing *M. aeruginosa* (0.06 ng MCs/larva/day) for 12 days. Compared with controls, mass and standard length of larvae from the 9th day were lower. Decreased SGR was also observed. Moreover, accumulation of MCs by the larvae was negatively correlated with growth performance between the two cyanobacterial treatments.

3.6.4. Amphibians

Body mass, and body width, measured as the spiracle and snout-vent length (SVL), of larval bullfrogs (*Rana catesbeiana*) fed with *M. incerta* for 8 days were significantly decreased compared to values of tadpoles at the start of feeding trials or those of controls maintained in growth medium containing no cyanobacteria or algae (Pryor, 2003). Similarly, body mass of *Xenopus laevis* fed with field *M. aeruginosa* for 21 days was smaller, compared with controls (Ziková et al., 2013). However, there was a significant increase of body mass of tadpoles of frogs (*Rana grylio*) fed either field *Microcystis* sp. for 3 days, or cultured *M. aeruginosa* NIES-90 for 7 days (Zhang et al., 2012). These inconsistent results could be due to different sensitivities among different species of amphibians.

3.7. Toxicity of *Microcystis* to animals

Unlike *Chlorella vulgaris*, MCs and other bioactive substances produced by *Microcystis* spp. and other cyanobacterial bloom-forming species, can cause toxicity to animals, including mollusks, crustaceans, fish, amphibians, mammals and birds.

3.7.1. Mollusks

Accumulation of MCs in the muscles and significantly greater activities of glutathione S-transferase (GST) in the gills, digestive glands and muscles were observed in *M. galloprovincialis* exposed to *M. aeruginosa* strain M6, which produces MC-FR, -LR and -WR, at 1.5×10^5 cells/mL for 4 days even after depuration for 14 days (Fernandes et al., 2009; Table 5).

Potential of 4 strains of *M. aeruginosa* to cause DNA damage in zebra mussels was studied after exposure for 7, 14, or 21 days (Juhel et al.,

Table 5
Effects of oral exposure to *Microcystis* spp. on mollusks.

Mollusks	<i>Microcystis</i> spp. characteristics and MC concentrations	Dose	Duration	Toxic effects	References
Mussel <i>M. galloprovincialis</i>	<i>M. aeruginosa</i> strain M6	1.5×10^5 cells/mL	4 days	Accumulation of MCs in the muscles, gill GST \uparrow , digestive glands \downarrow	Fernandes et al., 2009
Zebra mussel	<i>M. aeruginosa</i> SAG 17.85, CCAP 1450/10 and CCAP 1450/06, SAG 48.80	10^4 cell/mL	7, 14, 21 days	Haemocytes: no significant changes of cell viability, tail DNA \uparrow	Juhel et al., 2007
Freshwater snail <i>Bellamya aeruginosa</i>	<i>M. aeruginosa</i> FACHB-905	0.5×10^6 and 1×10^6 cells/mL	15 days	Hepatopancreas: accumulation of MCs, ACP \uparrow , ALP \uparrow , GST \uparrow , ultrastructural damage	Zhu et al., 2011
Pearl mussel	<i>M. aeruginosa</i> 905	40–60 μ g MC-LR/L	15 days	Accumulation of MC-LR: hepatopancreas > gonad > gill > muscle, 98 differentially expressed genes in the hepatopancreas	Yang et al., 2012
Freshwater snail <i>Sinotaia histrica</i>	<i>M. ichthyoblabe</i> TAC95, 13.7 μ g MC/mg dm	10^7 cells/mL	8 days	Microstructural changes in hepatopancreas	Xie et al., 2014
Zebra mussel	<i>M. aeruginosa</i> SAG 17.85, CCAP 1450/10, CCAP 1450/06, SAG 48.80	10^4 cell/mL	7, 14, 21 days	Total hemocyte count \downarrow , hyalinocyte percentage \downarrow , granulocyte percentage \uparrow , large basophilic hemocyte percentage \downarrow , phagocytic rate \downarrow , phagocytic index \downarrow , LZM \uparrow	Juhel et al., 2015
Triangle sail mussel	<i>M. aeruginosa</i> FACHB-905	25, 50 mg/L of particulate organic matters	14 days	Soft tissue: accumulation of MCs; haemolymph: GST \uparrow , SOD \uparrow , GPX \uparrow , CAT \downarrow , GSH \uparrow , LZM \uparrow	Hu et al., 2015
Triangle sail mussel	<i>M. aeruginosa</i> FACHB-905	1.8×10^7 cells/mL (50 mg/L)	3, 5, 7, 14, 21 days	Accumulation of MCs in soft tissue, pathological alterations in the gills, digestive diverticula, and stomach	Wu et al., 2017
Female freshwater snail <i>Parafossarulus striatulus</i>	<i>M. aeruginosa</i>	10^7 – 10^8 cells/mL	8 weeks	129 down-regulated and 147 up-regulated unigenes in whole soft bodies	Qiao et al., 2018
Triangle sail mussel	<i>M. aeruginosa</i> FACHB-905	5×10^5 cells/mL (1.4 mg/L), 5×10^6 cells/mL (14 mg/L)	1, 3, 7, and 14 days	Digestive gland: cellulase \uparrow , amylase \uparrow , lipase \uparrow ; stomach: cellulase \uparrow , amylase \uparrow , lipase \uparrow	Gu et al., 2019
Marine mussel <i>Mytilus galloprovincialis</i>	<i>M. aeruginosa</i> LEGE 91094, 0.023 pg MC-LR/cell	10^5 cells/mL	14 days	No significant changes of filtration rate, number of byssal threads, whole body dry mass, total protein or glycogen content, 4 up-regulated and 3 down-regulated proteins in digestive gland	Oliveira et al., 2020

\uparrow , increased; \downarrow , decreased; GST, glutathione S-transferase; ACP, acid phosphatase; ALP, alkaline phosphatase; SOD, superoxide dismutase; GPX, glutathione peroxidase; CAT, catalase; GSH, glutathione; LZM, lysozyme.

2007). Each of three strains (SAG 17.85, CCAP 1450/10 and CCAP 1450/06) produces unique profiles of MC-LF and MC-LR, whereas the 4th strain, SAG 48.80, does not produce MCs. Cell viability analysis showed that the doses of MCs to which the mussels were exposed did not cause cytotoxicity to the haemocytes from the posterior adductor muscle sinus. However, DNA damage in haemocytes, and the percentage of tail DNA (% tDNA) measured by the Comet assay, were increased in mussels fed the 3 MC-producing strains, but not in individuals fed with *M. aeruginosa* strain SAG 48.80. Moreover, the DNA damage appeared to be strain-dependent, and greater concentrations of MCs were associated with greater genotoxicity (Juhel et al., 2007).

Dietary exposure to *M. aeruginosa* also caused immunomodulatory effects in zebra mussels (Juhel et al., 2015). Fewer total hemocytes were observed after 14 days of exposure and thereafter, in mussels exposed to 3 MC-producing strains of *M. aeruginosa*, while there were no obvious changes of total hemocyte counts in mussels fed with non-MC-producing strain SAG 48.80. The ratio of granulocytes to hyalinocytes was increased only on the 7th day in mussels fed the two most toxic strains, which suggested a potential short-term inflammatory response. Phagocytic rate and index were smaller in mussels fed *M. aeruginosa* containing greater concentrations of MCs. Lysozyme (LZM) activities in cell-free hemolymph (CFH) were greater in mussels exposed to MC-producing *M. aeruginosa* compared with MC-free controls on the 14th day but subsequently exhibited a time- and dose-dependent increase (Table 5).

Exposure to *Microcystis* spp. also causes toxicity to freshwater snails. Accumulation of MCs, changes of activities of acid phosphatase (ACP), alkaline phosphatase (ALP) and GST, and ultrastructural damage including to nuclei, mitochondria, rough endoplasmic reticulum and lysosomes, were observed in the hepatopancreas of the freshwater snail (*Bellamya aeruginosa*) exposed to *M. aeruginosa* FACHB-905, which produces MCs at concentrations of 0.5×10^6 and 1×10^6 cells/mL for 15 days (Zhu et al., 2011). After a 15-day exposure to MC-producing *M. aeruginosa* 905, the hepatopancreas of the freshwater pearl mussel (*Hyriopsis cumingii*) contained the greatest MC-LR concentration, followed by gonad, gill and muscle (Yang et al., 2012). Ninety-eight differentially expressed genes involved in cytoskeletal assembly, transcription, signal transduction, cellular metabolism and apoptosis were observed in the hepatopancreas. Following exposure to MC-producing *M. ichthyoblabe* TAC95 for 8 days, the hepatopancreas of the freshwater snail *Sinotia histrica* showed microstructural changes including vacuolization and separation of the basal lamina from cells (Xie et al., 2014). The dose-dependent accumulation of MCs in soft tissue, increased activities of LZM, GST, glutathione peroxidase (GPX) and superoxide dismutase (SOD) in haemolymph, increased glutathione (GSH), and suppressive effects on catalase (CAT) activities, were observed in triangle sail mussels (*Hyriopsis cumingii*) fed MC-producing *M. aeruginosa* FACHB-905 for 14 days (Hu et al., 2015). Exposure to *M. aeruginosa* FACHB-905 for 21 days also resulted in the accumulation of MCs in soft tissue and pathological damage in the gills, digestive diverticula and stomach of triangle sail mussels (Wu et al., 2017). 129 down-regulated and 147 up-regulated unigenes were identified in whole soft bodies of the female freshwater snail (*Parafossarulus*

striatulus) fed cultured *M. aeruginosa* for 8 weeks, compared with those fed the green microalga, *Scenedesmus obliquus* (Qiao et al., 2018). Dose-dependent increased activities of amylase, cellulase and lipase were also observed in the digestive gland and stomach of triangle sail mussels fed *M. aeruginosa* FACHB-905 for 1, 3, 7, and 14 days (Gu et al., 2019). There were no significant changes of filtration rate, number of byssal threads, whole body dry mass, total protein or glycogen content in marine mussel *Mytilus galloprovincialis* fed *M. aeruginosa* LEGE 91094 for 14 days (Oliveira et al., 2020). However, 4 up-regulated and 3 down-regulated proteins were observed in digestive gland.

3.7.2. Crustaceans

Most of the MCs were found in the intestine and hepatopancreas of adult crayfish *Procambarus clarkii* exposed to *M. aeruginosa* IZANCYA2 (Vasconcelos et al., 2001; Table 6). In freshwater crayfish (*Paranephrops planifrons*) fed with diets containing *Microcystis* sp. CYN06 for 27 days (Clearwater et al., 2014), MCs preferentially accumulated in the hepatopancreas and muscle, and this accumulation was greater with increased dietary concentrations. The uptake of MC-WR, a MC congener from the diet was the smallest, with a greater uptake of MC-AR, -LA, -LR, and greatest uptake of MC-RR. In contrast to the positive correlation between MC hydrophobicity and *in vivo* toxicity to the protozoan *Tetrahymena pyriformis* (Ward and Codd, 1999), the uptake of different variants of MCs by *P. planifrons* did not appear to be related to their hydrophobicity.

Male grapsoid (burrowing) crabs (*Neohelice granulata*) were fed commercial rabbit feed for 7 weeks, supplemented with two strains of *M. aeruginosa*, NPDC1, which does not produce MCs, and NPJB, which produces mostly MC-LR (Sabatini et al., 2015). Concentrations of MC-LR in the hepatopancreas increased slightly during the first 3 weeks and subsequently began to decrease until 7 weeks. Contents of thiobarbituric acid-reactive substances (TBARS) were greater in exposed crabs than in controls during the first 3 weeks. GSH contents were less than in controls during the 6th and 7th weeks. Activities of SOD and GST were greater from the 3rd to 7th week in MC-exposed crabs compared to controls. These results suggested that accumulation of MCs and oxidative stress/injury are limited and reversed/recovered by a mechanism of depuration/detoxification based on conjugation with GSH, mediated by GST and activated activities of anti-oxidant and detoxifying enzymes, including SOD (Sabatini et al., 2015).

3.7.3. Fishes

Oral exposure of common carp to *M. aeruginosa* scum (50 µg/MCs/kg, bm/day) for 28 days resulted in significantly greater activities of serum alanine amino-transferase (ALT) and aspartate amino-transferase (AST) (Li et al., 2004; Table 7). Ultrastructural alterations in hepatocytes were also observed, including swelling of the endomembrane system, including mitochondria, endoplasmic reticulum and Golgi body, vacuolization of cytoplasm and accumulation of lipid droplets. Diets containing *Microcystis* spp. induced significant changes of biomarkers for oxidative stress and detoxification in Nile tilapia (*O. niloticus*) (Jos et al., 2005; Puerto et al., 2011), and loach (*Misgurnus mizolepis*) (Li et al., 2005), as shown by increased lipid peroxidation (LPO) and altered activities and expressions of genes/proteins

Table 6
Effects of oral exposure to *Microcystis* spp. on crustaceans.

Crustaceans	<i>Microcystis</i> sp. characteristics and MC concentrations	Dose	Duration	Toxic effects	References
Crayfish <i>Procambarus clarkii</i>	<i>M. aeruginosa</i> IZANCYA2, 11.3 µg MC/mg dm	3% cyanobacteria in diets	2 weeks	Accumulation of MCs in intestine and hepatopancreas	Vasconcelos et al., 2001
Crayfish (<i>Paranephrops planifrons</i>)	<i>Microcystis</i> sp. CYN06, 1400 µg MC/mg dm	6, 13, 33, 100 µg MC/kg dm diets	27 days	Moult↓, specific growth rate↓, accumulation of MCs in hepatopancreas and muscle	Clearwater et al., 2014
Male grapsoid crabs	<i>M. aeruginosa</i> NPJB, 230 µg MC/10 ⁶ cells; <i>M. aeruginosa</i> NPDC1	4×10^5 cells/mL	1, 2, 3, 4, 5, 6, 7 weeks	Hepatopancreas: accumulation of MCs, LPO↑, GSH↓, SOD↑, GST↑	Sabatini et al., 2015

†, increased; ↓, decreased; dm, dry mass; LPO, lipid peroxidation; GSH, glutathione; SOD, superoxide dismutase; GST, glutathione S-transferase.

Table 7
Effects of sub-chronic or chronic exposure to *Microcystis* spp. via oral route on fish.

