



OPEN

Influence of ligand's directional configuration, chrysenes as model compounds, on the binding activity with aryl hydrocarbon receptor

Taewoo Kim¹, Juyuan Zhen², Junghyun Lee¹, Robert Bauer², Changkeun Lee¹, Bong-Oh Kwon³, Keun Hwa Chae⁴, Seongjin Hong⁵, John P. Giesy^{6,7}, Gap Soo Chang²✉ & Jong Seong Kim¹✉

Understanding what and how physico-chemical factors of a ligand configure conditions for ligand-receptor binding is a key to accurate assessment of toxic potencies of environmental pollutants. We investigated influences of the dipole-driven orientation and resulting directional configuration of ligands on receptor binding activities. Using physico-chemical properties calculated by ab initio density functional theory, directional reactivity factors (DRF) were devised as main indicators of toxic potencies, linking molecular ligand-receptor binding to in vitro responses. The directional reactivity model was applied to predict variation of aryl hydrocarbon receptor-mediated toxic potencies among homologues of chrysene with structural modifications such as the numbers of constituent benzene rings, methylation and hydroxylation. Results of predictive models were consistent with empirical potencies determined by use of the H4IIE-*luc* transactivation bioassay. The experiment-free approach based on first principles provides an analytical framework for estimating molecular bioactivity in silico and complements conventional empirical approaches to studying molecular initiating events in adverse outcome pathways.

Protein receptors act as transcription factors that regulate expressions of genes and determine how cells in organisms respond to xenobiotics. When a receptor binds with exogenous substances (viz., ligand), ligand-receptor complexes are translocated to nuclei of cells where they regulate expressions of genes. In the case of xenobiotics, this includes up-regulation of genes coding for enzymes, such as cytochrome P450 1A1 (CYP1A1), which transforms xenobiotics¹. This receptor-mediated, enzyme regulation depends on ligand-specific binding affinity and especially since it is the first initiation event, provides the basis of toxicology. Although various receptors and signal transduction pathways are known, kinetics of how ligands interact with responsive receptors remains imperfect and is thus currently undergoing intensive research^{2,3}.

To understand such ligand-specific events involved in adverse outcome pathways (AOPs) or functioning of drugs, knowledge of affinities of binding to receptors is fundamental⁴. There are a number of nuclear receptors, conserved in a wide range of vertebrates with which contaminants or drugs can interact as agonists or antagonists⁵. Variations among amino acid sequences of the ligand binding domain can explain differences in potencies of chemicals and sensitivities among species⁶. Due to incomplete understanding of physico-chemical interactions involved in binding of ligands to protein receptors, accurate prediction of affinities of ligand-specific binding in addition to steric factors, predicting initiation of AOPs, and toxicities is challenging.

¹School of Earth and Environmental Sciences and Research Institute of Oceanography, Seoul National University, Seoul, Republic of Korea. ²Department of Physics and Engineering Physics, University of Saskatchewan, Saskatoon, SK, Canada. ³Department of Marine Biotechnology, Kunsan National University, Kunsan, Republic of Korea. ⁴Advanced Analysis Center, Korea Institute of Science and Technology, Seoul, Republic of Korea. ⁵Department of Ocean Environmental Sciences, Chungnam National University, Daejeon, Republic of Korea. ⁶Department of Veterinary Biomedical Sciences and Toxicology Centre, University of Saskatchewan, Saskatoon, SK, Canada. ⁷Environmental Sciences Department, Baylor University, Waco, TX, USA. ✉email: gapsoo.chang@usask.ca; jskocean@snu.ac.kr

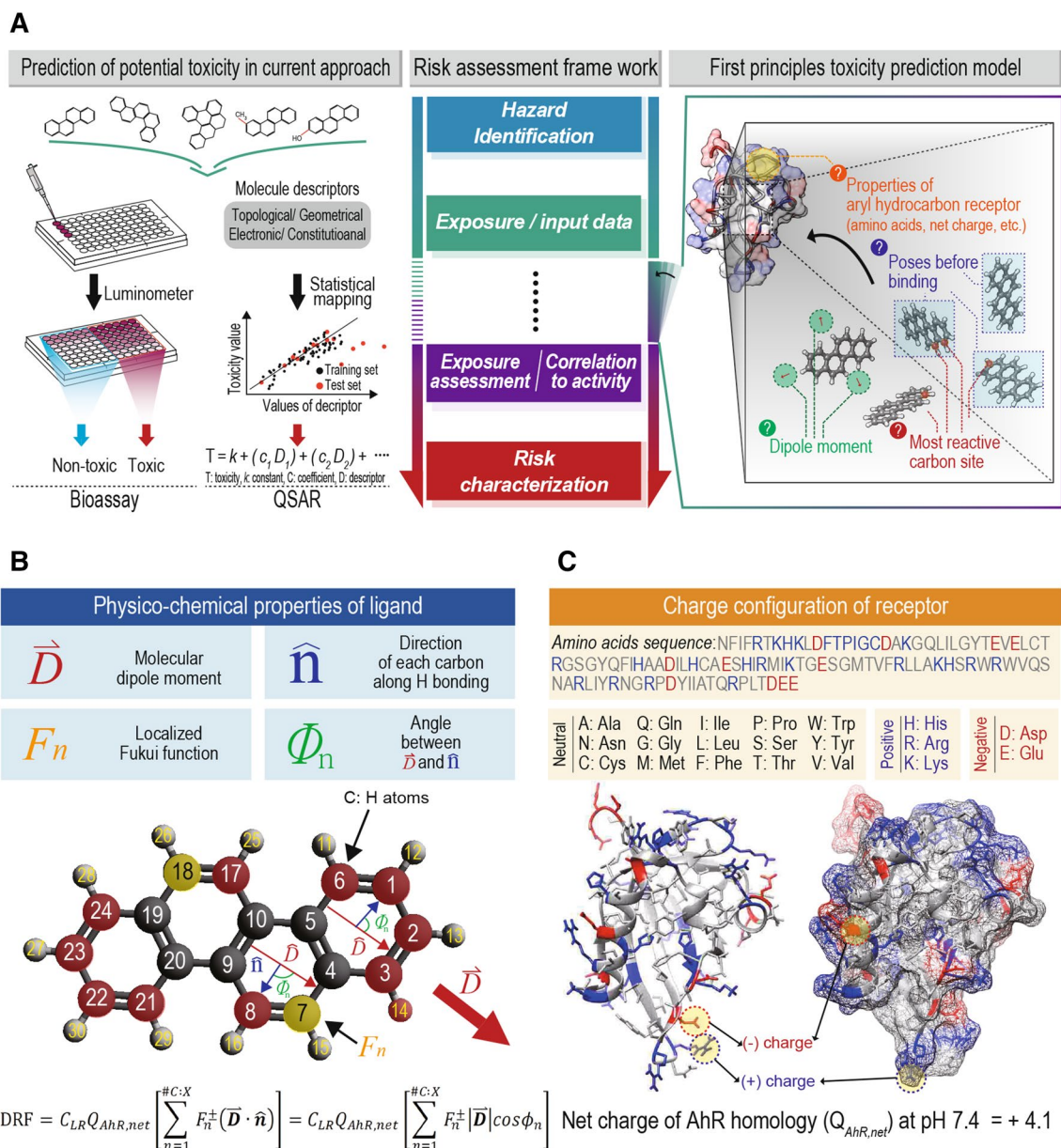


Figure 1. Directional reactive modeling for AhR mediated potency. **(A)** Comparison between contemporary experimental approaches, quantitative structure–activity relationship (QSAR) and first principles potential toxicity prediction model; **(B)** Physico-chemical properties of the model compound, chrysene, for the directional reactive factor (DRF); **(C)** Structural model of aryl hydrocarbon receptor (AhR) with ligand binding sites based on PDB ID: 4F3L and its amino acids sequence which is counted to 107. Net charge of AhR homology ($Q_{AhR,net}$) is determined to + 4.1.