Fish	<i>Microcystis</i> sp. characteristics and MC concentrations	Dose	Duration	Toxic effects	References
Common carp	<i>M. aeruginosa</i> (91.3%), collected from Naktong River, Korea, total MC (-RR, -LR, -YR, mainly -RR, 73.7%) 357.3 ± 26.8 µg MC/g dm	50 µg MC/kg bm/day	4 weeks	Body color darker, growth rate↓; serum: ALT↑, AST↑; liver: ultrastructural changes; MC contents in organs of fishes: liver 261.0 ± 108.3 ng MC-LReq/g wm, muscle 38.3 ± 12.3 ng MC-LReq/g wm	Li et al., 2004
Male Nile tilapia	<i>M. aeruginosa</i> , collected from River Guadiana, Portugal, 3230 µg MC-LR/g dm	About 1200 µg MC-LR/kg bm/day, crushed <i>M. aeruginosa</i> cells	14 days	No significant changes of LPO, SOD, CAT, GPX, or GR in liver, kidney, gill	Jos et al., 2005
Male Nile tilapia	<i>M. aeruginosa</i> , collected from River Guadiana, Portugal, 3340 µg MC-LR/g dm	About 1200 µg MC-LR/kg bm/day, crushed <i>M. aeruginosa</i> cells	21 days	Liver: LPO↑, SOD↑, CAT↑, GPX↑, GR↑; kidney: LPO↑, CAT↑, GPX↑, GR↑; gill: LPO↑, GPX↓	Jos et al., 2005
Male Nile tilapia	<i>M. aeruginosa</i> , collected from River Guadiana, Portugal, 3340 µg MC-LR/g dm	About 1200 µg MC-LR/kg bm/day, non-crushed <i>M. aeruginosa</i> cells	21 days	Liver: SOD↑, CAT↑, GPX↑, GR↑; kidney: LPO↑, CAT↑, GPX↑, GR↑; gill: SOD↑, GPX↓	Jos et al., 2005
Chinese muddy loach	<i>Microcystis</i> , collected from Naktong River, Korea, 133 µg MC-RR/g dm	75 mg <i>Microcystis</i> kg bm/day (10 µg MC-RR/kg bm/day)	4 weeks	Liver: SOD↑, CAT↑, GPX↑	Li et al., 2005
Male tilapia	<i>M. aeruginosa</i> , collected from River Guadiana, Portugal, 3230 µg MC-LR/g dm	About 1200 µg MC-LR/kg bm/day, crushed <i>M. aeruginosa</i> cells	14 days	Liver: micro- and ultra-structural changes, ACP↑; kidney: micro- and ultra-structural changes, ACP↑, ALP↑; gill: protein↓	Molina et al., 2005
Male tilapia	<i>M. aeruginosa</i> , collected from River Guadiana, Portugal, 2647 µg MC-LR/g dm	About 1200 µg MC-LR/kg bm/day, crushed <i>M. aeruginosa</i> cells	21 days	Liver: micro- and ultrastructural changes, protein↓, ACP↑, ALP↑; kidney: micro- and ultra-structural changes, ACP↑, ALP↑; gill: microstructural changes, protein↓; intestine: microstructural changes	Molina et al., 2005
Male tilapia	<i>M. aeruginosa</i> , collected from River Guadiana, Portugal, 2647 µg MC-LR/g dm	About 1200 µg MC-LR/kg bm/day, non-crushed <i>M. aeruginosa</i> cells	21 days	Liver: micro- and ultra-structural changes, protein↓, ACP↑, ALP↑; kidney: micro- and ultra-structural changes, ACP↑, ALP↑; gill: microstructural changes, protein↓; intestine: microstructural changes	Molina et al., 2005
Female Nile tilapia	<i>M. aeruginosa</i> (90%)	5%, 20% cyanobacteria in diets; 4.92, 19.54 µg MC-LR/g dm diets	1, 7, 28 days	Plasma: no significant changes of cortisol, glucose↑; liver: MCs accumulation, glycogen↓, no significant changes of GH and IGF-I mRNAs; gill and muscle: < detection limit of 3 ng/g fresh mass	Ziková et al., 2010
Male Nile tilapia	<i>M. aeruginosa</i> PCC7806, cultured in laboratory, 742 µg MC-LR/g dm	About 1200 µg MC-LR/kg bm/day, crushed <i>M. aeruginosa</i> cells	21 days	Liver: LPO↑, GPX↑, GST↑, GPX mRNA↑, GST mRNA↑, GST protein↑; kidney: LPO↑, GPX↓, GST mRNA↓, GST protein↓	Puerto et al., 2011
Sacramento splittail	<i>Microcystis</i> , collected from San Francisco Estuary, USA, 872 mg MC-LR/kg dm	6, 12, 23 g <i>Microcystis</i> /kg diets (120, 240, 460 mg <i>Microcystis</i> /kg bm/day); 3.55, 9.14, 17.13 mg MC/kg diets (71, 182.8, 342.6 µg MC/kg bm/day)	28 days	Liver: PP2A↓, microstructural changes; muscle: RNA/DNA↓	Acuña et al., 2012b
Female and male threadfin shad	<i>Microcystis</i> , collected from San Francisco Estuary, USA, 872 mg/kg MC-LR dm	5.73, 11.5 g <i>Microcystis</i> /kg diets (171.9, 345 mg <i>Microcystis</i> /kg bm/day); 4.4, 10 mg MC/kg in diet (132, 300 µg MC/kg bm/day)	57 days	Loss of muscle mass (cachexia), Fulton's condition factor↓, microstructural changes of liver in both females and males and ovary in females, accumulation of MCs in gut, liver and kidney	Acuña et al., 2012a
Male Nile tilapia	<i>M. aeruginosa</i> , 65.5 µg MC-LR/g dm	8, 16, 24 (×10 ⁴ cells/mL)	7 days	After depuration for 30 days, liver: GSH↑, LDH↑, SOD↑, MDA↑, DNA fragmentation↑, DNA-protein crosslinks↑	Khairy et al., 2012
Crucian carp	<i>Microcystis</i> , collected from Lake Dianchi, China, 1.41 mg MC/g dm (-RR, -LR, -YR: 0.84, 0.50, 0.07 mg/g)	20%, 40% cyanobacteria in diets	30 days	Blood: nitroblue tetrazolium activity↑↓, LZM activity↑↓, macrophage bactericidal activity↑; head kidney: relative mass↑, micro- and ultra-structural changes; spleen: relative mass↑, micro- and ultra-structural changes	Qiao et al., 2013
Indian major carp	<i>M. aeruginosa</i> (Kütz)	0.5, 1, 5 g <i>M. aeruginosa</i> /kg dm diets	30, 60, 90 days	Serum: total protein↑↓, albumin↑↓, globulin↑↓, albumin:globulin (A:G) ratio↑↓, LZM↑↓, bactericidal activity↑↓; blood: superoxide anion production↑↓	Das et al., 2013
Nile tilapia	<i>M. aeruginosa</i> , collected from Thenneri Lake, India, 245.33 µg/g MC-LR dm	0.15, 0.30 g <i>M. aeruginosa</i> /fish/day (about 600, 1200 µg MC-LR/kg bm/day)	14 days	Swim erratically, fins hard, body pigmentation faded, whole body GST↑; liver: microstructural changes, protein↓, LPO↑, GST↑; gill: microstructural changes, GST↑; intestine: microstructural changes	Preeti et al., 2016
Nile tilapia	<i>M. aeruginosa</i> , collected from Mariout Lake, Egypt, 3500 µg/g MC-LR dm	About 1200 µg MC-LR/kg bm/day	21 days	Muscle: microstructural changes; serum: ALT↑, AST↑; liver: microstructural changes, MDA, GR, GPX, CAT; gill: microstructural changes	Abdel-Latif and Khashaba, 2017
Blunt snout bream	<i>M. aeruginosa</i> (90%), collected from Lake Dianchi, China, 1.41 mg MC/g dm (-RR, -LR, -YR: 0.84, 0.50, 0.07 mg/g)	15%, 30% cyanobacteria in diets	30 days	Blood: white blood cells↓, phagocytosis activity↑↓, respiratory burst activity↑↓; head kidney: ultrastructural changes, slgM mRNA↑↓, mlgD mRNA↓, slgZ mRNA↓; spleen: slgM mRNA↑↓, mlgD mRNA↑↓, slgZ mRNA↑↓; MC contents in organs of fishes exposed to 30% cyanobacteria: head kidney 7.056 µg/g dm, liver 2.034 µg/g dm, intestine 1.933 µg/g dm, gonad 1.684 µg/g dm, spleen 1.254 µg/g dm	Xia et al., 2018

↑, increased; ↓, decreased; bm, body mass; dm, dry mass; wm, wet mass; PP, protein phosphatase; GH, growth hormone; IGF-I, insulin-like growth factor-I; LZM, lysozyme; LDH, lactate dehydrogenase; ACP, acid phosphatase; ALP, alkaline phosphatase; ALT, alanine amino-transferase; AST, aspartate amino-transferase; LPO, lipid peroxidation; MDA, malondialdehyde; CAT, catalase; SOD, superoxide dismutase; GSH, glutathione; GR, glutathione reductase; GPX, glutathione peroxidase; GST, glutathione S-transferase; mlgD, membrane-bound immunoglobulin class D; slgM/slgZ; secretory immunoglobulin class M/Z.

of antioxidant enzymes, including CAT, SOD, GR, GPX and GST. A significant decrease of Fulton's condition factor, loss of muscle mass (cachexia), micro- and ultra-structural changes in liver, kidney, gill, intestine, and ovary, greater activities of ACP, ALP, ALT and AST, lower activities of protein phosphatase 2A (PP2A), reduced protein contents, and a smaller ratio of RNA/DNA concentration in muscle, were also observed in Nile tilapia (Molina et al., 2005; Preeti et al., 2016; Abdel-Latif and Khashaba, 2017), Sacramento splittail (*Pogonichthys macrolepidotus*) (Acuña et al., 2012b), and threadfin shad (*Dorosoma petenense*) (Acuña et al., 2012a) exposed to *Microcystis* spp. via the diet. The inclusion of natural bloom biomass of cyanobacteria (mainly *M. aeruginosa*) in fish diets, also resulted in greater mortality, as well as reduced hepatosomatic index (HSI), hepatic ALT activities, and histopathological changes including vacuolization of hepatocytes in gibel carp (*Carassius auratus gibelio*) (Zhao et al., 2006a), hybrid tilapia (*Oreochromis niloticus* × *O. aureus*) (Dong et al., 2009), hybrid sturgeon (*Acipenser baeri* × *A. gueldenstaedtii*) (Dong et al., 2011), and yellow catfish (*Pelteobagrus fulvidraco*) (Dong et al., 2012). Even after periods of depuration for 30 days, dose-dependent greater activities of LDH and SOD, greater concentrations of GSH and malondialdehyde (MDA), increased DNA fragmentation and DNA-protein crosslinks were observed in male Nile tilapia exposed to cultured *M. aeruginosa* for 7 days (Khairy et al., 2012).

Fish also exhibit various immune responses after oral intake of *M. aeruginosa*. Lesser doses of *M. aeruginosa* resulted in greater immunity of crucian carp (*Carassius auratus*) and blunt snout bream (*Megalabrama amblycephala*), but greater doses were found to diminish immune function responses (Qiao et al., 2013; Xia et al., 2018). This discrepancy was attributed to hormetic responses (hormesis) to MCs (Qiao et al., 2013). Lesser doses of MC-producing *M. aeruginosa* promoted the growth of goldfish (*Carassius auratus*), but greater doses inhibited growth (Liang et al., 2015). In another study, altered concentrations of serum total protein, albumin, globulin, changes in albumin:globulin (A:G) ratio, and lysozyme activity, bactericidal activity, and blood superoxide anion production, were also observed in Indian major carp fed with field *M. aeruginosa* (Kütz) for 90 days (Das et al., 2013). However, compared to controls, higher survival percentages were observed after intraperitoneal infection of a lethal dose of *Aeromonas hydrophila*.

Phytoplanktivorous fishes, including silver carp, big-head carp (*Hypophthalmichthys nobilis*, previously: *Aristichthys nobilis*) and tilapia (*Oreochromis niloticus*), are direct consumers of phytoplankton (including MC-producing *M. aeruginosa*) and zooplankton, and thus are widely used in the non-traditional bio-manipulation of cyanobacterial blooms (Xie and Liu, 2001; Lu et al., 2006; Zhang et al., 2008). In a laboratory study, the body mass of silver carp fed only *M. viridis* for 80 days did not significantly increase, but no mortality was observed (Xie et al.,

2004). Silver carp and big-head carp exhibited rapid growth in fish pens located in Tai Lake (Ch: *Taihu*), where heavy cyanobacterial blooms (mainly *Microcystis* spp.) occurred during warm seasons (Ke et al., 2007, 2008; Guo et al., 2009, 2015; Zhou et al., 2009). The feeding intensities of these fishes were not decreased during blooms of *Microcystis* spp., but total gut content was the greatest. Since *Microcystis* species are usually less nutritious than zooplankton to big-head carp, the decreased growth in July and August may have been due to the increased percentage of *Microcystis* spp. (>75%) in the guts (Ke et al., 2007).

Although silver carp and big-head carp directly filter and feed on *Microcystis* spp. and therefore ingest MCs, which was shown by increased concentrations of MCs in intestinal contents and walls, these two fish species did not accumulate more MCs in liver and muscle and other tissues or organs, than other species, including herbivorous white amur bream (*Parabramis pekinensis*), grass carp (*Ctenopharyngodon idellus*), omnivorous crucian carp, common carp, muddy loach (*Misgurnus anguillicaudatus*), carnivorous redfin culter (*Chanodichthys erythropterus*, previously: *Culter erythropterus*), top mouth culter (*Culter alburnus*, previously: *Culter ilishaeformis*), lake anchovy (*Coilia ectenes*), Taihu new silverfish (*Neosalanx taihuensis*) and yellow catfish (*Pseudobagrus fulvidraco*) (Xie et al., 2005; Chen et al., 2009a,b; Zhang et al., 2009). During blooms of cyanobacteria, in comparison with carnivorous crucian carp and omnivorous top mouth culter, the phytoplanktivorous silver carp and big-head carp showed less liver damage (Li et al., 2007, 2008; Qiu et al., 2007). Biochemically, the phytoplanktivorous fish had greater basal GSH concentrations and showed higher correlations among the major antioxidant enzymes (CAT, SOD, GPX and GST) in liver, which may explain their increased resistance to MCs (Qiu et al., 2007). These results indicate that phytoplanktivorous fish may detoxify and degrade MCs more actively, and are probably more resistant to MC exposure, than other fish. These findings indicate potential to use these phytoplanktivorous species to help to tolerate and control toxic cyanobacterial blooms via non-traditional biomanipulation (Xie and Liu, 2001; Zhang et al., 2008).

3.7.4. Amphibians

Concentrations of corticosteroids, including aldosterone and corticosterone, were smaller on the 3rd day but greater on the 21st day in *Xenopus laevis* fed with *M. aeruginosa* bloom material (Ziková et al., 2013; Table 8). Surprisingly, no detectable accumulation of MC-LR or affected development was observed in tadpoles after exposure. Only minor, to negligible, up-regulation of transcriptions of luteinizing hormone (LH) in brain and heat-shock protein 70 (HSP 70) and multidrug resistance protein (MDR) in liver were found. These results suggested that *X. laevis* tadpoles have some mechanism(s) conveying resistance

Table 8

Effects of oral exposure to *Microcystis* spp. on amphibians.

Amphibians	<i>Microcystis</i> spp. characteristics and MC concentrations	Dose	Duration	Toxic effects	References
Bullfrog larvae	<i>Microcystis incerta</i>	–	8 days	dm↓, snout-vent length (SVL)↓, body width↓	Pryor, 2003
<i>Xenopus laevis</i> tadpoles	Cyanobacterial biomass consisting mainly of <i>M. aeruginosa</i> collected	10%, 50% <i>M. aeruginosa</i> in diets; 42.8, 187.0 µg/g diets	1, 3, 7, 21 days	bm↓, aldosterone↓, corticosterone↓, LH mRNA in brain↑, mRNA of HSP 70, MDR in liver↑; no significant changes of mortality, developmental stages, mRNA of FSH in brain, mRNA of SOD, CAT, CYP 1A1, muGST, piGST in liver	Ziková et al., 2013
Frog tadpoles	<i>M. aeruginosa</i> NIES-90, 2.768 µg MC-LR/mg dm	2, 4, 6 (×10 ⁶ cells/mL)	7 days	No significant changes of survival	Zhang et al., 2012
Frog tadpoles	<i>Microcystis</i> sp.	6 × 10 ⁶ cells/mL	3 days	bm↑	Zhang et al., 2012
Frog tadpoles	<i>M. aeruginosa</i> NIES-90, 2.768 µg MC-LR/mg dm	–	7 days	bm↑	Zhang et al., 2012
<i>Lithobates catesbeianus</i> tadpoles	<i>M. aeruginosa</i> NPLJ4	10 ⁵ cells/mL	16 days	Liver, skeletal muscle, intestinal tract: no free MCs bioaccumulation, but serious histopathological damages	Júnior et al., 2018

↑, increased; ↓, decreased; bm, body mass; dm, dry mass; FSH, follicle stimulating hormone; LH, luteinizing hormone; CYP, cytochrome P450 monooxygenase; HSP, heat shock protein; MDR, multidrug resistance protein; CAT, catalase; GST, glutathione S-transferase; SOD, superoxide dismutase.

to cyanobacterial biomass containing MC-LR. However, serious histopathological damage was observed in the liver, intestine and skeletal muscle of tadpoles of *Lithobates catesbeianus* exposed to *M. aeruginosa* NPLJ4, which produces [D-Leu¹]MC-LR, for 16 days (Júnior et al., 2018). Samples showed signs of recovery after 15 days of depuration, but still with severe damage. No accumulation of free MCs in the tadpoles was observed by use of high performance liquid chromatography with photometric diode array detection (HPLC-PDA) or mass spectrometry (MS) analysis.

3.7.5. Mammals

The median lethal dose (lethal dose, 50%, LD₅₀) of *M. aeruginosa* cells for mice exposed to a sub-chronic, oral dose for over 21 days was 2.6 g dry *M. aeruginosa*/kg, bm (Falconer et al., 1994; Table 9). Greater concentrations of plasma γ -glutamyl transpeptidase (GGT), ALP, total bilirubin (TBIL), lesser concentrations of plasma albumin, and microstructural changes in liver were found in pigs exposed to *M. aeruginosa* bloom material in drinking water for 44 days (Falconer et al., 1994; Chorus and Bartram, 1999). However, no significant changes of water or feed intakes, body mass gain, relative growth rates, plasma TBIL, GGT, ALP, glyceraldehyde dehydrogenase (GADH) or AST activities, or liver mass were observed in lactating Holstein-Friesian cows or beef cattle which were orally exposed to *M. aeruginosa* (containing 0–13 μ g MC-LR/kg bm/day) for 3–4 weeks (Orr et al., 2001, 2003; Feitz et al., 2002). While production of milk decreased during the experiment, the authors considered that this reduction might be due to stress, crush or the physiological response of mid-late lactating cows (Feitz et al., 2002). In the milk, <2 ng MC-LR/L (Orr et al., 2001) or 0.2 μ g MC-LR/L (Feitz et al., 2002) was detected, nevertheless as the authors suggested, enzyme linked immuno-sorbent assay (ELISA)

may generate false positive results and thus this finding required further investigation (Table 9).