In ecotoxicology, assessment of toxic potentials of biological responses through *in vivo* and later developed *in vitro* systems⁷. Aside from ethical issues involved in testing of whole animals and difficulty of studying threatened or endangered species or non-standard laboratory animals, the rapid upsurge in new chemicals either found in nature or synthesized in industry outpaces capacities of industries and governments to assess risks by use of traditional methods, which makes them impractical for routine screening and testing⁸. In an effort to complement traditional assays and efficiently screen new chemicals, this proliferation of new chemicals has led to increasing attention being given to *in silico* alternatives to empirical testing, especially *in vivo*⁹. Quantitative structure–activity relationships (QSAR) is one accepted alternative, *in silico* predictive method based on linear free energy models (LFEM) and statistical correlations between structure-related physico-chemical properties of ligands in form of molecular descriptors (numerical quantities) and previously experimented bioactivity data. Multiple linear regression equations developed by the QSAR approach are then used to extrapolate target endpoints, such as toxic potencies of new chemicals (Fig. 1A)^{10–12}. LFEMs include predictors of hydrophobic, electrical, and steric parameters. When structure-related properties are however, not linearly correlated to bioactivities, statistical inferences of target endpoints during semi-empirical

modeling are not applicable^{13–15}. Alternatively, recently, molecular docking models, based primarily on steric considerations, are being used to analytically estimate the binding affinity of ligands with receptors^{16,17}.

In a previous study, we introduced the directional reactivity factor to predict bioavailabilities of polychlorinated biphenyl congeners (PCBs) based on first principles. But physico-chemical information of participating receptors, such as electronic charge state was not considered in the predictive equation, which makes it limited for predicting receptor-dependent bioactivity¹⁸. Here, we report a novel, advanced directional reactive (DR) model that illustrates how physical properties of a ligand interplay to constitute optimal environments for biochemical interactions with respective receptors (Fig. 1A). The DR approach identifies three, stepwise key parameters and optimal conditions affecting ligand-receptor activity: (1) dipole-driven electrostatic interactions between ligands and receptors; (2) molecular orientation induced by physical ligand-receptor interaction and (3) resulting directional configuration of reactive sites in ligands. As a proof of concept, the DR model was applied to a ligand class containing homologues of polycyclic aromatic hydrocarbon (PAH), chrysene interacting with cytosolic protein, the aryl hydrocarbon receptor (AhR). Variation of the AhR-mediated toxic potencies across chrysene and its derivatives was well reproduced by the predictive model based on first-principles.

Results

Directional reactive modeling. For modeling interactions between ligands and receptor proteins, we explored two hypotheses; (a) different binding activities of various ligands with the same kind of receptor (AhR in this study) would solely depend on the ligand's physico-chemical properties and (b) different binding activities would be resulted from the interaction of ligands with common properties of the receptor. These hypotheses suggest that a proper identification of physical properties of the ligand compounds and their interplay configuring the optimal interaction conditions with the common receptor would provide a probabilistic estimation of the ligand-specific binding activity. Then, the estimated ligand-receptor binding activity can be verified by comparison with statistical observation of receptor-mediated toxic potency.

When a ligand (drug or xenobiotic) is introduced into the cytosol of cells, it encounters various forces such as electrostatic, hydrogen bonding, π - π stacking, and van der Waals forces¹⁹. Among these, electrostatic interaction is often the dominant force at large ligand-receptor separation before binding, while others are relatively weak or working over only small distances²⁰. This electrostatic interaction is mediated by dipole moments of ligands and charge state contributed from charged amino acids in a receptor. Considering larger molecular weights of receptor proteins than that of ligand molecule, charges on receptors can be assumed to be fixed in space and result in rotation of the ligand to make its dipole moment vector align along the line of force action, while the ligand moves toward the receptor. This process would determine the relative orientation of a ligand to "active binding" sites of the receptor. Another necessary condition to be considered for optimal reactions, resulting in ligand-receptor binding, is distribution of "frontier molecular orbitals (FMOs)" of a ligand, that is, the highest occupied molecular orbital (HOMO) and the lowest unoccupied molecular orbital (LUMO). Spatial distributions of HOMO and LUMO are related to charge densities in FMOs and thus, suggest which constituent atoms in a molecule serve as preferred sites for nucleophilic or electrophilic interactions with charged amino acids in a receptor. The DR model employs the atom-condensed Fukui function (F) as a FMO reactivity indicator. This approximation is valid for most cases of negligible orbital relaxation²¹. The optimal ligand-receptor reaction would then require the more reactive sites to be located at the front of the ligand, which is governed by the "dipole moment-driven alignment" (Fig. 1B). Considering these physico-chemical processes, we developed directional reactivity factors (DRFs), which describe favorable orientation configurations between ligands and receptors, given the charge-dipole interaction (Eq. 1).

$$\text{DRF} = C_{LR} Q_{R,\text{net}} \left[\sum_{n=1}^{\#A:X} F_n^{\pm} \left(\vec{D} \cdot \hat{n} \right) \right] = C_{LR} Q_{R,\text{net}} \left[\sum_{n=1}^{\#A:X} F_n^{\pm} \left| \vec{D} \right| \cos \phi_n \right] \quad (1)$$

where $Q_{R,\text{net}}$ is the receptor's net charge contributed from charged amino acids, F^+ (F^-) is the nucleophilic (electrophilic) Fukui function of each terminal atom bonded with hydrogen or other functional groups ($\#A:X$) in the ligand, \vec{D} is the ligand's dipole moment vector, and \hat{n} is the bonding direction of the terminal atom. The correlation coefficient C_{LR} is $1/4\pi\epsilon_0\epsilon_R r$ that represents Coulomb constant, relative permittivity of receptor cytosol (ϵ_R), and the inverse proportionality of charge-dipole interaction to the intermolecular distance (r). This coefficient becomes constant under the condition of equal ligand-receptor distance for the same type of receptor.

The DRF equation takes into account contributions of all atom-condensed Fukui functions along the line of action (\vec{D}). This describes reactive sites located behind the molecule ($\pi/2 < \phi_n < 3\pi/2$) with respect to the line of action that would reduce reaction affinity with the receptor. The use of either F^+ or F^- in the DRF calculation depends on the charged state of the responsive receptor. These quantities are computationally obtainable using the homology modeling method and the density functional theory (DFT)^{22,23}.

DRF as an indicator of ligand-binding reactivity. The DR model was applied to predict AhR mediated toxic potencies of homologues of chrysene by examining the DRF, which represents the degree of optimal reaction configuration toward ligand-receptor binding. Cytosolic AhR is a widely used ligand-activated transcript factor regulating expression of CYP450, xenobiotic metabolizing enzymes, which occur in diverse species and cell types²⁴. Chrysene is listed as one of 16 priority PAHs by the US Environmental Protection Agency (US EPA) and has a variety of homologues²⁵. The model compounds included 4 groups of differently modified chrysene homologues; benzo-, dibenzo-, methylated, and hydroxylated chrysenes. In each homologue group, 2–3 congeners were targeted to generally encompass homologous variations (Table S1 for full list of chrysene homologues). The net charge of AhR ($Q_{AhR,\text{net}}$) was obtained by building three-dimensional (3D) structures of

the ligand binding domain (LBD) from the sequence of amino acids in the rat AhR (GI: 7304873 in the NCBI sequence database) using SWISS-MODEL²⁶. The 3D AhR homology structure is composed of 107 amino acids in total and contains 5 histidines (His or H), 10 arginines (Arg or R), and 5 lysines (Lys or K), which possess positive charges at pH 7.4 (acidity condition of the in vitro bioassay in this study), and 5 aspartates (Asp or D) and 6 glutamates (Glu or E) of negatively charged acids (Fig. 1C and Figure S1). These charged amino acids result in a total net charge of +4.1 at pH 7.4 (Figure S2 and Table S2 for details of net charge calculation)²⁷.

Various physico-chemical properties of chrysenes, such as HOMO, LUMO, and molecular dipole moment, were determined by ab initio DFT calculations with Becke three-parameter Lee–Yang–Parr (B3LYP) exchange functional and the polarized triple-zeta valence (def2-TZVPP) basis set. The atom-condensed Fukui function was determined from the DFT-calculated Hirshfield charge population²⁸. Due to structural symmetry, when the chrysene base molecule is modified by electrophilic methyl and nucleophilic hydroxyl groups while the base molecule has a weak dipole moment almost perpendicular to the molecule plane, the dipole moment changes dramatically, both in direction and magnitude (Fig. 2A, Table 1, and Figure S3). The additive rings (benzo- and dibenzo-chrysenes) also result in rotation of dipole moment with small increase in magnitude. The hydroxy chrysene group has the largest dipole moment followed by the methyl group. The dipole moments of benzo and dibenzo groups are comparable to that of the base molecule. The electrophilic methyl and nucleophilic hydroxyl groups substantially increase the dipole moment magnitude. Alternatively, frontier orbitals (HOMO and LUMO) responsible for intermolecular interaction are additionally contributed from orbitals around carbon sites in benzo- and dibenzo-chrysenes (Fig. 2B and Figure S4).

Carbon in the methyl group does not possess LUMO and addition of the hydroxyl group suppress LUMO from nearby carbon atoms. The influence of molecular modification on frontier orbitals is supported by measurements of C 1s near-edge X-ray absorption fine structure (NEXAFS) (Fig. 2C). Two spectral features about 284.8 and 285.6 eV correspond to C 1s → LUMO transitions from carbon atoms bonded with hydrogen or other functional groups (peak a) and those bonded with neighboring carbon atoms (peak b)²⁹. Addition of a methyl group increases the spectral weight of a peak at the higher energy than LUMO of chrysene base molecule (284.8 eV) and thus the π^* state of carbon in the methyl group does not appear in the LUMO iso-surface. Suppression of the LUMO state due to the presence of the hydroxyl group is also supported by a decrease in the spectral weight of peak a, although there is no loss of a carbon atom through hydroxylation (Figure S5 for C 1s NEXAFS results for all chrysene homologues).

Using DFT-calculated properties and total net charge of the AhR, DRF values for all 12 homologues of chrysene were obtained (Table 1). The charge-dipole interaction would occur in the cell cytosol containing dissolved ions such as Na⁺, K⁺, and Cl⁻, not in free space³⁰. Results of a recent study suggested that these dissolved ions are attracted by and trapped in the vicinity of the charged amino acids in cytosolic proteins³¹. That is, Na⁺ and K⁺ ions are bound to negative Asp and Glu residues while Cl⁻ ions are to positive Arg, Lys, and His residues. After being introduced into the cell cytosol, the ligand thus recognizes the electrostatic potential by these ions surrounding the charged amino acids, which can be approximated to opposite charge of the AhR protein ($-Q_{AhR,net}$). The dipole moment vector and Fukui functions resulting in the DRFs are selected in this regard. The most striking result is that hydroxy-chrysene has the smallest DRF values despite their strongest dipole moments. This is because carbons with the large Fukui functions are located on the opposite side of the ligand from the receptor (Table 1 presents as an example, the location of carbon with the largest Fukui value). Methyl chrysene exhibited the largest DRF values, followed by benzo-chrysenes. In each homolog group, DRF values depend on the dipole moment and relative location of reactive carbons to the dipole orientation. The DRF shows the large value (27.97) for 2-methylchrysene (2MC) and decreases to 6.79 for 1-methylchrysene (1MC). In the case of benzo[c]chrysene (BcC), the DRF is larger than that of 1MC, because of the favorable distribution of reactive carbon sites despite the smaller F^+ and dipole moment.

AhR-mediated toxic potency of chrysene homologues. If the DR model describes the optimal configuration for ligand-receptor reaction properly, the DRF would be one promising indicator accurately estimating ligand-specific bioactivity. For robust validation of the DR model, a homogeneous set of AhR-mediated toxic potencies for chrysenes were examined experimentally by use of the in vitro, H4IIE-*luc* transactivation assay and compared with DRF values calculated from first-principles. The experimental dose–response curves for the groups of similarly structured chrysenes are presented (Fig. 3A). AhR-mediated potencies of all chrysene homologues are evidenced in the range of 51–147% BaP_{max}, which indicated significant potential to cause toxicity and suggests selected target chemicals are as potent as benzo[a]pyrene, which is a well-known in vitro agonist of the AhR. By group, methyl chrysene exhibits the greatest potency, followed by benzo-chrysene, while dibenzo-chrysene and hydroxy chrysene both exhibited similar or lesser potencies, compared to that of the base compound of unsubstituted chrysene.

For a detailed comparison with the DRF value for individual compounds, the half-maximal effective concentration (EC₅₀) values were calculated (Fig. 3B and Table 2). Among chrysene homologues tested, 2MC turns out to have the greatest toxic potency (EC₅₀ = 6.6 × 10⁻¹ μM) and 3-hydroxychrysene (3HC) was the least potent (EC₅₀ = 1400 × 10⁻¹ μM). Different EC₅₀ values were also observed within each homologue group. The EC₅₀ values of benzo[a]chrysene (BaC), benzo[b]chrysene (BbC), and BcC are 200 × 10⁻¹ μM, 28 × 10⁻¹ μM and 13 × 10⁻¹ μM, respectively, despite having the same molecular formula, mass, and octanol–water partition coefficient (Log K_{ow}; hydrophobicity). For the dibenzo-chrysene group, differences in EC₅₀ values were between dibenzo[b,def]chrysene (DbdC) (140 × 10⁻¹ μM) and dibenzo[def,p]chrysene (DdpC) (760 × 10⁻¹ μM). A similar trend was observed for the methyl- and hydroxy-chrysene groups. The experimentally derived bioactivity data was consistent with the optimal ligand-receptor reaction predicted by the DR model for toxicity rank among homologue groups and also among congeners within a homologue. This consistency suggests that accurate predictions