ELISA of the liver exposed beef cattle showed a concentration of 0.28–0.92 μ g MC-LR/g, wet mass (wm) (Orr et al., 2003). However, no MCs or 2-methyl-3-methoxy-4-phenyl butanoic acid (MMPB) was detected by HPLC or gas chromatography-mass spectrometry (GC-MS). Based on results of ELISA, concentrations of MCs in livers of exposed individuals were >1,000-fold greater than the limit of quantification (LOQ) by HPLC and GC-MS, which indicated that results of ELISA were due to cross-reaction with some chemicals other than MC-LR (Orr et al., 2003). This apparent anomaly may have been due to cross reactivity of MC antibodies with several naturally-occurring, metabolic MC-detoxification products (Metcalfe et al., 2000, 2002). Daily intake of *Microcystis* for 28 days caused greater concentrations of serum cholesterol, creatinine, phosphorus, lipases, and smaller concentrations of ALT, bilirubin, cholinesterase in rats (Adamovsky et al., 2011). Significantly less red blood cell (RBC) counts and mean corpuscular volumes (MCV), greater mean corpuscular haemoglobin (MCH) and mean corpuscular haemoglobin concentrations (MCHC), greater hepato-, spleen-, and testis- somatic indexes, and microstructural changes in liver were also observed (Adamovsky et al., 2013; Palikova et al., 2013).

3.7.6. Birds

Oral intake of *Microcystis* spp. resulted in histopathological damage of liver and testis, greater activities of plasma LDH, smaller concentrations of glucose, significant changes of antioxidant activities (GPX, GR, CAT), lipid peroxidation (LPO), and activities of activation phase (cytochrome P-450-dependent 7-ethoxyresorufin O-deethylase, EROD) and conjugation phase (GSH, GST) of detoxification in liver, brain, testis and heart of Japanese quail (*Coturnix coturnix japonica*) (Skocovska

Table 9
Effects of oral exposure to *Microcystis* spp. on mammals.

Animals	<i>Microcystis</i> spp. characteristics and MC concentrations	Dose	Duration	Toxic effects	References
Mice	<i>M. aeruginosa</i> , collected from Lake Mokoan, Australia	2.6 g algae/kg bm/day	21 days	50% of mortality	Falconer et al., 1994
Pigs	<i>M. aeruginosa</i> , collected from Lake Mokoan, Australia	80, 227, 374 mg algae/kg bm/day; 280, 796, 1312 μ g MC/kg bm; 100, 284, 469 μ g MC-LReq/kg bm	44 days	Plasma GGT \uparrow , ALP \uparrow , TBIL \uparrow , albumin \downarrow , microstructural changes in liver	Falconer et al., 1994; Chorus and Bartram, 1999
Lactating Holstein-Friesian cows	<i>M. aeruginosa</i>	1.21 \pm 0.07 μ g MC-LR/kg bm/day	3 weeks	No significant changes of blood bilirubin, GGT, GADH, AST, <2 ng MC-LR/L in milk	Orr et al., 2001
Lactating Holstein-Friesian cows	<i>M. aeruginosa</i> PCC7806	0–13 μ g MC-LR/kg bm/day	4 weeks	No significant changes of plasma TBIL, GGT, ALP, milk production \downarrow , <0.2 μ g MC-LR/L in milk	Feitz et al., 2002
Beef cattle	<i>M. aeruginosa</i>	1.42 \pm 0.3 μ g MC-LR/kg bm/day	28 days	No significant changes of water or feed intakes, body mass gain or relative growth rates, plasma GGT, GADH, AST, bilirubin, liver mass, 0.28–0.92 μ g MC-LR/kg wm in liver	Orr et al., 2003
Male Wistar albino rats	<i>Microcystis</i> , total MC 2.698 mg/g dw (-RR, -LR, -YR, 2 non identified: 1462, 1088, 96, 43, 9 μ g/g)	1% cyanobacteria in diet; 3.2 mg MC/kg bm/day	28 days	Serum CHOL \uparrow , CREA \uparrow , P \uparrow , LIP \uparrow , ALT \downarrow , BIL \downarrow , CHS \downarrow , CD3-8+ lymphocytes in peripheral blood \downarrow , MC conjugates with cysteine or GSH in liver	Adamovsky et al., 2011
Male Wistar albino rats	<i>M. aeruginosa</i> , 2500 μ g/g dm	1% cyanobacteria in diets; 26,572 μ g MC/kg diets (-LR, -RR, -YR, -LF, -LW: 8829, 15,425, 872, 671, 775 μ g/kg); 3000 μ g MC/kg bm/day	28 days	RBC count \downarrow , MCV \downarrow , MCH \uparrow , MCHC \uparrow , MCs around limit of detection (3–5 ng/g wm) in liver of 6/10 rats	Palikova et al., 2013
Male Wistar albino rats	<i>M. aeruginosa</i> s, 2500 μ g/g dm	1% cyanobacteria in diets; 26,572 μ g MC/kg diets (-LR, -RR, -YR, -LF, -LW: 8829, 15,425, 872, 671, 775 μ g/kg); 3033 μ g MC/kg bm/day	28 days	Microstructural changes in liver, hepato-, spleen-, and testes- somatic indexes \uparrow , serum BIL \uparrow , 5 ng MC-RR/g wm in liver of 6/10 rats	Adamovsky et al., 2013

\uparrow , increased; \downarrow , decreased; bm, body mass; dm, dry mass; wm, wet mass; RBC, red blood cells; MCV, mean corpuscular volume; MCH, mean corpuscular haemoglobin; MCHC, mean corpuscular haemoglobin concentrations; ALP, alkaline phosphatase; ALT, alanine amino-transferase; AST, aspartate amino-transferase; GADH, glyceraldehyde dehydrogenase; GGT, γ -glutamyl transpeptidase; BIL, bilirubin; TBIL, total bilirubin; CHOL, cholesterol; CREA, creatine; P, phosphorus; LIP, lipases; CHS, cholinesterase; GSH, glutathione.

et al., 2007; Pašková et al., 2008; Damkova et al., 2009, 2011; Pikula et al., 2010; Paskova et al., 2011; Table 10).

3.8. Accumulation of MCs in muscle and risk for human consumption

The results of several studies have shown that estimated daily intakes (EDI) of MCs via edible muscle of fish fed *Microcystis* spp. were close to or exceeding the tolerable daily intake (TDI) for the congener MC-LR of 0.04 µg/kg, bm per day for humans, established by the WHO and thus were not considered safe for human consumption (Li et al., 2004; Zhao et al., 2006b; Dong et al., 2009, 2011, 2012; Liang et al., 2015; Xia et al., 2018; Table 4). Even after periods of depuration of 43, 55 or 60 days, the ingestion of MCs would still have been greater than the WHO provisional guideline for health protection. It was therefore inferred that edible (muscle) tissue of fish exposed to *Microcystis* spp. under these conditions was not safe for human consumption (Dong et al., 2009, 2011, 2012). Results of some studies have indicated that the consumption of muscle of silver carp and big-head carp might be safe for humans (Chen et al., 2009b; Zhang et al., 2013), but others have indicated that these products are not always safe (Xie et al., 2005; Chen et al., 2006, 2007; Guo et al., 2015), especially during blooms of *Microcystis*.

3.9. Cyanotoxins and cyanobacterial metabolites beyond MCs

Microcystis spp. which do not produce MCs can also be toxic to animals. For example, diets containing the non-MC-producer *M. aeruginosa* IZANCYA6 caused a decrease in growth of juvenile crayfish (Vasconcelos et al., 2001). Biomass of *M. wesenbergii*, lacking detectable MCs, also caused significant lethality to embryos of the African clawed frog, and the estimated lethal concentration 25% (LC₂₅) value was 232 mg biomass/L (Dvořáková et al., 2002). *M. wesenbergii* also resulted in malformations, with effective concentration 50% (EC₅₀) of about 300 mg biomass/L and EC₂₅ of 75 mg biomass/L. Results showed that *M. wesenbergii* biomass can present a significant risk of teratogenicity, and that the teratogenic index (TI value) of 3.1 was even greater than that of *M. aeruginosa* (2.4). Acute oral exposure of Patagonian

pejerrey to *M. aeruginosa* NPDC1 which did not produce MCs induced an increase in GST activity in liver (Bieczynski et al., 2013). Decreased percentages of hyalinocytes and large basophilic hemocytes, and increased lysozyme activities were observed in the hemolymph of zebra mussels after dietary exposure to *M. aeruginosa* SAG 48.80, which did not produce MCs (Juhel et al., 2015).

Certainly, MCs are not the only, mainly or necessarily the major toxic components in *M. aeruginosa* (Falconer, 2007; Chen et al., 2016a). In addition to MCs, *Microcystis* spp. can synthesize and release a wide range of other metabolites including anatoxin-a, retinoic acids, microviridins, anabaenopeptins, aeruginosins, microginins, piricyclamides and cyanopeptolins (Park et al., 1993; Park and Watanabe, 1995; Osswald et al., 2009; Wu et al., 2012; Gemma et al., 2016; Otten et al., 2017). Some strains of *Microcystis* spp. can produce at least two types of cyanotoxins (Park et al., 1993). Furthermore, as Gram negative prokaryotes, cyanobacteria, including *Microcystis* spp. characteristically contain lipopolysaccharide (LPS) endotoxin as components of their cell walls (Stewart et al., 2006; Metcalf and Codd, 2012; Monteiro et al., 2017). Also, multiple *Microcystis* species or strains can co-occur within a population/community as a general condition (Metcalf and Codd, 2012; Otten et al., 2017). While the monitoring of MCs in *Microcystis* biomass, intended for animal feed, is clearly a priority requirement, analysis for additional cyanotoxins may also be necessary. Other toxins produced by *Microcystis* spp. and additional cyanobacterial genera which might co-occur in biomass, may be additive or synergistic to the toxicity of MCs. Quantitative evaluation of the total toxicity induced by MCs and other secondary metabolites is a major scientific challenge and an important issue which needs to be addressed in future toxicological studies on *Microcystis* spp. and further cyanobacteria (Pavagadhi and Balasubramanian, 2013; Chen et al., 2016a).

3.10. Additional toxic chemicals

Cyanobacteria can also bioaccumulate xenobiotics from aquatic environments, including heavy metals, polycyclic aromatic hydrocarbons (PAHs), organochlorine pesticides (OCPs), polychlorinated biphenyls

Table 10
Effects of oral exposure to *Microcystis* spp. on birds.

Birds	<i>Microcystis</i> spp. characteristics and MC concentrations	Dose	Duration	MC contents in organs	Toxic effects	References
Male Japanese quail	<i>M. aeruginosa</i> 80%, <i>M. viridis</i> 15%, <i>M. flos-aquae</i> 5%, collected from Brno reservoir, Czech Republic, 373.3 µg MC/g dm (-RR, -YR, -LR, unidentified MC: 141.8, 33.7, 141.7, 56.1 µg/g)	0.2, 2.2, 22.4, 224.3 µg/kg bm/day	10, 30 days	MC contents in liver: 10 days >30 days	Plasma: LDH↑, glucose↓; liver: micro- and ultra-structural changes; testis: microstructural changes	Skocovska et al., 2007
Male Japanese quail	<i>M. aeruginosa</i> , collected from Brno reservoir, Czech Republic, 373.3 µg MC/g dm (-RR, -YR, -LR, unidentified MC: 141.8, 33.7, 141.7, 56.1 µg/g)	0.2, 2.24, 22.46, 224.6 µg/kg bm/day	10, 30 days	MC contents in liver and muscle: 10 days >30 days	Liver: GST↑, GSH↑, GPX↓, GR↑, LPO↑; heart: EROD↑, GST↑, GSH↑, GR↑, LPO↑; brain: EROD↑, GSH↓↑, GPX↑, LPO↑	Pašková et al., 2008
Male and female Japanese quail	<i>M. aeruginosa</i> 90% and <i>M. ichthyoblabe</i> 10%, collected from Musovska reservoir, Czech Republic	61.62 µg MC/bird/day (-RR, -YR, -LR: 26.54, 7.62, 27.39 µg)	8 weeks	NA	Mass of eggs↓, mass of chicks at hatching and 14 days after hatching↓, fertilization rates (viability)↑, overall effect of hatching↑, number of 14-day old survivors per hen per day↑, plasma LDH↑	Damkova et al., 2009
Male Japanese quail	<i>M. aeruginosa</i> 90%, <i>M. ichthyoblabe</i> 10%	46.044 µg MC/bird/day (-RR, -YR, -LR: 15.36, 12.70, 17.98 µg); 210 µg/kg bm/day	10, 30 days	Liver: 39.94 ± 17.75 ng MC/g wm	Liver: microstructural changes	Pikula et al., 2010
Male Japanese quail	<i>M. aeruginosa</i> 90% and <i>M. ichthyoblabe</i> 10%, collected from Musovska reservoir, Czech Republic	61.62 µg MC/bird/day (-RR, -YR, -LR: 26.54, 7.62, 27.39 µg)	8 weeks	NA	Testis: microstructural changes, LPO↓, GPX↓, CAT↑	Damkova et al., 2011
Male Japanese quail	<i>M. aeruginosa</i>	46 µg MC/bird/day (-RR, -YR, -LR: 15.36, 12.70, 17.98 µg), 210 µg/kg bm/day	30 days	Liver: 39.9 ± 17.7 ng MC/g wm	Liver: GR↑; heart: GSH↑, GR↑	Paskova et al., 2011

↑, increased; ↓, decreased; bm, body mass; dm, dry mass; wm, wet mass; LDH, lactate dehydrogenase; EROD, 7-ethoxyresorufin O-deethylase; LPO, lipid peroxidation; CAT, catalase; GSH, glutathione; GR, glutathione reductase; GPX, glutathione peroxidase; GST, glutathione S-transferase.

(PCBs) and endocrine disrupting chemicals (EDCs) (Chen et al., 2017b; Shi et al., 2017; Jia et al., 2018, 2019). As part of the selection process for the use of *Microcystis* spp. biomass as a potential source of protein (amino acids) in feed for animals, early testing to determine potential chemical contaminants is warranted (Fig. 5). These tests continue to be necessary after a decision to proceed is made and throughout biomass harvesting and processing. The need for analysis of xenobiotics is increased due to the ability of cyanobacteria to bioaccumulate these compounds in aquatic environments. Also, due to the potential for toxic potencies of as yet unidentified compounds and/or interactions among chemicals in mixtures, testing should not be based solely on instrumental monitoring of MCs and other toxicants, but also, should include bioassays, particularly including oral toxicity determination using mice or other animals. Finally, doses used in the bioassays with mice or other animals need to be scaled appropriately to allow proper interpretation and extrapolation of the results to fish, poultry, livestock, etc.

3.11. Different susceptibilities among species and optimum proportion of *Microcystis* biomass for feed

Examples of the different susceptibilities of different animal species to oral dosing with *Microcystis* biomass were given in Section 3.6. A single oral dose of *M. aeruginosa* 7806 (370 mg *M. aeruginosa* 7806/kg bm; 1700 µg MC/kg bm) caused mortality in common carp (*Cyprinus carpio*) within 24–48 h, while the same dose resulted in no mortality of rainbow trout (Tencalla et al., 1994; Fischer and Dietrich, 2000). The phytoplanktivorous fishes, silver carp and bighead carp had greater basal GSH concentrations and higher correlations among the activities of major anti-oxidant enzymes (CAT, SOD, GPX and GST) in liver, enabling them to detoxify and degrade MCs more actively and probably conferring more resistance to MC exposure than in other fish species (Li et al., 2007, 2008; Qiu et al., 2007). The absolute lethal dose (LD₁₀₀) of *M. aeruginosa* NRC-1 to mice, guinea pigs, and rabbits were all 3.2 g/kg, bm via the oral route (Konst et al., 1965). However, on an

equivalent body mass basis, lethal doses for a lamb and two calves were 3–10 fold greater, i.e. 16 and 9.6–32 g/kg, bm, respectively. Symptoms were less pronounced and survival time was longer in the large ruminants than in laboratory animals. In chickens, oral doses required for lethality were 8–16 g/kg, bm (Konst et al., 1965).

In addition to this experimentally-determined variability, the established requirements for the identification and incorporation of environmental toxicology “safety factors” (Hughes, 1996; Chorus and Bartram, 1999) need to be included in decision-making on the suitability of *Microcystis* biomass for animal consumption. These factors include allowance for uncertainties regarding differences in susceptibility of animals to toxicants, both between- and within (e.g. versus sex and age) species. It is necessary to include these factors, typically 10-fold differences, in determining the health safety of material including cyanobacteria and the optimum ratios of addition for animal consumption.

3.12. Effects of *Microcystis* on quality of muscle

Goldfish and blunt snout bream fed with *M. aeruginosa* were of poor nutritional value and reduced quality of muscle (Liang et al., 2015; Wang et al., 2017). Significantly reduced concentrations of crude protein, crude fat and ash were observed in the muscle of goldfish which had consumed dietary *M. aeruginosa* (10%, 20%, 30%, 40% of cyanobacteria in diet) for 16 weeks (Liang et al., 2015). *M. aeruginosa* also resulted in greater concentrations of saturated fatty acids (SFA) in fish muscle, but lesser concentrations of n-3 poly-unsaturated fatty acids (PUFAs) and collagen, decreased pH and myofibril length, increased fiber diameters and cooking loss. Also, the concentrations of flavor compounds, including amino acids, nucleotides, organic acids and carnosine exhibited significant changes in fish exposed to *Microcystis* and concentrations of compounds affecting taste and odor, including geosmin (GSM) and 2-methylisoborneol (MIB) were greater (Liang et al., 2015). Dietary *M. aeruginosa* (30% of cyanobacteria in diet) also caused lesser concentrations of total amino acids in dorsal muscle, total essential amino acids (EAAs), decreased

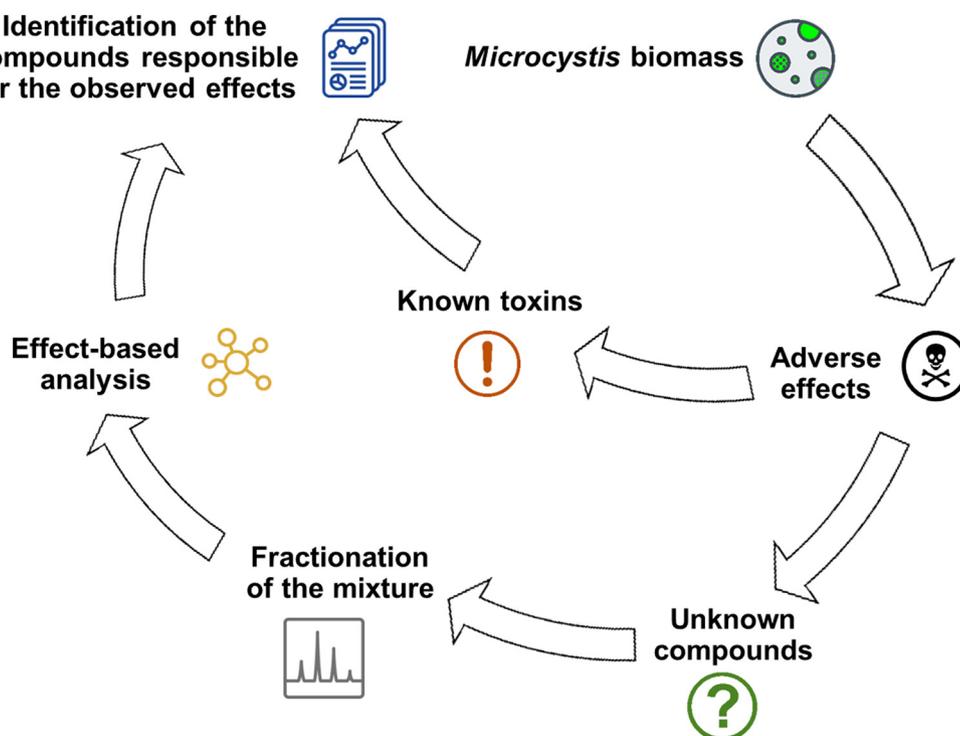


Fig. 5. The proposed workflow to identify toxic compounds responsible for the adverse effects caused by biomass of *Microcystis* spp. In scenarios, where the toxic effect is not associated with known toxins, effect-based analysis may be applied. To identify compounds responsible for the observed effects, the complex *Microcystis* biomass has to be fractionized and the individual fractions tested. Positive fractions can be further sub-fractioned to obtain individual compound or set of chemically similar compounds.

gumminess, and increased pH of muscle of blunt snout bream (Wang et al., 2017). These results suggested decreased palatability, nutritional values and quality of muscle.

3.13. Microbial pathogens in *Microcystis* bloom material

In addition to the presence of MCs, blooms of cyanobacteria and their mucilaginous layers can contain microbial pathogens that, if not removed or killed, may limit use of biomass as a source of protein in feed. Blooms of cyanobacteria create a microenvironment that provides protection and a physical substrate, as well as nutrients and useful molecules, such as sugars and amino acids, which can be used as sources of carbon and energy, for associated microbes, including bacteria, fungi and virus. For instance, some bacteria benefit from and live in close association with heterocystous cyanobacteria capable of nitrogen fixation (Paerl, 1976). Also, the general possibility of carbon-scavenging within blooms of cyanobacteria promotes growth of heterotrophic bacteria (Paerl, 1978).

Cyanobacterial blooms can contain pathogenic bacteria such as *Vibrio cholerae* (Islam et al., 1990, 1994; Epstein, 1993). Potential environmental reservoirs of *V. cholerae* were recently reviewed by Islam et al. (2020). The authors evaluated several possible inter-epidemic reservoirs of *V. cholerae*, such as aquatic fauna, including zooplankton and crustaceans, and flora, including macrophytes, cyanobacteria and microalgae. Some species of cyanobacteria, including *Anabaena variabilis* and *M. aeruginosa*, are able to act as inter-epidemic reservoirs of *V. cholerae*. Screening of cyanobacterial (*Spirulina*) and microalgal biomass harvested from open systems for faecal coliforms was negative, but positive for faecal streptococci (Jaquet, 1976), which is an early indication of the need for microbial assessment of such biomass (Jassby, 1988). Much of the work on the association between *V. cholerae* and cyanobacteria has concentrated on surface waters in Bangladesh, but there is no reason that such associations are unique to conditions in this location. Besides Bangladesh, cholera is endemic in parts of India and outbreaks of the disease have been reported in Yemen, as well as

in regions of sub-Saharan Africa and Haiti (Deen et al., 2019). Also, due to the cultural eutrophication of limited freshwater resources and high water temperatures, cyanobacteria can thrive in many of these countries. While China is not among the countries hardest hit by cholera (Deen et al., 2019), many faecal-associated and pathogenic bacteria occur in the Tiaoxi River, a major inflow entering Lake Taihu in China, where they can become associated with *Microcystis* blooms (Vadde et al., 2019).

There is thus abundant evidence that pathogenic microorganisms including bacteria, fungi and virus can be present in *Microcystis* blooms. Thus, before and/or after processing, bloom biomass intended for use as feed for animals, must be analyzed for microbial pathogens as part of the decision-making process.

4. Measures to avoid or ameliorate toxic effects caused by *Microcystis*

Due to reports of poisonings/intoxications of animals and humans and evidence that *Microcystis* spp. and MCs cause or contribute to these events, MCs have received increasing attention, particularly as a threat to public health (Metcalf and Codd, 2012; Chen et al., 2016a; Buratti et al., 2017; Svirčev et al., 2017, 2019). Cyanotoxins and xenobiotics should be removed or degraded sufficiently to assure that the threshold for toxicity of the *Microcystis* biomass is not exceeded in products intended for use as a primary (main) or supplemental protein source in animal feed. As an ameliorative measure, antidotes or detoxicants can be used to avoid or reduce the toxic effects caused by MCs and additional cyanotoxins. Here, some trials to ameliorate toxic effects caused by *M. aeruginosa* are reviewed (Fig. 6).

4.1. Removal of MCs and other bioactive metabolites produced by *Microcystis* spp. and xenobiotics accumulated by *Microcystis* spp.

Recently, the acidolysis product of materials from blooms of *M. aeruginosa* in Lake Tai by hydrochloric acid (HCl) treatment and

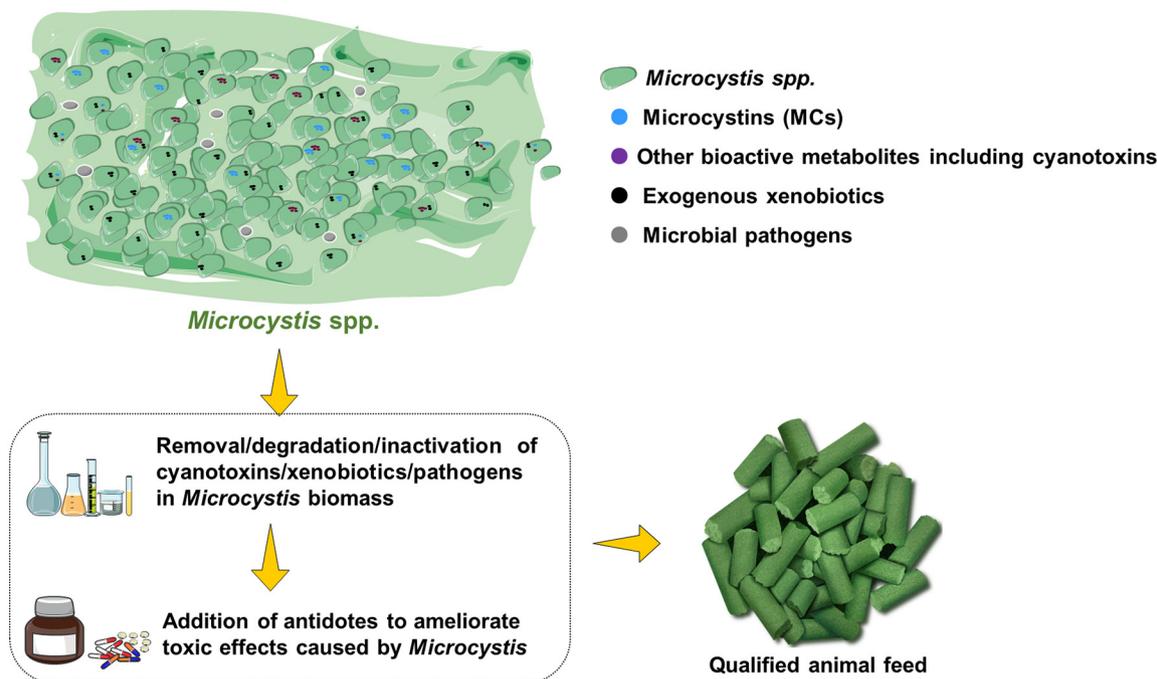


Fig. 6. Measures to avoid or ameliorate toxic effects caused by *Microcystis* spp. *Microcystis* has some limitations for use in animal feed: 1) microcystin (MC)-producing and non-MC-producing strains co-exist in a bloom; 2) *Microcystis* can produce other bio-active metabolites beyond MCs; 3) *Microcystis* can also bioaccumulate exogenous xenobiotics in aquatic environments, including heavy metals, polycyclic aromatic hydrocarbons (PAHs), organochlorine pesticides (OCPs); 4) microbial pathogens can occur in *Microcystis* blooms. Therefore, cyanotoxins/xenobiotics/pathogens in *Microcystis* biomass should be removed/degraded/inactivated sufficiently to assure safety for use as a primary/main/supplemental ingredient in animal feed. As an ameliorative measure, antidotes/detoxicants can be used to avoid/reduce the toxic effects.

heating was reported to be safe for use as an ingredient of animal feed (Han et al., 2015). After extraction of *M. aeruginosa* by heating and acidolysis using HCl, the concentrations of amino acids, residual MCs, and heavy metals of the acidolysis product were determined. After 18 h of heating and acidification, the product contained 17 identified amino acids. These amino acids accounted for 51% of total acidolysis product, and 30% of the remainder were essential amino acids (EAAs) for livestock and poultry. The concentration of residual MC-LR was 0.94 µg/kg, dm, which was less than the WHO provisional guideline value of MCs for human drinking water limit (1 µg MC-LR/L, equal to 1 µg MC-LR/kg). Concentrations of As, Pb, Hg and Cr in the feed were in compliance with National Standards of China for feed (Han et al., 2015). Furthermore, the results of Horn's method showed that the LD₅₀ dose of the acidolysis product to ICR mice via the oral route was >9.09 g/kg, bm, which was classified as being non-toxic (5001–15,000 mg/kg) (Han et al., 2015). None of the oral dosages of 2.15, 4.64, 10 or 21.5 g/kg, bm of the *M. aeruginosa* acidolysis product resulted in significantly greater activities of hepatic ALP, serum LDH or γ-glutamyltransferase (γ-GT). The acidolysis product exhibited neither mutagenicity nor effects on sperm malformation. It was concluded that the product was safe for use as an animal feed ingredient (Han et al., 2015).

Water treatment technologies to remove toxic compounds may be used for the treatment of *Microcystis* spp. biomass, including disinfection processes and the destruction of cyanotoxins by common oxidants, including free chlorine (NaOCl), chlorine dioxide (ClO₂), chloramines (mainly monochloramine, NH₂Cl), permanganate (KMnO₄) and ferrate (FeO₄²⁻), adsorption by activated carbon, iron-based adsorbents, and advanced oxidation processes including ultrasound, ultraviolet (UV), UV/H₂O₂, and ozone (He et al., 2016; Zhou et al., 2019). However, unlike in water treatment technologies, the cyanobacterial proteins and other nutrients intended for animal nutrition should not be affected.

Properly processed products containing *Microcystis* spp. might be safe for use as an ingredient in animal feed if the following two requirements are met:

- 1) Cyanotoxins and other bioactive metabolites produced by *Microcystis* spp., including MCs, and xenobiotics accumulated by

blooms of *Microcystis* spp., e.g. heavy metals and pesticides, are sufficiently removed to prevent significant oral exposure to cause adverse health outcomes in the animals being fed and in subsequent human consumers of products of those animals;

- 2) No toxicants are produced or introduced during the harvesting and processing of the *Microcystis* spp. biomass.

4.2. Potential use of antidotes to ameliorate toxic effects caused by *Microcystis*

Microcystis or MC-induced toxicity in fish and mammals can be ameliorated to different degrees by antidotes or detoxicants, including: 1) transporter inhibitors, e.g. bile acids, cyclosporin A, rifampicin, cytochalasin, trypan blue, trypan red and naringin; 2) anti-inflammatory agents, e.g. several glucocorticoids, including flucinolone, dexamethasone, and hydrocortisone; 3) osmotic agents, e.g. D-glucose, mannitol, and dihydroxy acetone; and 4) antioxidants, e.g. GSH, N-acetylcysteine (NAC), L-cysteine, vitamin C, vitamin E, selenium, melatonin, flavonoids, quercetin, silybin, and morin, and green tea polyphenols, primarily catechins, and sulforaphane (Guzmán-Guillén et al., 2017). Some results concerning the elimination or alleviation of MC effects by antidotes have been summarized by Guzmán-Guillén et al. (2017). Trolox, a soluble vitamin E analogue (Prieto et al., 2008, 2009; Table 11), sodium selenite (Na₂SeO₃) pentahydrate (Atencio et al., 2009), and NAC, a precursor of glutathione (GSH) (Puerto et al., 2009, 2010), can ameliorate histopathological damage and oxidative stress caused by *Microcystis* spp. in fish. Naringin can inhibit the uptake of MC-LR in the freshwater snail exposed to *M. ichthyoblabe* (Xie et al., 2014).

4.2.1. Naringin

Naringin, a flavonoid isolated from grape and citrus fruit species, exhibits therapeutic potential, including anti-oxidant, anti-inflammatory, anti-ulcer, anti-atherogenic, anti-cancer, hepato-protective and neuro-protective activities (Xie et al., 2014). It is also a clinical inhibitor of the organic anion-transporting polypeptide 1A2 (OATP1A2).

Naringin was shown to inhibit accumulation of MC-LR by freshwater snails (Xie et al., 2014). Initial treatment with 1 mM (0.581 g/L) naringin only on the first day resulted in prevention of accumulation of MC-LR in

Table 11
Protective role of antidotes versus toxic effects caused by acute oral exposure to *Microcystis* spp.