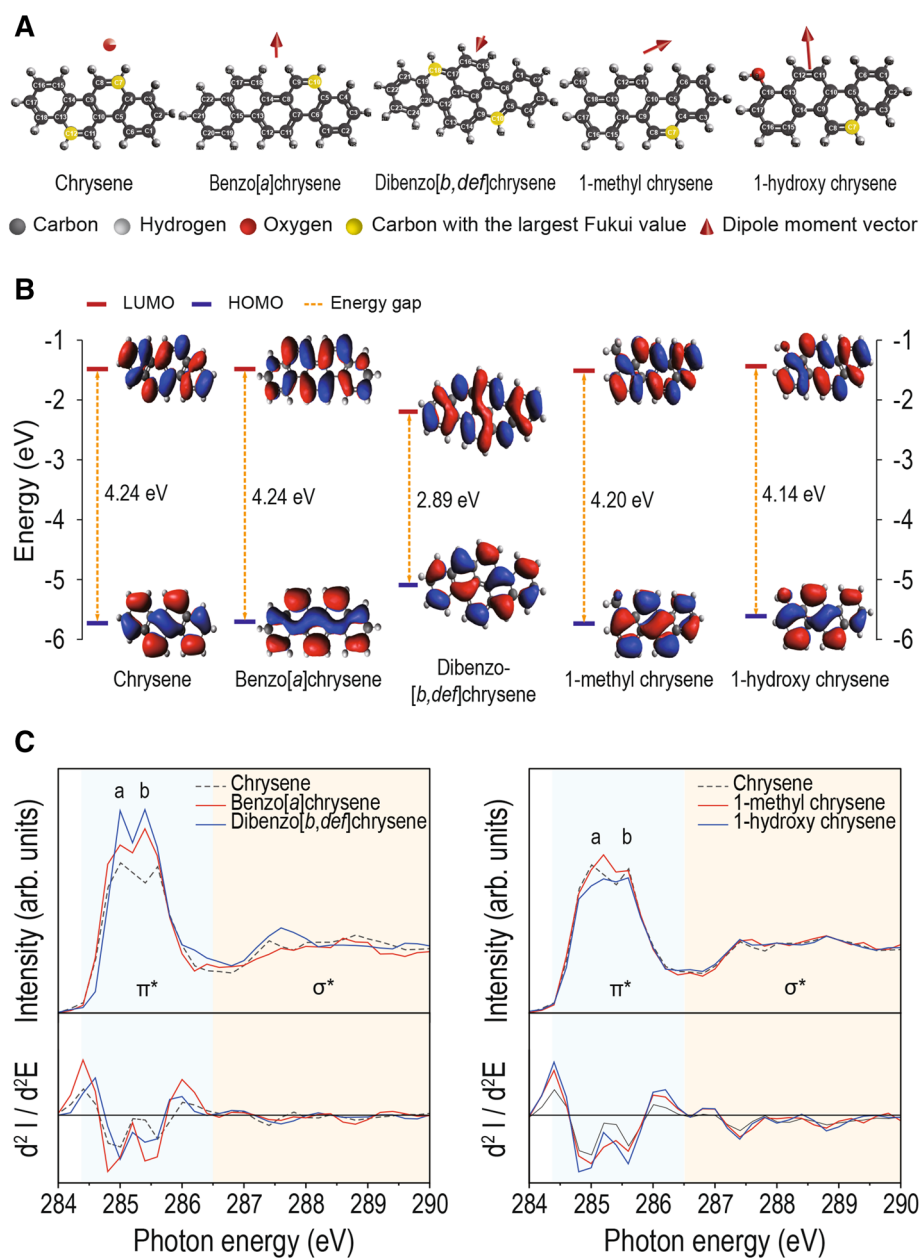


Figure 2. Molecular orbital structures of chrysene and its homologues (benzo, dibenzo, methyl and hydroxy chrysenes). **(A)** Change in dipole moment by addition of a benzene ring and functional group (methyl- and hydroxy-) of chrysene. Carbon atoms with the largest Fukui value in each chrysene homologue are highlighted in yellow and the red arrow denotes the direction and magnitude of dipole moment; **(B)** The HOMO and LUMO energies and corresponding orbital iso-surfaces for the chrysene homologues; **(C)** C 1s near-edge X-ray absorption fine structure (NEXAFS) spectra of chrysene homologues collected from pristine samples. Photon energy ranges exhibiting spectral features of carbon atoms bonded with hydrogen or other functional groups, C-H or C-X (peak **a**) and those bonded with neighboring carbon atoms, C-C (peak **b**).

of ligand-receptor binding and resulting ligand-specific bioactivity requires contributions involving not only physico-chemical properties but also interplay between and among them.

Comparison of predicted potencies: DR model versus current in silico models. DRF values and experimental toxic potency data (inverse of EC_{50}) were compared with the toxicity-relevant characteristics predicted by currently used in silico predictive models; QSAR (inverse of the median effective concentration, $1/EC_{50}$), a structure-based statistical regression method, and molecular docking models (free energy and binding distance), a structure-based analytical approach searching for the best ligand-receptor binding pose in terms of binding free energy^{32,33}. Target toxicants can be input as the unambiguous notation of the structure such as SMILES in VEGA-QSAR. The results of predictive EC_{50} were obtained from the Instituto di Ricerche Farma-

Compounds	HOMO (eV)	LUMO (eV)	HOMO–LUMO energy gap (eV)	Dipole moment magnitude (10^2 Debye)	Fukui value	Carbon atoms with the largest Fukui value	Location of carbon* to AhR	Directional reactivity factor (DRF)
					F^+			
Chrysene	-5.75	-1.51	4.24	0.03	0.068	7C, 12C	P	-0.02
Benzo[a]chrysene	-5.73	-1.50	4.24	3.30	0.062	10C	B	-2.98
Benzo[b]chrysene	-5.43	-1.94	3.49	1.89	0.056	8C	F	1.60
Benzo[c]chrysene	-5.77	-1.59	4.18	10.10	0.060	10C	F	9.75
Dibenzo[b,def]chrysene	-5.11	-2.22	2.89	0.03	0.062	10C, 18C	B	-0.02
Dibenzo[def,p]chrysene	-5.36	-1.97	3.39	3.43	0.059	10C	B	-3.57
1-methylchrysene	-5.69	-1.49	4.20	39.58	0.069	7C	F	6.79
2-methylchrysene	-5.68	-1.47	4.21	61.40	0.067	7C	F	27.97
3-methylchrysene	-5.67	-1.44	4.23	57.52	0.067	7C	F	13.66
1-hydroxychrysene	-5.56	-1.43	4.14	131.50	0.070	7C	B	-33.19
2-hydroxychrysene	-5.66	-1.49	4.17	145.70	0.069	7C	B	-51.44
3-hydroxychrysene	-5.62	-1.48	4.14	94.04	0.059	12C	B	-58.75

Table 1. Structural and electronic configuration parameters of the chrysene homologues and the potential toxicity predicted by directional reactivity factor (DRF). *The carbon with the largest Fukui value; forward (F), perpendicular (P), and backward (B) to the AhR.

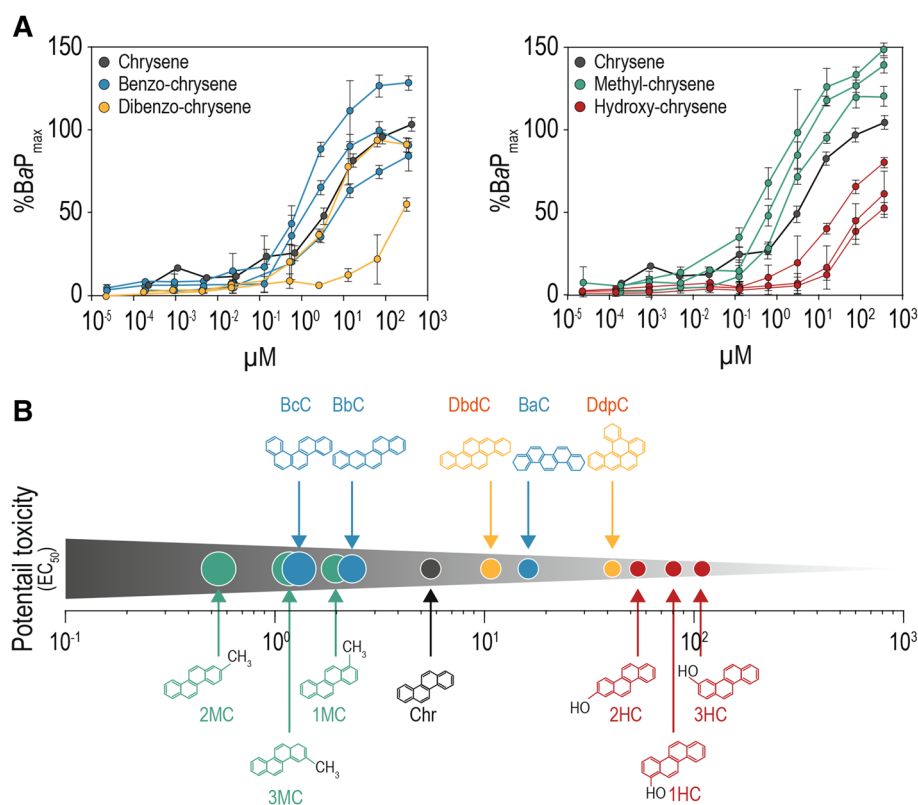


Figure 3. Experimental AhR-mediated potencies of chrysene homologues. (A) Dose-dependent activation of chrysene base compound and its homologues; benzo ($n=3$), dibenzo ($n=2$), methyl ($n=3$), and hydroxy chrysenes ($n=3$). Relative responses were normalized against the positive control (benzo[a]pyrene at the following concentrations; 100, 20, 4, 0.8, 0.16, 0.032, 0.0064, 0.00128, 0.000256, or 0.0000512 $\mu\text{g mL}^{-1}$; set as 1.0 of potential toxicity). Dose expressed as nmol chrysene homologues/mL media (μM) present in the test well. Error bars represent the standard error; (B) The order of potential toxicity of chrysene homologue based on EC_{50} calculated for H4IIE-*luc* transactivation assay. Examined chrysene homologues were ranked in descending order of potential toxicity. The methyl chrysene group exhibits the highest potency followed by benzo chrysene one while dibenzo chrysene and hydroxy-chrysene groups show similar or lower potencies than other chrysene homologue groups.

Chemical	In vitro bioassay	QSAR ^a	Docking model ^b			Directional reactivity factor (DRF)	
	H4IIE- <i>luc</i> EC ₅₀	<i>Daphnia magna</i> EC ₅₀	Free energy (kcal mol ⁻¹)	Binding distance to LBD (Å)			
	(μM)	(μM)		Total	H285 ^c		F318 ^d
2-methylchrysene	6.6 × 10 ⁻¹	7.4 × 10 ⁻¹	-7.1	6.3	3.0	3.3	28.0
3-methylchrysene	11 × 10 ⁻¹	7.4 × 10 ⁻¹	-10.5	6.3	3.0	3.4	13.7
Benzo[c]chrysene	13 × 10 ⁻¹	5.4 × 10 ⁻¹	-8.0	7.4	3.6	3.8	9.8
1-methylchrysene	19 × 10 ⁻¹	7.4 × 10 ⁻¹	-9.8	6.6	3.1	3.4	6.8
Benzo[b]chrysene	28 × 10 ⁻¹	5.4 × 10 ⁻¹	-6.6	22.8	9.1	13.8	1.6
Chrysene	72 × 10 ⁻¹	9.1 × 10 ⁻¹	-9.7	6.6	3.2	3.4	-0.02
Dibenzo[b,def]chrysene	140 × 10 ⁻¹	4.2 × 10 ⁻¹	-6.1	23.2	9.2	14.0	-0.02
Benzo[a]chrysene	200 × 10 ⁻¹	5.4 × 10 ⁻¹	-5.6	25.1	10.6	14.5	-3.0
Dibenzo[def,p]chrysene	760 × 10 ⁻¹	4.2 × 10 ⁻¹	-7.8	33.0	18.4	14.6	-3.6
2-hydroxychrysene	590 × 10 ⁻¹	32 × 10 ⁻¹	-8.0	31.9	17.7	14.2	-51.4
1-hydroxychrysene	870 × 10 ⁻¹	32 × 10 ⁻¹	-8.9	32.3	17.7	14.6	-33.2
3-hydroxychrysene	1400 × 10 ⁻¹	32 × 10 ⁻¹	-8.1	6.7	2.9	3.8	-58.8

Table 2. Comparison of the directional reactivity (DR) model with other predictive models and experimental potential toxicity data. ^aPredicted data from VEGA-QSAR. ^bGalaxydock. ^cHistidine 285. ^dPhenylalanine 318 in the AhR homology.

cologiche Mario Negri (IRFMN) toxicity model (*Daphnia magna* acute) based on the training sets including aromatic hydrocarbons³⁴. The same AhR homology used in the DR model was employed for calculations with docking models. Overall, the results suggest that both QSAR and molecular docking models do not accurately estimate the experimental potencies determined by use of in vitro bioassays, say anti-empirical data.

The QSAR model predicted greatest toxic potency of dibenzo- and benzo-chrysene groups followed by chrysene base compound and methylated chrysenes. Different experimental toxicities within each homologue group are also not reproduced, rather the similar degree of toxicity among congeners was observed (benzo-chrysene: 5.4 × 10⁻¹ μM; dibenzo-chrysene: 4.2 × 10⁻¹ μM; methyl-chrysene: 7.4 × 10⁻¹ μM; hydroxy-chrysene: 32 × 10⁻¹ μM) (Fig. 4A and Table 2). This results in statistically insignificant correlations between experimental 1/EC₅₀ and QSAR-estimated 1/EC₅₀ (R² = 0.20, *p* > 0.05), which implies considerable uncertainty in prediction accuracy of toxicity (Fig. 4B). Unlike the QSAR approach, the docking analysis predicts different toxic potencies of congeners in each homologue group and high binding affinity behaviors (i.e., low free energy and short binding distance) of the group of methylated chrysene, which was consistent with empirical toxicity data (Fig. 4B and Figure S6).

Chrysene homologues composed of benzene rings (benzo- and dibenzo-chrysene groups) exhibited lesser free energies than those of methyl- and hydroxy-chrysene groups. Another notable feature is that hydroxy-chrysenes show comparable free energy to methyl chrysenes while experimental toxic potency due to addition of the hydroxyl group is least among chrysene homologues tested. Considering the fact that hydroxylated benzene compounds are usually known as metabolites by degradation or metabolism in abiotic and biotic responses, the docking model seems to overestimate toxicities of these model compounds (Table S3 presents a result of potential toxicity from other in silico models to predict toxic potencies)³⁵. The statistical correlation between calculated optimal binding affinity and experimental 1/EC₅₀ values is found to be insignificant (R² = 0.09, *p* > 0.05) (Fig. 4B). Meanwhile, although a positive correlation between 1/EC₅₀ and total binding distance between the ligand and either of two binding domains (H285 and F318) was observed (R² = 0.61, *p* < 0.01), the short binding distances calculated for chrysene and 3HC compounds are inconsistent with experimental toxicity results.

Discrepancies of experimental bioactivity data from toxicity-relevant characteristics predicted by current in silico models are not unexpected. In fact, QSAR approaches are based on algorithms composed of a linear regression relationship between logarithms of aquatic toxicity of known chemicals and their structural molecular descriptors (training set). Effective toxic concentrations of individual chemicals were estimated by use of regression equations after training and the measure or calculated values of corresponding chemical from the descriptors³⁶. As a consequence, the congeners in each homologue group with the same molecular weight shows the same toxic potency according to the QSAR model.

Discussion

Toxicity testing has become central to hazard assessments of (in)organic substances associated with causal exposure in environments. Thus, accurate evaluation of adverse effects of chemicals on bio-organisms have been a profound problem in environmental and pharmaceutical toxicology³⁷. Development of a model to reliably predict bioactivity complementing traditional bioassays is directly related with a fundamental question about how a chemical trigger the ligand-receptor binding that results in the following expression of xenobiotic metabolizing genes. Current computational approaches have focused on statistical correlations between structure-based physico-chemical properties and the observed bioactivity endpoints using multiple linear regression (i.e., QSAR),