Animals	<i>Microcystis</i> spp. characteristics and MC concentrations	MC-LR dose and duration	Antidote dose and duration	Protective role	References
Freshwater snail <i>Sinotia histrica</i>	<i>M. ichthyoblabe</i> TAC95, 13.7 µg MC/mg dm	10 ⁷ cells/mL, 8 days	Naringin, 0.581, 5.81 g/L; 1, 8 days	Hepatopancreas: inhibition of MC-LR accumulation	Xie et al., 2014
male Nile tilapia	<i>M. aeruginosa</i> , 1350 µg MC-LR/g dm	about 2400 µg MC-LR/kg bm, crushed <i>M. aeruginosa</i> cells; single dose; 24 h	Trolox, a vitamin E analog; about 1.2, 4.2 mg/kg bm/day; 7 days	Liver: LPO, protein oxidation, CAT, SOD; kidney: LPO, CAT, SOD, GR; gill: LPO, CAT, SOD, GPX	Prieto et al., 2008
Male Nile tilapia	<i>M. aeruginosa</i> , 1350 µg MC-LR/g dm	about 2400 µg MC-LR/kg bm, crushed <i>M. aeruginosa</i> cells; single dose; 24, 48, 72 h	Trolox; about 4.2 mg/kg bm/day; 7 days	Liver: micro- and ultra-structural changes, LPO, SOD, CAT, GPX, GR, GST, GSH/GSSG; kidney: micro- and ultra-structural changes, LPO, CAT, GPX, GR; gill: LPO, SOD, CAT, GPX, GR; heart and gastrointestinal tract: micro- and ultra-structural changes	Prieto et al., 2009
Male Nile tilapia	<i>Microcystis</i> , collected from Guadiana River, Portugal, 2885 µg MC-LR/g dm	about 2400 µg MC-LR/kg bm, crushed <i>Microcystis</i> cells; single dose; 24 h	Sodium selenite (Na ₂ SeO ₃); 10, 18, 36 µg selenium (Se)/kg/day; 7 days	Liver: micro- and ultra-structural changes, CAT, GPX, GST; kidney: micro- and ultra-structural changes, CAT, SOD, GPX; heart: micro- and ultra-structural changes; gastrointestinal tract: micro- and ultra-structural changes	Atencio et al., 2009
Male Nile tilapia	<i>Microcystis</i> , collected from Guadiana River, Portugal, 2885 µg MC-LR/g dm	about 2400 µg MC-LR/kg bm, crushed <i>Microcystis</i> cells; single dose; 24 h	N-acetylcysteine (NAC), a GSH precursor; 400, 880, 1936 mg NAC/kg/day; 7 days	Liver: LPO, protein oxidation, protein, CAT, GPX, GST, GSH/GSSG; kidney: LPO, CAT, SOD, GPX, GR	Puerto et al., 2009
Male Nile tilapia	<i>Microcystis</i> , collected from Guadiana River, Portugal, 2885 µg MC-LR/g dm	about 2400 µg MC-LR/kg bm, crushed <i>Microcystis</i> cells; single dose; 24 h	NAC; 400, 880, 1936 mg NAC/kg/day; 7 days	Liver, kidney, heart, gastrointestinal tract, and gill: micro- and ultra-structural changes	Puerto et al., 2010

bm, body mass; dm, dry mass; LPO, lipid peroxidation; CAT, catalase; SOD, superoxide dismutase; GSH, glutathione; GSSG, oxidized glutathione; GR, glutathione reductase; GPX, glutathione peroxidase; GST, glutathione S-transferase.

the hepatopancreas of the freshwater snail *S. histrica* exposed to *M. ichthyoblabe* TAC95 by approximately 60% over a subsequent 8 days. Initial treatment with 10 mM naringin only on the first day suppressed MCs accumulation in the first 2 days, but concentrations of MC-LR increased in the animals from the 5th to 8th day. With continuous treatment of 10 mM of naringin for 8 days, the MCs uptake was completely prevented.

4.2.2. Vitamin E

Vitamin E is a generic descriptor for all compounds which have the biopotency of α -tocopherol (Prieto et al., 2008). All of the natural forms of vitamin E are D-stereoisomers and consist of a substituted aromatic ring and a long isoprenoid side chain. Among the 8 natural molecules with vitamin E activity, α -tocopherol exhibits the greatest biological activities. Vitamin E is a lipid-soluble anti-oxidant which can decrease LPO and thus can protect biological membranes against free radical-induced damage.

Effects of pre-treatment with Trolox (6-hydroxy-2,5,7,8-tetramethylchromane-2-carboxylic acid), a water-soluble vitamin E analogue, on MC-induced oxidative stress/damage in the liver, kidney, and gill of male Nile tilapia were investigated (Prieto et al., 2008). The fish were fed Trolox (at about 1.2, and 4.2 mg vitamin E/kg, bm/day) for 7 days and then were fed a single oral dosage of *M. aeruginosa* (about 2400 μ g MC-LR/kg, bm) in the diet. *M. aeruginosa* caused greater LPO values in liver, while the fish pre-treated with vitamin E showed no alteration. Activities of anti-oxidant enzymes, including CAT, SOD, GPX and GR, were also ameliorated by vitamin E-dosing, while protein oxidation and the GSH/GSSG ratio did not exhibit obvious changes. The larger dosage of vitamin E (about 4.2 mg vitamin E/kg, bm/day) had a greater protective efficacy, especially upon LPO and CAT activity (Prieto et al., 2008).

The time-dependent protective actions of Trolox on oxidative stress/injury and histopathological damage caused by MC-producing *M. aeruginosa*, were also studied (Prieto et al., 2009). Male Nile tilapia were fed Trolox supplement (about 4.2 mg vitamin E/kg, bm/day) for 7 days and then fed a single oral dose of *M. aeruginosa* (about 2400 μ g MC-LR/kg, bm) in the diet, and sacrificed after 24, 48, or 72 h. For the biomarkers of oxidative stress including LPO, SOD, CAT, GR and GST in liver, kidney and gill changed by *M. aeruginosa*, a greater protective efficacy of vitamin E was observed 24 h post exposure to *M. aeruginosa*, although protection extended for up to 48 h in gills for some biomarkers. Treatment with Trolox also reduced micro- and ultra-structural injury in liver, kidneys, gastrointestinal tract and heart, and the protective abilities were more obvious after 72 h (Prieto et al., 2009).

4.2.3. Selenium (Se)

Selenium (Se) plays dual but contradictory roles in animals because it is both a nutrient and a toxicant (Atencio et al., 2009). Se can protect cells against oxidative stress as an integral component of GPX which contains 4 sub-units and every sub-unit includes a Se atom. Se is also a component of iodothyronine deiodinase and thioredoxin reductase, which is the only enzyme found to catalyze the reduction of thioredoxin and is involved in modulation of cellular redox homeostasis and protection against oxidative stress/injury. However, Se also can be toxic by reaction with sulfhydryl groups to generate biologically active reactive oxygen species (ROS). As a nutrient, the requirement of dietary Se for fish is 0.1–0.5 μ g/g dm. However, Se becomes toxic at only 7–30 times the required dietary concentrations for nutrition (>3 μ g/g, dm). Therefore, it is critical for homeostatic regulation that optimal amounts of Se are available to protect cells/tissues/organs from ROS-induced oxidative stress and to maintain overall health.

The protective effects of dietary Se supplement (sodium selenite, Na₂SeO₃) against morphological alterations and oxidative stress/damage caused by MC-containing *Microcystis* sp. in male Nile tilapia were investigated (Atencio et al., 2009). Fish were fed NaSeO₃ (about 10, 18, and 36 μ g Se/kg, bm/day) for 7 days and then fed a single oral dose of *Microcystis* sp. (about 2400 μ g MC-LR/kg, bm), and killed after 24 h.

The protective actions of Se were dose-dependent. The lesser dosage of Se caused CAT and GR activities in liver and CAT and SOD activities kidney to converge to baseline values, but the reversal of LPO values and SOD and GST activities in liver required the greater dosage. However, Se induced a pro-oxidant effect with increased LPO in kidney and GPX activities in liver and kidney of fish which were not exposed to *Microcystis*. Amelioration of *Microcystis* sp.-caused pathological changes in liver, gastrointestinal tract, kidney and heart was observed by the greatest dosage of Se. Therefore, careful attention should be paid to control the amounts of Se supplementation to promote beneficial effects and to avoid potential detrimental outcomes.

4.2.4. N-acetylcysteine

N-acetylcysteine (NAC) is a derivative of the natural amino acid L-cysteine (L-Cys). It is a thiolic anti-oxidant, and a mucolytic agent for various respiratory diseases, and can act against cellular degeneration (Puerto et al., 2009). It is a precursor as a supplier of cysteine (Cys) in the synthesis of GSH and can stimulate activities of cytosolic enzymes involved in the GSH cycle, including GR, enhancing the rate of GSH regeneration from GSSG. NAC can also protect cells by reacting with ROS.

The protective role of NAC against the oxidative stress/damage (Puerto et al., 2009) and pathological changes (Puerto et al., 2010) induced by MC-containing *Microcystis* sp. in male Nile tilapia were investigated. Tilapia were fed NAC (about 400, 880 and 1936 mg/kg, bm/day) for 7 days and then fed a single oral dose of *Microcystis* sp. (about 2400 μ g MC-LR/kg, bm), and killed after 24 h. NAC exerted a dose-dependent protective role (Puerto et al., 2009). Smaller doses of NAC ameliorated the increased LPO in liver and kidney and the decreased protein content and GSH/GSSG in liver caused by *Microcystis* sp. Greater activities of anti-oxidant enzymes, including SOD, GPx, and GR, induced by *Microcystis* sp., were also recovered by NAC pretreatment. However, the greatest dose of NAC caused changes of some enzyme activities. Similarly, results of histopathology in the liver, kidneys, gills, intestine and heart suggested protective actions of NAC mainly at the median dose (880 mg NAC/kg, bm/day) probably due to its anti-oxidant activity (Puerto et al., 2010). However, the greatest dose of NAC caused toxic effects in fish not exposed to *Microcystis*. Thus, NAC can be used as a chemo-protectant in the prophylaxis, prevention and treatment of MC-induced toxicity induced by *Microcystis* sp. in fish. However, the optimal dose of NAC must be selected carefully because of its own pro-oxidant and pathological activities, which were induced by 1936 mg NAC/kg, bm/day in these studies (Puerto et al., 2009, 2010).

Finally, it is emphasized that the concentrations of antidotes applied to ameliorate toxic effects caused by *Microcystis* spp. in dosed animals should be well controlled to avoid possible risks to health of humans, or other consumers of livestock, poultry or fish, which were fed diets containing *Microcystis*.

5. Cost control

Cyanobacteria such as *Spirulina platensis* and microalgae including *Chlorella vulgaris*, are expensive to culture and produce, although efforts are under way addressing cost-efficient mass cultivation of these organisms (Vanthoor-Koopmans et al., 2014). Compared with cyanobacteria such as *Spirulina platensis* and microalgae including *Chlorella vulgaris*, *Microcystis* biomass is highly available from naturally occurring scums and blooms, and it is not necessary to culture *Microcystis*, so there are no costs for production of *Microcystis* biomass. However, cost control for harvesting and treatment of *Microcystis* biomass, including dewatering, drying and feed production is needed. Also, costs for measures to avoid/ameliorate effects of toxic *Microcystis* biomass, including removal of MCs, other bioactive metabolites and xenobiotics accumulated by *Microcystis*, and use of antidotes, are needed to be considered. Pragmatic large-scale techniques must be devised to reduce the costs of treatment of *Microcystis* biomass and to compete with conventional protein sources.

6. Conclusions

Microcystis spp., including *M. aeruginosa*, are rich in protein and have been suggested as a source of protein in animal feeds. However, in addition to protein, *Microcystis* spp. need to be considered as sources of a wider range of nutrients and metabolites including nutritionally beneficial components but also as animal and human health hazards. They can produce unpleasant taste and odor compounds, and potent toxins including MCs, which can cause multiple toxicities to animals and humans, including hepatotoxicity, reproductive toxicity, developmental toxicity, nephrotoxicity, neurotoxicity, immunomodulation, endocrine disruption, and death. Without effective procedures to monitor and, if necessary, to remove or detoxify cyanotoxins, the presence of these potent toxins restricts the use of *M. aeruginosa* as a source of protein in animal feedstuffs. Here results of studies of nutritional and toxicological aspects of *Microcystis* spp. on fish, mollusks, crustaceans, amphibians, mammals and birds have been reviewed. Inclusion of *M. aeruginosa* in diets has resulted in increased mortality, decreased growth, and caused toxicities, such as hepatotoxicity, gastrointestinal toxicity, nephrotoxicity, cardiotoxicity, neurotoxicity, immunotoxicity and animal death. The estimated daily intake (EDI) values of MCs in muscle of fish fed with *Microcystis* spp. are close to or exceed the tolerable daily intake (TDI) derivation of 0.04 µg/kg, bm/day for humans, established by the WHO, and are thus not safe for consumption by humans. In addition, microbial pathogens can be present in *Microcystis* blooms. Properly processed products containing, or originating from *Microcystis* spp. might be safe for use as an ingredient in animal feed if the following three requirements are met: 1) toxicants produced by *Microcystis* spp., including MCs, additional cyanotoxins, and xenobiotics accumulated by *Microcystis* spp., including heavy metals and pesticides, and microbial pathogens are sufficiently removed or inactivated, to prevent significant exposure to cause adverse health outcomes in the animals being fed, or in consumers of products of those animals; 2) no toxicants or microbial pathogens are produced or introduced during processing of *Microcystis* spp.; 3) taste and odor compounds produced by *Microcystis* spp. should be removed. Antidotes or detoxicants can be used to avoid or reduce toxic effects caused by MCs and *Microcystis*. Before the use of *Microcystis* spp. as sources of protein for animal feed, further, well-designed and relevant investigations and safety and health evaluation are required. Also, cost control for harvesting and treatment of *Microcystis* biomass, feed production and measures to avoid/ameliorate toxic effects needs to be considered.

CRedit authorship contribution statement

Liang Chen conceived the idea, wrote the first draft manuscript and prepared the tables. John P. Giesy, Ondrej Adamovsky, Zorica Svirčev, Jussi Meriluoto, Geoffrey A. Codd and Biljana Mijovic wrote parts of the manuscript. Liang Chen, Ting Shi, Ondrej Adamovsky and Geoffrey A. Codd drew the figures. Liang Chen, John P. Giesy, Geoffrey A. Codd, Xun Tuo, Shang-Chun Li, Jun Chen and Ping Xie revised then edited the manuscript. All co-authors contributed to and checked the final draft.

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Declaration of competing interest

None.