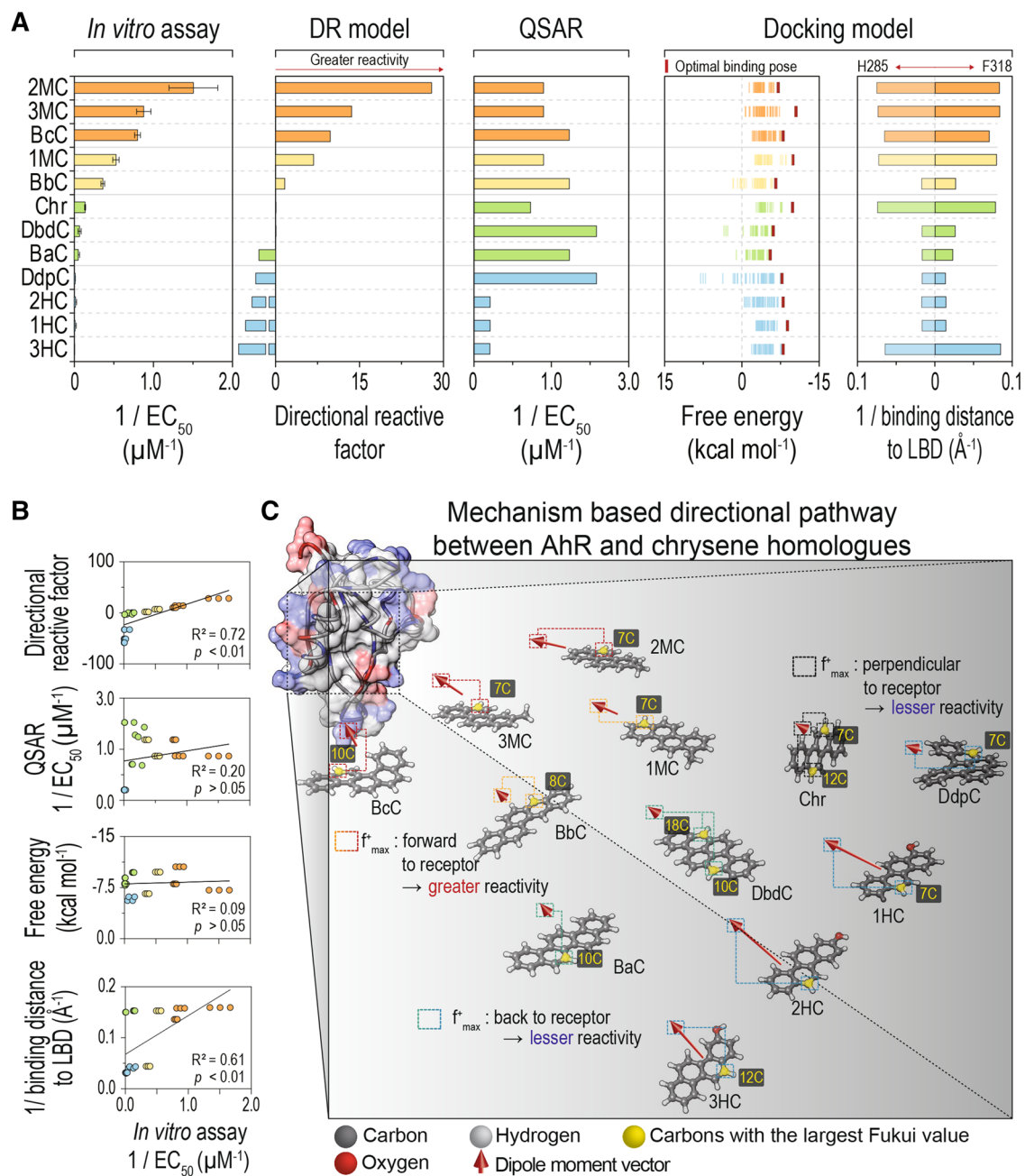


Figure 4. Prediction efficiency of DR model and other in silico assays. **(A)** Experimental potency results of in vitro bioassay (H4IIE-*luc*) and predicted toxicity-relevant characteristics by in silico models (QSAR, molecular docking model, and the DR model introduced here). The QSAR and the docking model calculations were made using VEGA-QSAR and GalaxyDock, respectively. Each toxicant in the graph was sorted in descending order of $1/EC_{50}$ value. **(B)** Predicted versus observed values represented from each method and the regression coefficient, R^2 with significance (p). **(C)** Visualization of first principles prediction model for AhR-mediated potency of chrysene homologues before their binding to AhR.

and characteristics of the ligand-receptor binding after minimization of free energy of binding (i.e., molecular docking)^{38–40}. However, here, we have shown that structural similarity does not always guarantee that bioactivities will also be similar, but small structural differences, like different functionalization sites between 1MC and 2MC, can result in dramatic differences in toxic potency. Furthermore, since long-range dipole-charge interactions predetermine orientation and relative distribution of ligand's reactive sites before binding to the receptor, the best ligand-receptor binding pose with the minimum free energy must not be always plausible.

The DR model suggests that individual physico-chemical properties of specific ligands do not contribute independently to binding, but interplay between them is also important. This kinetic approach begins with consideration of the “molecular dipole moment” that is sensitive to functionalization, as the main property of ligand interacting with a receptor (mediated by trapped cytosolic ions) after a ligand is introduced into the cell cytosol.

However, the favorable binding environment is achieved by combined influence from the molecular orientation and relative location of reactive sites driven by “alignment” of the dipole moment (Fig. 4C). In this study, the DRF devised from the molecular level kinetic process could estimate the macroscopic toxic potency of chrysene homologues and therefore serve as a bioactivity indicator of chemicals. Once the amino acid sequence and constituent, charged acids are identified, the DR model can be applied to investigate species-specific bioactivity.

Unlike other current models, this model does not predict toxicity through statistical methods, but predicts the possibility of toxic effects, assuming the pre-conjugation state in the case of toxicants combined with receptors. Based on this model, it was confirmed that the reactivity (Fukui value) of molecules in each ligand group was different depending on the position despite their structural similarity, and the contribution in dipole moment of the ligand was confirmed according to the position of the reactive molecule. Thus, we found that the shape and orientation configuration of toxicants before the ligand-receptor interaction were generally consistent with EC₅₀ values. This model was based on a good understanding on the toxic mechanisms and could apply to compounds with different action mechanisms explaining different reactivity of similarly structured chemicals to AhR.

Computational studies have been increasing substantially in the last few years. These computational techniques with toxicity prediction model highly rely on QSAR analysis. While the DR model is an *in silico*-based approach like QSAR, but it employs more mechanism-driven consideration of toxic reaction rather than statistical inference and extrapolation. The DRF could bring up a topic of understanding the more scientifically robust basis for risk assessment by providing detailed mechanistic pathways and serve as a complementary to existing QSAR methodology. Using various efficient tools to calculate electronic excitation such as time dependent density functional theory (TDDFT), the DR model have the potential to extend to predict chemical interaction with biomolecules (e.g., fatty acids, DNA, RNA, and micro RNA)⁴¹. Therefore, further improvement of the DR model will keep pace with mutual advances in molecular biology and computational structural biology and provide a useful tool for discovery of drugs or predictive toxicology.

Materials and methods

Selection of model chemicals. As a first step toward identification of the major physical factors affecting toxic potencies, we selected the model chemicals considering four criteria: First, the selection began with a pool of high priority 16 PAHs regulated by the US EPA which have played an exceptionally large role in environmental sciences. Second, candidates were narrowed to 7 PAHs exhibiting acute toxic potency since the DR model does not consider absorption, distribution, metabolism, and excretion, and thus the model compounds should show perceivable variation in the toxic potency for a short period of exposure⁴². Third, for prediction of the binding activity with AhR, the most potent AhR agonist in the environment would be desirable. Finally, various homologues of the selected AhR agonist were added in order to understand how the structural modification of ligand induces different binding activity with AhR and resulting toxic potency. Following these steps, chrysene and its homologues were selected as the model compounds for application of our bio-physical communication model and comparison with experimental results of toxic potency testing. Chrysene homologues cover a broad set of compounds ranging from four- to six-membered benzene to methylated- and hydroxylated benzene; chrysene (Chr), BaC, BbC, BcC, DbdC, DdpC, 1MC, 2MC, 3-methylchrysene (3MC), 1-hydroxychrysene (1HC), 2-hydroxychrysene (2HC), and 3HC. Their full names, providers, and purities are present in the Supplementary information (Table S1).