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References

- Abdel-Latif, H.M.R., Khashaba, A.M.A., 2017. Subchronic toxicity of Nile tilapia with different exposure routes to *Microcystis aeruginosa*: histopathology, liver functions, and oxidative stress biomarkers. *Veterinary World* 10, 955–963.
- Acuña, S., Baxa, D., Teh, S., 2012a. Sublethal dietary effects of microcystin producing *Microcystis* on threadfin shad, *Dorosoma petenense*. *Toxicol.* 60, 1191–1202.
- Acuña, S., Deng, D.-F., Lehman, P., Teh, S., 2012b. Sublethal dietary effects of *Microcystis* on Sacramento splittail, *Pogonichthys macrolepidotus*. *Aquat. Toxicol.* 110, 1–8.
- Adamovsky, O., Kopp, R., Zikova, A., Blaha, L., Kohoutek, J., Ondrackova, P., Paskerova, H., Mares, J., Palikova, M., 2011. The effect of peroral administration of toxic cyanobacteria on laboratory rats (*Rattus norvegicus* var. alba). *Neuroendocrinol. Lett.* 32, 101–111.
- Adamovsky, O., Palikova, M., Ondrackova, P., Zikova, A., Kopp, R., Mares, J., Pikula, J., Paskerova, H., Kohoutek, J., Blaha, L., 2013. Biochemical and histopathological responses of Wistar rats to oral intake of microcystins and cyanobacterial biomass. *Neuroendocrinol. Lett.* 34, 11–20.
- Adamovsky, O., Moosova, Z., Pekarova, M., Basu, A., Babica, P., Svihalkova, S.L., Kubala, L., Blaha, L., 2015. Immunomodulatory potency of microcystin, an important water-polluting cyanobacterial toxin. *Environmental Science & Technology* 49, 12457–12464.
- Atencio, L., Moreno, I., Jos, Á., Prieto, A.I., Moyano, R., Blanco, A., Cameán, A.M., 2009. Effects of dietary selenium on the oxidative stress and pathological changes in tilapia (*Oreochromis niloticus*) exposed to a microcystin-producing cyanobacterial water bloom. *Toxicol.* 53, 269–282.
- Balder, H.F., de Vogel, J., Jansen, M.C.J.F., Weijenberg, M.P., van den Brandt, P.A., Westenbrink, S., van der Meer, R., Goldbohm, R.A., 2006. Heme and chlorophyll intake and risk of colorectal cancer in the Netherlands cohort study. *Cancer Epidemiology and Prevention Biomarkers* 15, 717–725.
- Baxa, D.V., Kurobe, T., Ger, K.A., Lehman, P.W., Teh, S.J., 2010. Estimating the abundance of toxic *Microcystis* in the San Francisco Estuary using quantitative real-time PCR. *Harmful Algae* 9, 342–349.
- Becker, E.W., 2007. Micro-algae as a source of protein. *Biotechnol. Adv.* 25, 207–210.
- Bernard, C., Ballot, A., Thomazeau, S., Malufi, S., Furey, A., Mankiewicz-Boczek, J., Pawlik-Skowronska, B., Capelli, C., Salmasso, N., 2017. Appendix 2: cyanobacteria associated with the production of cyanotoxins. In: Meriluoto, J., Spoof, L., Codd, G.A. (Eds.), *Handbook of Cyanobacterial Monitoring and Cyanotoxin Analysis*. Wiley, Chichester UK, pp. 500–525.
- Bickel, H., Lyck, S., Utkilen, H., 2000. Energy state and toxin content - experiments on *Microcystis aeruginosa* (Chroococcales, Cyanophyta). *Phycologia* 39, 212–218.
- Bieczynski, F., Bianchi, V.A., Luquet, C.M., 2013. Accumulation and biochemical effects of microcystin-LR on the Patagonian pejerrey (*Odontesthes hatcheri*) fed with the toxic cyanobacteria *Microcystis aeruginosa*. *Fish Physiol. Biochem.* 39, 1309–1321.
- Bouaicha, N., Miles, C.O., Beach, D.G., Labidi, Z., Djabri, A., Benayache, N.Y., Nguyen-Quang, T., 2019. Structural diversity, characterization and toxicology of microcystins. *Toxins* 11, 714.
- Boyd, C.E., 1973. Amino acid composition of freshwater algae. *Archiv fuer Hydrobiologie* 72, 1–9.
- Buratti, F.M., Manganeli, M., Vichi, S., Stefanelli, M., Scardala, S., Testai, E., Funari, E., 2017. Cyanotoxins: producing organisms, occurrence, toxicity, mechanism of action and human health toxicological risk evaluation. *Arch. Toxicol.* (3), 1049–1130.
- Cardozo, K.H.M., Guaratini, T., Barros, M.P., Falcão, V.R., Tonon, A.P., Lopes, N.P., Campos, S., Torres, M.A., Souza, A.O., Colepicolo, P., Pinto, E., 2007. Metabolites from algae with economical impact. *Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology* 146, 60–78.
- Catherine, A., Bernard, C., Spoof, L., Bruno, M., 2017. Chapter 11: Cyanotoxins and nodularins. In: Meriluoto, J., Spoof, L., Codd, G.A. (Eds.), *Handbook of Cyanobacterial Monitoring and Cyanotoxin Analysis*. Wiley, Chichester UK, pp. 109–126.
- Chen, L., Xie, P., 2016. Mechanisms of microcystin-induced cytotoxicity and apoptosis. *Mini-Rev. Med. Chem.* 16, 1018–1031.
- Chen, J., Xie, P., Zhang, D.W., Ke, Z.X., Yang, H., 2006. *In situ* studies on the bioaccumulation of microcystins in the phytoplanktivorous silver carp (*Hypophthalmichthys molitrix*) stocked in Lake Taihu with dense toxic *Microcystis* blooms. *Aquaculture* 261, 1026–1038.

- Chen, J., Xie, P., Zhang, D.W., Lei, H.H., 2007. *In situ* studies on the distribution patterns and dynamics of microcystins in a biomanipulation fish–bighead carp (*Aristichthys nobilis*). *Environ. Pollut.* 147, 150–157.
- Chen, J., Xie, P., Li, L., Xu, J., 2009a. First identification of the hepatotoxic microcystins in the serum of a chronically exposed human population together with indication of hepatocellular damage. *Toxicol. Sci.* 108, 81–89.
- Chen, J., Zhang, D.W., Xie, P., Wang, Q., Ma, Z.M., 2009b. Simultaneous determination of microcystin contaminations in various vertebrates (fish, turtle, duck and water bird) from a large eutrophic Chinese lake, Lake Taihu, with toxic *Microcystis* blooms. *Sci. Total Environ.* 407, 3317–3322.
- Chen, W., Jia, Y.L., Li, E.H., Zhao, S., Zhou, Q.C., Liu, L.M., Song, L.R., 2012. Soil-based treatments of mechanically collected cyanobacterial blooms from Lake Taihu: efficiencies and potential risks. *Environmental Science & Technology* 46, 13370–13376.
- Chen, L., Zhang, X.Z., Zhou, W.S., Qiao, Q., Liang, H.L., Li, G.Y., Wang, J.H., Cai, F., 2013. The interactive effects of cytoskeleton disruption and mitochondria dysfunction lead to reproductive toxicity induced by microcystin-LR. *PLoS One* 8, e53949.
- Chen, L., Zhang, X., Chen, J., Zhang, X.Z., Fan, H.H., Li, S.C., Xie, P., 2014. NF- κ B plays a key role in microcystin-RR-induced HeLa cell proliferation and apoptosis. *Toxicol.* 87, 120–130.
- Chen, L., Chen, J., Zhang, X.Z., Xie, P., 2016a. A review of reproductive toxicity of microcystins. *J. Hazard. Mater.* 301, 381–399.
- Chen, L., Li, S.C., Guo, X.C., Xie, P., Chen, J., 2016b. The role of GSH in microcystin-induced apoptosis in rat liver: involvement of oxidative stress and NF- κ B. *Environmental Toxicology* 31, 552–560.
- Chen, L., Hu, Y.F., He, J., Chen, J., Giesy, J.P., Xie, P., 2017a. Responses of the proteome and metabolome in livers of zebrafish exposed chronically to environmentally relevant concentrations of microcystin-LR. *Environmental Science & Technology* 51, 596–607.
- Chen, W., Jia, Y.L., Liu, A.Y., Zhou, Q.C., Song, L.R., 2017b. Simultaneous elimination of cyanotoxins and PCBs via mechanical collection of cyanobacterial blooms: an application of “green-bioadsorption concept”. *J. Environ. Sci.* 57, 118–126.
- Chen, L., Wang, Y.K., Giesy, J.P., Chen, F., Shi, T., Chen, J., Xie, P., 2018. Microcystin-LR affects the hypothalamic-pituitary-inter-renal (HPI) axis in early life stages (embryos and larvae) of zebrafish. *Environ. Pollut.* 241, 540–548.
- Chorus, I., Bartram, J. (Eds.), 1999. *Toxic Cyanobacteria in Water: A Guide to Their Public Health Consequences, Monitoring and Management*. E & FN Spon, London, on behalf of the World Health Organization, Geneva, p. 416.
- Christaki, E., Florou-Paneri, P., Bonos, E., 2011. Microalgae: a novel ingredient in nutrition. *Int. J. Food Sci. Nutr.* 62, 794–799.
- Clearwater, S.J., Wood, S.A., Phillips, N.R., Parkyn, S.M., Van Ginkel, R., Thompson, K.J., 2014. Toxicity thresholds for juvenile freshwater mussels *Echyridella menziesii* and crayfish *Paranephrops planifrons*, after acute or chronic exposure to *Microcystis* sp. *Environ. Toxicol.* 29, 487–502.
- Codd, G.A., Meriluoto, J., Metcalf, J.S., 2017. Chapter 1: introduction: cyanobacteria, cyanotoxins, their human impact, and risk management. In: Meriluoto, J., Spoof, L., Codd, G.A. (Eds.), *Handbook of Cyanobacterial Monitoring and Cyanotoxin Analysis*. Wiley, Chichester, UK, pp. 3–8.
- Colla, E., Menegotto, A.L.L., Kalschne, D.L., da Silva-Buzanello, R.A., Canan, C., Drunkler, D.A., 2020. Chapter 32 - microalgae: a new and promising source of food. In: Konur, O. (Ed.), *Handbook of Algal Science, Technology and Medicine*. Elsevier, pp. 507–518.
- Da Rós, P.C.M., Silva, C.S.P., Silva-Stenico, M.E., Fiore, M.F., de Castro, H.F., 2012. *Microcystis aeruginosa* lipids as feedstock for biodiesel synthesis by enzymatic route. *J. Mol. Catal. B Enzym.* 84, 177–182.
- Damkova, V., Sedlackova, J., Bandouchova, H., Peckova, L., Vitula, F., Hilscherova, K., Paskova, V., Kohoutek, J., Pohanka, M., Pikula, J., 2009. Effects of cyanobacterial biomass on avian reproduction: a Japanese quail model. *Neuroendocrinol. Lett.* 30, 205–210.
- Damkova, V., Paskova, V., Bandouchova, H., Hilscherova, K., Sedlackova, J., Novotny, L., Peckova, L., Vitula, F., Pohanka, M., Pikula, J., 2011. Testicular toxicity of cyanobacterial biomass in Japanese quails. *Harmful Algae* 10, 612–618.
- Das, B.K., Pradhan, J., Sahu, S., Marhual, N.P., Mishra, B.K., Eknath, A.E., 2013. *Microcystis aeruginosa* (Kütz) incorporated diets increase immunity and survival of Indian major carp *Labeo rohita* (Ham.) against *Aeromonas hydrophila* infection. *Aquac. Res.* 44, 918–927.
- de la Fuente, G., Flores, A., Molina, M.R., Almengor, L., Bressani, R., 1977. Some nutritional characteristics of a naturally occurring alga (*Microcystis* sp.) in a Guatemalan Lake. *Appl. Environ. Microbiol.* 33, 6–9.
- de Moor, F.C., Scott, W.E., 1985. Digestion of *Microcystis aeruginosa* by *Oreochromis mossambicus*. *Journal of the Limnological Society of South Africa* 11, 14–19.
- de Vogel, J., Jonker-Termont, D.S.M.L., Van Lieshout, E.M.M., Katan, M.B., van der Meer, R., 2005. Green vegetables, red meat and colon cancer: chlorophyll prevents the cytotoxic and hyperproliferative effects of haem in rat colon. *Carcinogenesis* 26, 387–393.
- Deen, J., Mengel, M.A., Clemens, J.D., 2019. Epidemiology of cholera. *Vaccine* 38, A31–A40.
- Deng, X.W., Chen, J., Hansson, L.-A., Zhao, X., Xie, P., 2020. Eco-chemical mechanisms govern phytoplankton emissions of dimethylsulfide in global surface waters. *Natl. Sci. Rev.* nwa140.
- Dong, G.F., Zhu, X.M., Han, D., Yang, Y.X., Song, L.R., Xie, S.Q., 2009. Effects of dietary cyanobacteria of two different sources on growth and recovery of hybrid tilapia (*Oreochromis niloticus* \times *O. aureus*). *Toxicol.* 54, 208–216.
- Dong, G.F., Zhu, X.M., Han, D., Yang, Y.X., Song, L.R., Xie, S.Q., 2011. Response and recovery of hybrid sturgeon from subchronic oral administration of cyanobacteria. *Environ. Toxicol.* 26, 161–170.
- Dong, G.F., Xie, S.Q., Zhu, X.M., Han, D., Yang, Y.X., Song, L.R., Gan, L.Q., Chen, W., 2012. Responses of yellow catfish (*Pelteobagrus fulvidraco* Richardson) exposed to dietary cyanobacteria and subsequent recovery. *Toxicol.* 60, 1298–1306.
- Du, X.D., Liu, H.H., Yuan, L., Wang, Y.Q., Ma, Y., Wang, R., Chen, X.H., Losiewicz, M.D., Guo, H.X., Zhang, H.Z., 2019. The diversity of cyanobacterial toxins on structural characterization, distribution and identification: a systematic review. *Toxins* 11, 530.
- Dvořáková, D., Dvořáková, K., Bláha, L., Maršálek, B., Knotková, Z., 2002. Effects of cyanobacterial biomass and purified microcystins on malformations in *Xenopus laevis*: teratogenesis assay (FETAX). *Environ. Toxicol.* 17, 547–555.
- Epstein, P.R., 1993. Algal blooms in the spread and persistence of cholera. *Biosystems* 31, 209–221.
- Falconer, I.R., 2007. Cyanobacterial toxins present in *Microcystis aeruginosa* extracts - more than microcystins! *Toxicol.* 50, 585–588.
- Falconer, I.R., Burch, M.D., Steffensen, D.A., Choice, M., Coverdale, O.R., 1994. Toxicity of the bluegreen alga (cyanobacterium) *Microcystis aeruginosa* in drinking water to growing pigs, as an animal model for human injury and risk assessment. *Environ. Toxicol. Water Qual.* 9, 131–139.
- Feitz, A.J., Lukondeh, T., Moffitt, M.C., Burns, B.P., Naidoo, D., Vedova, D.J., Gooden, J.M., Neilan, B.A., 2002. Absence of detectable levels of the cyanobacterial toxin (microcystin-LR) carry-over into milk. *Toxicol.* 40, 1173–1180.
- Fernandes, S., Welker, M., Vasconcelos, V.M., 2009. Changes in the GST activity of the mussel *Mytilus galloprovincialis* during exposure and depuration of microcystins. *J. Exp. Zool. A Ecol. Genet. Physiol.* 311, 226–230.
- Fischer, W.J., Dietrich, D.R., 2000. Pathological and biochemical characterization of microcystin-induced hepatopancreas and kidney damage in carp (*Cyprinus carpio*). *Toxicol. Appl. Pharmacol.* 164, 73–81.
- Fogg, G.E., Stewart, W.D.P., Fay, P., Walsby, A.E., 1973. *The Blue-Green Algae*. Academic Press, London.
- Gantar, M., Svirčev, Z., 2008. Microalgae and cyanobacteria: food for thought. *J. Phycol.* 44, 260–268.
- Gemma, S., Molteni, M., Rossetti, C., 2016. Lipopolysaccharides in cyanobacteria: a brief overview. *Advances in Microbiology* 6, 391–397.
- Ghazali, I.E., Saqrane, S., Carvalho, A.P., Ouahid, Y., Del Campo, F.F., Vasconcelos, V., Oudra, B., 2010. Effects of the microcystin profile of a cyanobacterial bloom on growth and toxin accumulation in common carp *Cyprinus carpio* larvae. *J. Fish Biol.* 76, 1415–1430.
- Gouveia, L., Batista, A.P., Sousa, I., Raymundo, A., Bandarra, N.M., 2008. Chapter 2. Microalgae in novel food products. In: Papadopoulos, K.N. (Ed.), *Food Chemistry Research Developments*. Nova Science Publishers, Inc., Hauppauge NY, USA, pp. 75–112.
- Gu, H.X., Hu, M.H., Wei, S.S., Kong, H., Huang, X.Z., Bao, Y.B., Wang, Y.J., 2019. Combined effects of toxic *Microcystis aeruginosa* and hypoxia on the digestive enzyme activities of the triangle sail mussel *Hyriopsis cumingii*. *Aquat. Toxicol.* 212, 241–246.
- Guo, L.G., Ke, Z.X., Xie, P., Ni, L.Y., 2009. Food consumption by *In Situ* pen-cultured planktivorous fishes and effects on an algal bloom in Lake Taihu, China. *J. Freshw. Ecol.* 24, 135–143.
- Guo, L.G., Wang, Q., Xie, P., Tao, M., Zhang, J., Niu, Y., Ma, Z.M., 2015. A non-classical biomanipulation experiment in Gonghu Bay of Lake Taihu: control of *Microcystis* blooms using silver and bighead carp. *Aquac. Res.* 46, 2211–2224.
- Guzmán-Guillén, R., Puerto, M., Gutiérrez-Praena, D., Prieto, A.I., Pichardo, S., Jos, Á., Campos, A., Vasconcelos, V., Cameán, A.M., 2017. Potential use of chemoprotectants against the toxic effects of cyanotoxins: a review. *Toxins* 9, 175.
- Ha, J.H., Hidaka, T., Tsuno, H., 2009. Quantification of toxic *Microcystis* and evaluation of its dominance ratio in blooms using real-time PCR. *Environmental Science & Technology* 43, 812–818.
- Han, S.Q., Zhou, Q., Xu, Y.D., Vanogtrop, F., Guo, Q.J., Liu, G.F., Yan, S.H., 2015. Valuable ingredients and feed toxicity evaluation of *Microcystis aeruginosa* acidolysis product in mice. *Exp. Biol. Med.* 240, 1333–1339.
- Harke, M.J., Steffen, M.M., Gobler, C.J., Otten, T.G., Wilhelm, S.W., Wood, S.A., Paerl, H.W., 2016. A review of the global ecology, genomics, and biogeography of the toxic cyanobacterium, *Microcystis* spp. *Harmful Algae* 54, 4–20.
- He, X.X., Liu, Y.-L., Conklin, A., Westrick, J., Weavers, L.K., Dionysiou, D.D., Lenhart, J.J., Mouser, P.J., Szlag, D., Walker, H.W., 2016. Toxic cyanobacteria and drinking water: impacts, detection, and treatment. *Harmful Algae* 54, 174–193.
- Hu, M.H., Wu, F.L., Yuan, M.Z., Li, Q.Z., Gu, Y.D., Wang, Y.J., Liu, Q.G., 2015. Antioxidant responses of triangle sail mussel *Hyriopsis cumingii* exposed to harmful algae *Microcystis aeruginosa* and hypoxia. *Chemosphere* 139, 541–549.
- Hu, Y.F., Chen, J., Fan, H.H., Xie, P., He, J., 2016. A review of neurotoxicity of microcystins. *Environ. Sci. Pollut. Res.* 23, 7211–7219.
- Hughes, W.W., 1996. *Essentials of Environmental Toxicology*. Taylor & Francis, London, p. 176.
- Huisman, J., Codd, G.A., Paerl, H.W., Ibelings, B.W., Verspagen, J.M.H., Visser, P.M., 2018. Cyanobacterial blooms. *Nat. Rev. Microbiol.* 16, 471–483.
- IARC (International Agency for Research on Cancer), 2010. *IARC Monographs on the Evaluation of Carcinogenic Risks to Humans, Ingested Nitrate and Nitrite and Cyanobacterial Peptide Toxins*. vol. 94. International Agency for Research on Cancer, Lyon.
- Islam, M.S., Drasar, B.S., Bradley, D.J., 1990. Long-term persistence of toxigenic *Vibrio cholerae* O1 in the mucilaginous sheath of a blue-green alga, *Anabaena variabilis*. *The Journal of Tropical Medicine and Hygiene* 93, 133–139.
- Islam, M.S., Miah, M.A., Hasan, M.K., Sack, R.B., Albert, M.J., 1994. Detection of non-culturable *Vibrio cholerae* O1 associated with a cyanobacterium from an aquatic environment in Bangladesh. *Trans. R. Soc. Trop. Med. Hyg.* 88, 298–299.
- Islam, M.S., Zaman, M.H., Islam, M.S., Ahmed, N., Clemens, J.D., 2020. Environmental reservoirs of *Vibrio cholerae*. *Vaccine* 38, A52–A62.
- Janssen, E.M.-L., 2019. Cyanobacterial peptides beyond microcystins - a review on co-occurrence, toxicity, and challenges for risk assessment. *Water Res.* 151, 488–499.
- Jaquet, J., 1976. Microflora of *Spirulina* preparations (in French). *Annales de la Nutrition et de l'Alimentation* 29, 589–601.