Density functional theory calculations. Physical properties of chrysene and its homologues to be implemented in the DR model were obtained from DFT calculations, which were performed using the ORCA (version 4.1.1) program package⁴³. All stages of DFT computations were made by use of B3LYP exchange functional with the def2-TZVPP basis set. Meanwhile, very tight, self-consistent field (SCF) convergence (the energy change is 10⁻⁹ a.u.) and fine integration grid of 5.0 were employed. The first stage (spin-restricted) geometric optimizations were performed for all structures of the homologues of chrysene. Next, $\pm e$, charge was added to each compound as cationic or anionic case for the second stage (spin-unrestricted) optimizations. Based on the corresponding optimized structures, the molecular orbitals, Hirshfeld charge populations, density of orbital states, molecular dipole moments, and vibrational amplitudes for neutral, cationic and anionic states of each chrysene homologue were computed. The atom-condensed Fukui functions for all carbon atoms bonded with hydrogen or other functional groups in each homologue were determined by Hirshfeld population analysis²⁸.

C 1s NEXAFS spectroscopy. Since physical properties were determined from the DFT-calculated molecular orbitals of each model compound, the calculated molecular orbital structures were verified through comparison with experimental unoccupied molecular orbital structures measured by NEXAFS spectroscopy. Measurements of C 1s NEXAFS spectra were made at the 10D HR XAS KIST beamline of the Pohang Light Source in Korea. All NEXAFS spectra were measured in total electron yield (TEY) mode. After data acquisition, the spectra were energy calibrated with the Π^* (C=C) transition at 285.6 eV of highly oriented pyrolytic graphite (HOPG). Samples were tested for radiation damage considering the sensitivity of carbon-based materials to X-ray radiation. Spectra were intensity-normalized to the incoming photon flux, as recorded by an Au mesh.

H4IIE-*luc* transactivation bioassay for evaluating AhR-mediated potencies and calculation of EC₅₀ of homologues of chrysene. H4IIE cell line derived from rat hepatoma has been reported to continuously express cytochrome P450 protein/mRNA. For this study, H4IIE cells transfected with the luciferase (*Luc*) reporter gene plasmid (pGudLuc1.1) were used, originally obtained from Jac Aarts, University of Wageningen, The Netherlands and further developed by John Giesy, University of Saskatchewan^{44,45}. Upon ligand binding, AhR in H4IIE-*luc* is activated and translocated to the nucleus, where heat shock proteins are dissociated from

the complex and they form a dimer with the AhR nuclear translocator (ARNT) protein. This complex binds to specific DNA sequences with high affinity, the dioxin-responsive element (DRE). The binding to the DRE results in DNA bending transcriptional activation of adjacent responsive genes. The detailed mechanism is provided in Figure S7.

Acute toxic potencies of chrysene homologues were determined based on an AhR-mediated activity using H4IIE-*luc* cells, by the method previously reported⁴⁶. Since PAHs could be degraded by metabolic activity during the H4IIE-*luc* transactivation bioassay with longer exposure time, the 4 h exposure time was chosen for calculation of AhR-mediated toxic potency. After the exposure, the activity results were expressed as relative luminescence units that were quantified using a Victor X3 multi label plate reader (PerkinElmer, Waltham, MA). Benzo[*a*]pyrene was used as a positive control ligand. We converted responses of the H4IIE-*luc* bioassay to the percentages of the maximum response (%BaP_{max}) observed for a 50 nM BaP (= 100%BaP_{max}). All bioassays were repeated four times in triplicate (Table S4 for detail information of in vitro bioassay conditions). The half effective concentrations for the AhR-mediated effects of individual compounds were determined by use of H4IIE-*luc* bioassays. Compounds with 10 concentrations using fivefold serial dilutions (viz., 100, 20, 4, 0.8, 0.16, 0.032, 0.0064, 0.00128, 0.000256, and 0.0000512 µg mL⁻¹) were tested. Estimations of EC₅₀ values of target compounds from the dose–response relationships were basically assumed equal efficacy and parallelism between target compounds and reference compound (benzo[*a*]pyrene).

In silico analysis: quantitative structure–activity relationship (QSAR) and molecular docking model. Prediction capability of current in silico models for chrysene homologues were tested with QSAR model and molecular docking analysis. VEGA-QSAR program and the GalaxyDock program were used for QSAR and molecular docking analysis, respectively. VEGA-QSAR has been performed and showed a large set of toxicological data and endpoints convincing to derive predictive toxicity according to the previous studies⁴⁷. In our case, EC₅₀ values were obtained using SMILES notation of VEGA-QSAR and obtained the toxicity data corresponding to *D. magna*.

For molecular docking analysis, the GalaxyDock was used to simulate the optimized configuration of AhR in the presence of model chemicals and determine the binding affinity for receptor–ligand docking at each ligand. The receptor–ligand docking was then simulated using the GalaxyDock in GalaxyWEB^{34,48}. For each model chemical, 50 candidate configurations of the ligand–receptor docking were calculated and the configuration with the lowest formation energy was chosen as the optimal binding structure. Two aromatic residues of histidine 285 (H285) and phenylalanine 318 (F318) in the homology modeling were recognized as the most important for ligand binding⁴⁹. The distances between the nearest aromatic ring or residues of chrysene homologues and each of the ligand binding domains were calculated from the predicted binding pose generated by GalaxyDock. For calculations, the three-dimensional (3D) structure of the ligand binding domain (LBD) of the AhR was built from the sequence of amino acids in the rat AhR (GI: 7304873 in the NCBI sequence database) by the validation using the SWISS-MODEL, Rampage web servers, and ProSA-web^{26,50}. A crystal structure of 4F3L (sequence identity = 27.8%, residues 252–357) was selected as a template. The structure model, built using Ramachandran plot from Rampage web server, was validated since 98% of the total residues were in the favored region with 2.4% of the residues in the allowed region. Furthermore, results from ProSA analyses showed the z-score for the 3D model to be within the range of score typically found for native proteins of the similar size.

Data availability

All data are available in the main text or the Supplementary information.

Received: 19 March 2020; Accepted: 30 July 2020

Published online: 14 August 2020

References

- Rothhammer, V. & Quintana, F. J. The aryl hydrocarbon receptor: an environmental sensor integrating immune responses in health and disease. *Nat. Rev. Immunol.* **19**, 184–197 (2019).
- Bernetti, M., Cavalli, A. & Mollica, L. Protein–ligand (un) binding kinetics as a new paradigm for drug discovery at the crossroad between experiments and modelling. *Med. Chem. Comm.* **8**, 534–550 (2017).
- the case of the adenosine receptors. Guo, D., Heitman, L. H. & IJzerman, A. P. Kinetic aspects of the interaction between ligand and G protein-coupled receptor. *Chem. Rev.* **117**, 38–66 (2016).
- Ankley, G. T. *et al.* Adverse outcome pathways: a conceptual framework to support ecotoxicology research and risk assessment. *Environ. Sci. Pollut. Res. Int.* **29**, 730–741 (2010).
- Zhao, Y., Zhang, K., Giesy, J. P. & Hu, J. Families of nuclear receptors in vertebrate models: characteristic and comparative toxicological perspective. *Sci. Rep.* **5**, 8554 (2015).
- Doering, J. A. *et al.* Differences in activation of aryl hydrocarbon receptors of white sturgeon relative to lake sturgeon are predicted by identities of key amino acids in the ligand binding domain. *Environ. Sci. Technol.* **49**, 4681–4689 (2015).
- Legler, J. *et al.* Comparison of in vivo and in vitro reporter gene assays for short-term screening of estrogenic activity. *Environ. Sci. Technol.* **36**, 4410–4415 (2002).
- Ferdowsian, H. R. & Beck, N. Ethical and scientific considerations regarding animal testing and research. *PLoS ONE* **6**, e24059 (2011).
- Cumming, J. G. *et al.* Chemical predictive modelling to improve compound quality. *Nat. Rev. Drug Discov.* **12**, 948–962 (2013).
- Chen, Q. *et al.* Molecular initiating events of Bisphenols on androgen receptor-mediated pathways provide guidelines for in silico screening and design of substitute compounds. *Environ. Sci. Technol. Lett.* **6**, 205–210 (2019).
- Gramatica, P. & Sangion, A. A historical excursus on the statistical validation parameters for QSAR models: a clarification concerning metrics and terminology. *J. Chem. Inf. Model.* **56**, 1127–1131 (2016).
- Kirchmair, J. *et al.* Predicting drug metabolism: experiment and/or computation?. *Nat. Rev. Drug Discov.* **14**, 387–404 (2015).
- Cash, G. Prediction of chemical toxicity to aquatic organisms: ECOSAR vs. Microtox® Assay. *Environ. Toxic. Water.* **13**, 211–216 (1998).