- Jassby, A., 1988. Chapter 8: some public health aspects of microalgal products. In: Lambi, C.A., Waaland, J.R. (Eds.), *Algae and Human Affairs*. Cambridge University Press, Cambridge, UK, pp. 181–202.
- Jia, Y.L., Chen, W., Zuo, Y.X., Lin, L.Z., Song, L.R., 2018. Heavy metal migration and risk transference associated with cyanobacterial blooms in eutrophic freshwater. *Sci. Total Environ.* 613, 1324–1330.
- Jia, Y.L., Chen, Q.Q., Crawford, S.E., Song, L.R., Chen, W., Hammers-Wirtz, M., Strauss, T., Seiler, T.-B., Schäffer, A., Hollert, H., 2019. Cyanobacterial blooms act as sink and source of endocrine disruptors in the third largest freshwater lake in China. *Environ. Pollut.* 245, 408–418.
- Jin, X.-C., Chu, Z.-S., Yi, W.-L., Hu, X.-Z., 2005. Influence of phosphorus on *Microcystis* growth and the changes of other environmental factors. *J. Environ. Sci.* 17, 937–941.
- Jos, Á., Pichardo, S., Prieto, A.I., Repetto, G., Vázquez, C.M., Moreno, I., Cameán, A.M., 2005. Toxic cyanobacterial cells containing microcystins induce oxidative stress in exposed tilapia fish (*Oreochromis* sp.) under laboratory conditions. *Aquat. Toxicol.* 72, 261–271.
- Juhel, G., Davenport, J., O'halloran, J., Culloty, S.C., O'Riordan, R.M., James, K.F., Furey, A., Allis, O., 2006. Impacts of microcystins on the feeding behaviour and energy balance of zebra mussels, *Dreissena polymorpha*: a bioenergetics approach. *Aquat. Toxicol.* 79, 391–400.
- Juhel, G., O'Halloran, J., Culloty, S.C., O'Riordan, R.M., Davenport, J., O'Brien, N.M., James, K.F., Furey, A., Allis, O., 2007. *In vivo* exposure to microcystins induces DNA damage in the haemocytes of the zebra mussel, *Dreissena polymorpha*, as measured with the comet assay. *Environ. Mol. Mutagen.* 48, 22–29.
- Juhel, G., Ramsay, R.M., Davenport, J., O'halloran, J., Culloty, S.C., 2015. Effect of the microcystin-producing cyanobacterium, *Microcystis aeruginosa*, on immune functions of the zebra mussel *Dreissena polymorpha*. *J. Shellfish Res.* 34, 433–443.
- Júnior, O.R.P., de Oliveira, N.B., Bosque, R.J., Ferreira, M.F.N., da Silva, V.M.A., Magalhães, A.C.M., de Santana, C.J.C., de Souza Castro, M., 2018. Histopathological evaluation of the exposure by cyanobacteria culture containing [d-Leu1] microcystin-LR on *Lithobates catesbeianus* tadpoles. *Toxins* 10, 318.
- Kamjunke, N., Mendonca, R., Hardeewig, I., Mehner, T., 2002a. Assimilation of different cyanobacteria as food and the consequences for internal energy stores of juvenile roach. *J. Fish Biol.* 60, 731–738.
- Kamjunke, N., Schmidt, K., Pflugmacher, S., Mehner, T., 2002b. Consumption of cyanobacteria by roach (*Rutilus rutilus*): useful or harmful to the fish? *Freshw. Biol.* 47, 243–250.
- Ke, Z.X., Xie, P., Guo, L.G., Liu, Y.Q., Yang, H., 2007. *In situ* study on the control of toxic *Microcystis* blooms using phytoplanktivorous fish in the subtropical Lake Taihu of China: a large fish pen experiment. *Aquaculture* 265, 127–138.
- Ke, Z., Xie, P., Guo, L., 2008. *In situ* study on effect of food competition on diet shifts and growth of silver and bighead carps in large biomanipulation fish pens in Meiliang Bay, Lake Taihu. *J. Appl. Ichthyol.* 24, 263–268.
- Keshavanath, P., Beveridge, M.C.M., Baird, D.J., Lawton, L.A., Nimmo, A., Codd, G.A., 1994. The functional grazing response of a phytoplanktivorous fish *Oreochromis niloticus* to mixtures of toxic and non-toxic strains of the cyanobacterium *Microcystis aeruginosa*. *J. Fish Biol.* 45, 123–129.
- Khairy, H.M., Ibrahim, M.A., Ibrahim, M.D., 2012. Phenological and liver antioxidant profiles of adult Nile tilapia (*Oreochromis niloticus*) exposed to toxic live cyanobacterium (*Microcystis aeruginosa* Kützinger) cells. *Zeitschrift für Naturforschung C* 67, 620–628.
- Kong, C., Adeola, O., 2014. Evaluation of amino acid and energy utilization in feedstuff for swine and poultry diets. *Asian Australas. J. Anim. Sci.* 27, 917–925.
- Konst, H., Mc Kercher, P.D., Gorham, P.R., Robertson, A., Howell, J., 1965. Symptoms and pathology produced by toxic *Microcystis aeruginosa* NRC-1 in laboratory and domestic animals. *Canadian Journal of Comparative Medicine and Veterinary Science* 29, 221–228.
- Krüger, G.H.J., De Wet, H., Kock, J.L.F., Pieterse, A.J.H., 1995. Fatty acid composition as a taxonomic characteristic for *Microcystis* and other coccoid cyanobacteria (blue-green algae) isolates. *Hydrobiologia* 308, 145–151.
- Li, X.-Y., Chung, I.-K., Kim, J.-I., Lee, J.-A., 2004. Subchronic oral toxicity of microcystin in common carp (*Cyprinus carpio* L.) exposed to *Microcystis* under laboratory conditions. *Toxicol.* 44, 821–827.
- Li, X.-Y., Chung, I.-K., Kim, J.-I., Lee, J.-A., 2005. Oral exposure to *Microcystis* increases activity-augmented antioxidant enzymes in the liver of loach (*Misgurnus mizolepis*) and has no effect on lipid peroxidation. *Comparative Biochemistry and Physiology, Part C* 141, 292–296.
- Li, L., Xie, P., Chen, J., 2007. Biochemical and ultrastructural changes of the liver and kidney of the phytoplanktivorous silver carp feeding naturally on toxic *Microcystis* blooms in Taihu Lake, China. *Toxicol.* 49, 1042–1053.
- Li, L., Xie, P., Guo, L.G., Ke, Z.X., Zhou, Q., Liu, Y.Q., Qiu, T., 2008. Field and laboratory studies on pathological and biochemical characterization of microcystin-induced liver and kidney damage in the phytoplanktivorous bighead carp. *Sci. World J.* 8, 121–137.
- Li, M., Nkrumah, P.N., Xiao, M., 2014. Biochemical composition of *Microcystis aeruginosa* related to specific growth rate: insight into the effects of abiotic factors. *Inland Waters* 4, 357–362.
- Liang, H.L., Zhou, W.S., Zhang, Y.L., Qiao, Q., Zhang, X.Z., 2015. Are fish fed with cyanobacteria safe, nutritious and delicious? A laboratory study. *Sci. Rep.* 5, 15166.
- Lu, K.H., Jin, C.H., Dong, S.L., Gu, B.H., Bowen, S.H., 2006. Feeding and control of blue-green algal blooms by tilapia (*Oreochromis niloticus*). *Hydrobiologia* 568, 111–120.
- Lum, K.K., Kim, J., Lei, X.G., 2013. Dual potential of microalgae as a sustainable biofuel feedstock and animal feed. *Journal of Animal Science and Biotechnology* 4, 53.
- McLellan, N.L., Manderville, R.A., 2017. Toxic mechanisms of microcystins in mammals. *Toxicology Research* 6, 391–405.
- Metcalfe, J.S., Codd, G.A., 2012. Chapter 24: cyanotoxins. In: Whitton, B.A. (Ed.), *Ecology of Cyanobacteria II. Their Diversity in Space and Time*. Springer, Dordrecht, pp. 651–675.
- Metcalfe, J.S., Beattie, K.A., Pflugmacher, S., Codd, G.A., 2000. Immuno-crossreactivity and toxicity assessment of conjugation products of the cyanobacterial toxin, microcystin-LR. *FEMS Microbiol. Lett.* 189, 155–158.
- Metcalfe, J.S., Beattie, K.A., Ressler, J., Gerbersdorf, S., Pflugmacher, S., Codd, G.A., 2002. Cross-reactivity and performance assessment of four microcystin immunoassays with detoxication products of the cyanobacterial toxin microcystin-LR. *J. Water Supply Res. Technol. AQUA* 51, 145–151.
- Mišurcová, L., Buňka, F., Ambrožová, J.V., Machů, L., Samek, D., Kráčmar, S., 2014. Amino acid composition of algal products and its contribution to RDI. *Food Chem.* 151, 120–125.
- Molina, R., Moreno, I., Pichardo, S., Jos, A., Moyano, R., Monterde, J.G., Cameán, A., 2005. Acid and alkaline phosphatase activities and pathological changes induced in Tilapia fish (*Oreochromis* sp.) exposed subchronically to microcystins from toxic cyanobacterial blooms under laboratory conditions. *Toxicol.* 46, 725–735.
- Monteiro, S., Santos, R., Blaha, L., Codd, G.A., 2017. Chapter 17: lipopolysaccharide endotoxins. In: Meriluoto, J., Spoof, L., Codd, G.A. (Eds.), *Handbook of Cyanobacterial Monitoring and Cyanotoxin Analysis*. Wiley, Chichester UK, pp. 164–172.
- Morales-Covarrubias, M.S., Alonso-Rodríguez, R., Velázquez-Garay, J.A., Flores-Chavarría, A.M.G., Alva-Martínez, A.F., 2016. Effect of *Microcystis aeruginosa* on organs and tissues of white shrimp (*Litopenaeus vannamei*) postlarvae in low-salinity water. *Revista Científica* 26, 181–187.
- Oliveira, F., Diez-Quijada, L., Turkina, M.V., Morais, J., Felpeto, A.B., Azevedo, J., Jos, A., Camean, A.M., Vasconcelos, V., Martins, J.C., Campos, A., 2020. Physiological and metabolic responses of marine mussels exposed to toxic cyanobacteria *Microcystis aeruginosa* and *Chrysochloris ovalisporum*. *Toxins* 12, 196.
- Orr, P.T., Jones, G.J., Hunter, R.A., Berger, K., De Paoli, D.A., Orr, C.L.A., 2001. Ingestion of toxic *Microcystis aeruginosa* by dairy cattle and the implications for microcystin contamination of milk. *Toxicol.* 39, 1847–1854.
- Orr, P.T., Jones, G.J., Hunter, R.A., Berger, K., 2003. Exposure of beef cattle to sub-clinical doses of *Microcystis aeruginosa*: toxin bioaccumulation, physiological effects and human health risk assessment. *Toxicol.* 41, 613–620.
- Osswald, J., Rellán, S., Gago-Martínez, A., Vasconcelos, V., 2009. Production of anatoxin-a by cyanobacterial strains isolated from Portuguese fresh water systems. *Ecotoxicology* 18, 1110–1115.
- Otten, T.G., Paerl, H.W., Dreher, T.W., Kimmerer, W.J., Parker, A.E., 2017. The molecular ecology of *Microcystis* sp. blooms in the San Francisco Estuary. *Environ. Microbiol.* 19, 3619–3637.
- Paerl, H.W., 1976. Specific associations of the blue-green algae *Anabaena* and *Aphanizomenon* with bacteria in freshwater blooms. *J. Phycol.* 12, 431–435.
- Paerl, H.W., 1978. Microbial organic carbon recovery in aquatic ecosystems. *Limnol. Oceanogr.* 23, 927–935.
- Palikova, M., Ondrackova, P., Mares, J., Adamovsky, O., Pikula, J., Kohoutek, J., Navratil, S., Blaha, L., Kopp, R., 2013. *In vivo* effects of microcystins and complex cyanobacterial biomass on rats (*Rattus norvegicus* var. alba): changes in immunological and haematological parameters. *Toxicol.* 73, 1–8.
- Park, H.-D., Watanabe, M.F., 1995. Chapter: toxic *Microcystis* in eutrophic lakes. In: Watanabe, M.F., K-i Harada, Carmichael, W.W., Fujita, H. (Eds.), *Toxic Microcystis*. CRC Press, Boca Raton, Florida, pp. 57–77.
- Park, H.D., Watanabe, M.F., Harada, K.I., Nagai, H., Suzuki, M., Watanabe, M., Hayashi, H., 1993. Hepatotoxic (microcystin) and neurotoxin (anatoxin-a) contained in natural blooms and strains of cyanobacteria from Japanese freshwaters. *Nat. Toxins* 1, 353–360.
- Pašková, V., Adamovský, O., Pikula, J., Skočovská, B., Band'ouchová, H., Horáková, J., Babica, P., Maršálek, B., Hilscherová, K., 2008. Detoxification and oxidative stress responses along with microcystins accumulation in Japanese quail exposed to cyanobacterial biomass. *Sci. Total Environ.* 398, 34–47.
- Paskova, V., Paskerova, H., Pikula, J., Bandouchova, H., Sedlackova, J., Hilscherova, K., 2011. Combined exposure of Japanese quails to cyanotoxins, Newcastle virus and lead: oxidative stress responses. *Ecotoxicol. Environ. Saf.* 74, 2082–2090.
- Pavagadhi, S., Balasubramanian, R., 2013. Toxicological evaluation of microcystins in aquatic fish species: current knowledge and future directions. *Aquat. Toxicol.* 142–143, 1–16.
- Pikula, J., Bandouchova, H., Hilscherova, K., Paskova, V., Sedlackova, J., Adamovsky, O., Knotkova, Z., Lany, P., Machat, J., Marsalek, B., Novotny, L., Pohanka, M., Vitula, F., 2010. Combined exposure to cyanobacterial biomass, lead and the Newcastle virus enhances avian toxicity. *Sci. Total Environ.* 408, 4984–4992.
- Piorreck, M., Baasch, K.-H., Pohl, P., 1984. Biomass production, total protein, chlorophylls, lipids and fatty acids of freshwater green and blue-green algae under different nitrogen regimes. *Phytochemistry* 23, 207–216.
- Polyak, Y., Zaytseva, T., Medvedeva, N., 2013. Response of toxic cyanobacterium *Microcystis aeruginosa* to environmental pollution. *Water Air Soil Pollut.* 224, 1494.
- Preeti, T., Hariharan, G., Rajarajeswari, G.R., 2016. Histopathological and biochemical effects of cyanobacterial cells containing microcystin-LR on Tilapia fish. *Water and Environment Journal* 30, 135–142.
- Prieto, A.I., Jos, Á., Pichardo, S., Moreno, I., Cameán, A.M., 2008. Protective role of vitamin E on the microcystin-induced oxidative stress in tilapia fish (*Oreochromis niloticus*). *Environ. Toxicol. Chem.* 27, 1152–1159.
- Prieto, A.I., Jos, Á., Pichardo, S., Moreno, I., de Sotomayor, M.Á., Moyano, R., Blanco, A., Cameán, A.M., 2009. Time-dependent protective efficacy of Trolox (vitamin E analog) against microcystin-induced toxicity in tilapia (*Oreochromis niloticus*). *Environ. Toxicol.* 24, 563–579.
- Pryor, G.S., 2003. Growth rates and digestive abilities of bullfrog tadpoles (*Rana catesbeiana*) fed algal diets. *J. Herpetol.* 37, 560–567.
- Puerto, M., Prieto, A.I., Pichardo, S., Moreno, I., Jos, Á., Moyano, R., Cameán, A.M., 2009. Effects of dietary N-acetylcysteine on the oxidative stress induced in tilapia

- (*Oreochromis niloticus*) exposed to a microcystin-producing cyanobacterial water bloom. *Environ. Toxicol. Chem.* 28, 1679–1686.
- Puerto, M., Prieto, A.I., Jos, A., Moreno, I., Moyano, R., Blanco, A., Cameán, A.M., 2010. Dietary N-acetylcysteine (NAC) prevents histopathological changes in tilapias (*Oreochromis niloticus*) exposed to a microcystin-producing cyanobacterial water bloom. *Aquaculture* 306, 35–48.
- Puerto, M., Gutiérrez-Praena, D., Prieto, A.I., Pichardo, S., Jos, A., Miguel-Carrasco, J.L., Vazquez, C.M., Cameán, A.M., 2011. Subchronic effects of cyanobacterial cells on the transcription of antioxidant enzyme genes in tilapia (*Oreochromis niloticus*). *Ecotoxicology* 20, 479–490.
- Qiao, Q., Liang, H.L., Zhang, X.Z., 2013. Effect of cyanobacteria on immune function of crucian carp (*Carassius auratus*) via chronic exposure in diet. *Chemosphere* 90, 1167–1176. <https://doi.org/10.1016/j.chemosphere.2012.09.025>.
- Qiao, F., Lei, K., Li, Z.C., Wei, Z.L., Liu, Q., Yang, L.B., He, J.W., An, L.H., Qi, H.L., Cui, S., 2018. Transcriptomic responses of the freshwater snail (*Parafossarulus striatulus*) following dietary exposure to cyanobacteria. *Sci. Total Environ.* 624, 153–161.
- Qiu, T., Xie, P., Ke, Z.X., Li, L., Guo, L.G., 2007. *In situ* studies on physiological and biochemical responses of four fishes with different trophic levels to toxic cyanobacterial blooms in a large Chinese lake. *Toxicol. Environ. Chem.* 50, 365–376.
- Rinta-Kanto, J.M., Ouellette, A.J.A., Boyer, G.L., Twiss, M.R., Bridgeman, T.B., Wilhelm, S.W., 2005. Quantification of toxic *Microcystis* spp. during the 2003 and 2004 blooms in western Lake Erie using quantitative real-time PCR. *Environmental Science & Technology* 39, 4198–4205.
- Sabatini, S.E., Brena, B.M., Pirez, M., de Molina, M.C.R., Luquet, C.M., 2015. Oxidative effects and toxin bioaccumulation after dietary microcystin intoxication in the hepatopancreas of the crab *Neohelice (Chasmagnathus) granulata*. *Ecotoxicol. Environ. Saf.* 120, 136–141.
- Sathasivam, R., Radhakrishnan, R., Hashem, A., Abd_Allah, E.F., 2019. Microalgae metabolites: a rich source for food and medicine. *Saudi Journal of Biological Sciences* 26, 709–722.
- Šejnová, L., Maršálek, B., 2012. Chapter 7: *Microcystis*. In: Whitton, B.A. (Ed.), *Ecology of Cyanobacteria II. Their Diversity in Space and Time*. Springer, Dordrecht, pp. 195–228.
- Servaites, J.C., Faeth, J.L., Sidhu, S.S., 2012. A dye binding method for measurement of total protein in microalgae. *Anal. Biochem.* 421, 75–80.
- Sharathchandra, K., Rajashekhar, M., 2011. Total lipid and fatty acid composition in some freshwater cyanobacteria. *J. Algal Biomass Utilization* 2, 83–97.
- Shi, W., Yu, N.Y., Jiang, X., Han, Z.H., Wang, S.H., Zhang, X.W., Wei, S., Giesy, J.P., Yu, H.X., 2017. Influence of blooms of phytoplankton on concentrations of hydrophobic organic chemicals in sediments and snails in a hyper-eutrophic, freshwater lake. *Water Res.* 113, 22–31.
- Skocovska, B., Hilscherova, K., Babica, P., Adamovsky, O., Bandouchova, H., Horakova, J., Knotkova, Z., Marsalek, B., Paskova, V., Pikula, J., 2007. Effects of cyanobacterial bloom on the Japanese quail. *Toxicol. Environ. Chem.* 49, 793–803.
- Smith, J.L., Boyer, G.L., Zimba, P.V., 2008. A review of cyanobacterial odorous and bioactive metabolites: impacts and management alternatives in aquaculture. *Aquaculture* 280, 5–20.
- Spoof, L., Catherine, A., 2017. Appendix 3: tables of microcystins and nodularins. In: Meriluoto, J., Spoof, L., Codd, G.A. (Eds.), *Handbook of Cyanobacterial Monitoring and Cyanotoxin Analysis*. Wiley, Chichester, UK, pp. 527–537.
- Stewart, I., Schluter, P.J., Shaw, G.R., 2006. Cyanobacterial lipopolysaccharides and human health—a review. *Environ. Health* 5, 7.
- Svirčev, Z., Drobac, D., Tokodi, N., Mijović, B., Codd, G.A., Meriluoto, J., 2017. Toxicology of microcystins with reference to cases of human intoxications and epidemiological investigations of exposures to cyanobacteria and cyanotoxins. *Arch. Toxicol.* 91, 621–650.
- Svirčev, Z., Lalić, D., Savić, G.B., Tokodi, N., Backović, D.D., Chen, L., Meriluoto, J., Codd, G.A., 2019. Global geographical and historical overview of cyanotoxin distribution and cyanobacterial poisonings. *Arch. Toxicol.* 93, 2429–2481.
- Tencalla, F.G., Dietrich, D.R., Schlatter, C., 1994. Toxicity of *Microcystis aeruginosa* peptide toxin to yearling rainbow trout (*Oncorhynchus mykiss*). *Aquat. Toxicol.* 30, 215–224.
- Tibbetts, S.M., Milley, J.E., Lall, S.P., 2015. Chemical composition and nutritional properties of freshwater and marine microalgal biomass cultured in photobioreactors. *J. Appl. Phycol.* 27, 1109–1119.
- Torres-Tijj, Y., Fields, F.J., Mayfield, S.P., 2020. Microalgae as a future food source. *Biotechnol. Adv.* 41, 107536.
- Vadde, K.K., Feng, Q.L., Wang, J.J., McCarthy, A.J., Sekar, R., 2019. Next-generation sequencing reveals fecal contamination and potentially pathogenic bacteria in a major inflow river of Taihu Lake. *Environ. Pollut.* 254, 113108.
- van Apeldoorn, M.E., van Egmond, H.P., Speijers, G.J.A., Bakker, G.J.I., 2007. Toxins of cyanobacteria. *Mol. Nutr. Food Res.* 51, 7–60.
- Vanthoor-Koopmans, M., Cordoba-Matson, M.V., Arredondo-Vega, B.O., Lozano-Ramírez, C., García-Trejo, J.F., Rodríguez-Palacio, M.C., 2014. Chapter 8: microalgae and cyanobacteria production for feed and food supplements. In: Guevara-Gonzalez, R., Torres-Pacheco, I. (Eds.), *Biosystems Engineering: Biofactories for Food Production in the Century XXI*. Springer, Cham, pp. 253–275.
- Vasconcelos, V., Oliveira, S., Teles, F.O., 2001. Impact of a toxic and a non-toxic strain of *Microcystis aeruginosa* on the crayfish *Procambarus clarkii*. *Toxicol. Environ. Chem.* 39, 1461–1470.
- Villarruel-López, A., Ascencio, F., Nuño, K., 2017. Microalgae, a potential natural functional food source - a review. *Polish Journal of Food and Nutrition Sciences* 67, 251–263.
- Walsh, K., Jones, G.J., Dunstan, R.H., 1997. Effect of irradiance on fatty acid, carotenoid, total protein composition and growth of *Microcystis aeruginosa*. *Phytochemistry* 44, 817–824.
- Wang, L., Chen, C.Y., Liu, W.J., Xia, H., Li, J., Zhang, X.Z., 2017. Effects of toxic cyanobacteria and ammonia on flesh quality of blunt snout bream (*Megalobrama amblycephala*). *J. Sci. Food Agric.* 97, 1200–1206.
- Ward, C.J., Codd, G.A., 1999. Comparative toxicity of four microcystins of different hydrophobicities to the protozoan *Tetrahymena pyriformis*. *J. Appl. Microbiol.* 86, 874–882.
- Whitton, B.A. (Ed.), 2012. *Ecology of Cyanobacteria II. Their Diversity in Space and Time*. Springer, Dordrecht.
- WHO (World Health Organization), 1998. *Guidelines for Drinking-Water Quality - Second Edition - Volume 2 - Health Criteria and Other Supporting Information - Addendum*. World Health Organization, Geneva.
- Wood, S.A., Rueckert, A., Hamilton, D.P., Cary, S.C., Dietrich, D.R., 2011. Switching toxin production on and off: intermittent microcystin synthesis in a *Microcystis* bloom. *Environ. Microbiol. Rep.* 3, 118–124.
- Wu, X.Q., Jiang, J.Q., Wan, Y., Giesy, J.P., Hu, J.Y., 2012. Cyanobacteria blooms produce teratogenic retinoic acids. *Proc. Natl. Acad. Sci.* 109, 9477–9482.
- Wu, F.L., Kong, H., Shang, Y.Y., Zhou, Z.Q., Gul, Y., Liu, Q.G., Hu, M.H., 2017. Histopathological alterations in triangle sail mussel (*Hyriopsis cumingii*) exposed to toxic cyanobacteria (*Microcystis aeruginosa*) under hypoxia. *Aquaculture* 467, 182–189.
- Xia, H., Song, T., Wang, L., Jiang, L.S., Zhou, Q.T., Wang, W.M., Liu, L.G., Yang, P.H., Zhang, X.Z., 2018. Effects of dietary toxic cyanobacteria and ammonia exposure on immune function of blunt snout bream (*Megalobrama amblycephala*). *Fish and Shellfish Immunology* 78, 383–391.
- Xie, P., Liu, J., 2001. Practical success of biomanipulation using filter-feeding fish to control cyanobacteria blooms: a synthesis of decades of research and application in a subtropical hypereutrophic lake. *Sci. World J.* 1, 337–356.
- Xie, L.Q., Xie, P., Ozawa, K., Honma, T., Yokoyama, A., Park, H.-D., 2004. Dynamics of microcystins-LR and-RR in the phytoplanktonic silver carp in a sub-chronic toxicity experiment. *Environ. Pollut.* 127, 431–439.
- Xie, L.Q., Xie, P., Guo, L.G., Li, L., Miyabara, Y., Park, H.-D., 2005. Organ distribution and bioaccumulation of microcystins in freshwater fish at different trophic levels from the eutrophic Lake Chaohu, China. *Environ. Toxicol.* 20, 293–300.
- Xie, L.Q., Hanyu, T., Futatsugi, N., Komatsu, M., Steinman, A.D., Park, H.-D., 2014. Inhibitory effect of naringin on microcystin-LR uptake in the freshwater snail *Sinotaita histrica*. *Environ. Toxicol. Pharmacol.* 38, 430–437.
- Yang, Z.Y., Wu, H.J., Li, Y., 2012. Toxic effect on tissues and differentially expressed genes in hepatopancreas identified by suppression subtractive hybridization of freshwater pearl mussel (*Hyriopsis cumingii*) following microcystin-LR challenge. *Environ. Toxicol.* 27, 393–403.
- Zang, X.M., Wang, L., Xiao, J.R., Zhang, H.Y., Li, L., Zhang, X.Z., 2020. Harvesting colonial *Microcystis flos-aquae* using two-stage filtration: influence of pre-filtration on harvesting performance. *Sep. Purif. Technol.* 245, 116736.
- Zhang, X., Xie, P., Huang, X.P., 2008. A review of nontraditional biomanipulation. *Sci. World J.* 8, 1184–1196.
- Zhang, D.W., Xie, P., Liu, Y.Q., Qiu, T., 2009. Transfer, distribution and bioaccumulation of microcystins in the aquatic food web in Lake Taihu, China, with potential risks to human health. *Sci. Total Environ.* 407, 2191–2199.
- Zhang, X.M., Song, L.R., Zhang, P.C., He, J.W., Liu, Y.D., Matsuura, H., Watanabe, M.M., 2012. Grazing on toxic cyanobacterial blooms by tadpoles of edible frog *Rana grylio*. *Phycol. Res.* 60, 20–26.
- Zhang, D.W., Deng, X.W., Xie, P., Chen, J., Guo, L.G., 2013. Risk assessment of microcystins in silver carp (*Hypophthalmichthys molitrix*) from eight eutrophic lakes in China. *Food Chem.* 140, 17–21.
- Zhao, M., Xie, S., Zhu, X., Yang, Y., Gan, L., Song, L., 2006a. Effect of inclusion of blue-green algae meal on growth and accumulation of microcystins in gibel carp (*Carassius auratus gibelio*). *J. Appl. Ichthyol.* 22, 72–78.
- Zhao, M., Xie, S.Q., Zhu, X.M., Yang, Y.X., Gan, N.Q., Song, L.R., 2006b. Effect of dietary cyanobacteria on growth and accumulation of microcystins in Nile tilapia (*Oreochromis niloticus*). *Aquaculture* 261, 960–966.
- Zhou, Q., Xie, P., Xu, J., Ke, Z.X., Guo, L.G., 2009. Growth and food availability of silver and bighead carps: evidence from stable isotope and gut content analysis. *Aquac. Res.* 40, 1616–1625.
- Zhou, J.H., Zhao, Z.W., Liu, J., Peng, W., Peng, X., Han, Y.T., Xiao, P., 2019. Removal of *Microcystis aeruginosa* and control of algal organic matter by potassium ferrate (VI) pre-oxidation enhanced Fe(II) coagulation. *Korean J. Chem. Eng.* 36, 1587–1594.
- Zhu, J.Y., Lu, K.H., Zhang, C.J., Liang, J.J., Hu, Z.Y., 2011. Biochemical and ultrastructural changes in the hepatopancreas of *Bellamyia aeruginosa* (Gastropoda) fed with toxic cyanobacteria. *Sci. World J.* 11, 2091–2105.
- Ziková, A., Trubiroha, A., Wiegand, C., Wuertz, S., Rennert, B., Pflugmacher, S., Kopp, A., Mareš, J., Kloas, W., 2010. Impact of microcystin containing diets on physiological performance of Nile tilapia (*Oreochromis niloticus*) concerning stress and growth. *Environ. Toxicol. Chem.* 29, 561–568.
- Ziková, A., Lorenz, C., Lutz, I., Pflugmacher, S., Kloas, W., 2013. Physiological responses of *Xenopus laevis* tadpoles exposed to cyanobacterial biomass containing microcystin-LR. *Aquat. Toxicol.* 128, 25–33.
- Zuo, Z.J., Ni, B.B., Yang, L., 2018. Production of primary metabolites in *Microcystis aeruginosa* in regulation of nitrogen limitation. *Bioresour. Technol.* 270, 588–595.