14. Ribeiro, F. A. L. & Ferreira, M. M. C. QSAR model of the phototoxicity of polycyclic aromatic hydrocarbons. *J. Mol. Struct. Theoret. Chem.* **719**, 191–200 (2005).
15. Zvinavashe, E. *et al.* QSAR models for predicting in vivo aquatic toxicity of chlorinated alkanes to fish. *Chem. Res. Toxicol.* **21**, 739–745 (2008).
16. Yang, W. *et al.* Molecular docking and comparative molecular similarity indices analysis of estrogenicity of polybrominated diphenyl ethers and their analogues. *Ecotox. Environ. Toxicol. Chem.* **29**, 660–668 (2010).
17. Wang, X. *et al.* Docking and CoMSIA studies on steroids and non-steroidal chemicals as androgen receptor ligands. *Ecotox. Environ. Safe.* **89**, 143–149 (2013).
18. Forrest, J. *et al.* A comprehensive model for chemical bioavailability and toxicity of organic chemicals based on first principles. *Front. Mar. Sci.* **1**, 31 (2014).
19. Helm, C. A., Knoll, W. & Israelachvili, J. N. Measurement of ligand-receptor interactions. *Proc. Natl. Acad. Sci. USA* **88**, 8169–8173 (1991).
20. Petukh, M., Stefl, S. & Alexov, E. The role of protonation states in ligand-receptor recognition and binding. *Curr. Pharm. Des.* **19**, 4182–4190 (2013).
21. Chattaraj, P. K. *Chemical Reactivity Theory: A Density Functional View* (CRC Press, Cambridge, 2009).
22. Hirano, M. *et al.* In silico analysis of the interaction of avian aryl hydrocarbon receptors and dioxins to decipher isoform-, ligand-, and species-specific activations. *Environ. Sci. Technol.* **49**, 3795–3804 (2015).
23. Parr, R. G. & Yang, W. Density-functional theory of the electronic structure of molecules. *Annu. Rev. Phys. Chem.* **46**, 701–728 (1995).
24. Tagliabue, S. G., Faber, S. C., Motta, S., Denison, M. S. & Bonati, L. Modeling the binding of diverse ligands within the Ah receptor ligand binding domain. *Sci. Rep.* **9**, 1–14 (2019).
25. Zhang, S. *et al.* Uptake and translocation of polycyclic aromatic hydrocarbons (PAHs) and heavy metals by maize from soil irrigated with wastewater. *Sci. Rep.* **7**, 12165 (2017).
26. Waterhouse, A. *et al.* SWISS-MODEL: homology modelling of protein structures and complexes. *Nucl. Acids. Res.* **46**, W296–W303 (2018).
27. Lear, S. & Cobb, S. L. Pep-Calc. com: a set of web utilities for the calculation of peptide and peptoid properties and automatic mass spectral peak assignment. *J. Comput. Aided. Mol. Des.* **30**, 271–277 (2016).
28. Bultinck, P., Van Alsenoy, C., Ayers, P. W. & Carbó-Dorca, R. Critical analysis and extension of the Hirshfeld atoms in molecules. *J. Chem. Phys.* **126**, 144111 (2007).
29. Fronzoni, G. *et al.* Vibrationally resolved high-resolution NEXAFS and XPS spectra of phenanthrene and coronene. *J. Chem. Phys.* **141**, 044313 (2014).
30. Faundez, V. & Hartzell, H. C. Intracellular chloride channels: determinants of function in the endosomal pathway. *Sci. Signal.* **233**, re8 (2004).
31. Friedman, R. Ions and the protein surface revisited: extensive molecular dynamics simulations and analysis of protein structures in alkali-chloride solutions. *J. Phys. Chem. B.* **115**, 9213–9223 (2011).
32. Cappelli, C. I., Benfenati, E. & Cester, J. Evaluation of QSAR models for predicting the partition coefficient (log P) of chemicals under the REACH regulation. *Environ. Res.* **143**, 26–32 (2015).
33. Du, X. *et al.* Insights into protein–ligand interactions: mechanisms, models, and methods. *Int. J. Mol.* **17**, 144 (2016).
34. Shin, W.-H. & Seok, C. GalaxyDock: protein–ligand docking with flexible protein side-chains. *J. Chem. Inf. Model.* **52**, 3225–3232 (2012).
35. Geier, M. C. *et al.* Comparative developmental toxicity of a comprehensive suite of polycyclic aromatic hydrocarbons. *Arch. Toxicol.* **92**, 571–586 (2018).
36. Gissi, A. *et al.* Integration of QSAR models for bioconcentration suitable for REACH. *Sci. Total. Environ.* **456**, 325–332 (2013).
37. Krewski, D. *et al.* Toxicity testing in the 21st century: a vision and a strategy. *J. Toxicol. Environ. Health B.* **13**, 51–138 (2010).
38. Floris, M., Manganaro, A., Nicolotti, O., Medda, R. & Mangiatordi, G. F. A generalizable definition of chemical similarity for read-across. *J. Cheminform.* **6**, 39 (2014).
39. Deng, Y. & Roux, B. Calculation of standard binding free energies: Aromatic molecules in the T4 lysozyme L99A mutant. *J. Chem. Theory Comput.* **2**, 1255–1273 (2006).
40. Mobley, D. L. *et al.* Predicting absolute ligand binding free energies to a simple model site. *J. Mol. Biol.* **371**, 1118–1134 (2007).
41. Varsano, D. *et al.* A TDDFT study of the excited states of DNA bases and their assemblies. *J. Phys. Chem. B.* **110**(14), 7129–7138 (2006).
42. Kim, J. *et al.* Newly identified AhR-active compounds in the sediments of an industrial area using effect-directed analysis. *Environ. Sci. Technol.* **53**, 10043–10052 (2019).
43. Neese, F. The ORCA program system. *WIREs Comput. Mol. Sci.* **2**(1), 73–78 (2012).
44. Giesy, J. P. *et al.* Cell bioassays for detection of aryl hydrocarbon (AhR) and estrogen receptor (ER) mediated activity in environmental samples. *Mar. Pollut. Bull.* **45**(1–12), 3–16 (2002).
45. Denison, M. S. *et al.* Production of novel recombinant cell line bioassay systems for detection of 2,3,7,8-tetrachlorodibenzo-p-dioxin-like chemicals. *Organohalogen Compd.* **13**, 365–368 (1993).
46. Louiz, I. *et al.* Monitoring of dioxin-like, estrogenic and anti-androgenic activities in sediments of the Bizerta lagoon (Tunisia) by means of in vitro cell-based bioassays: contribution of low concentrations of polynuclear aromatic hydrocarbons (PAHs). *Sci. Total. Environ.* **402**, 318–329 (2008).
47. Benfenati, E., A. & Manganaro, G. VEGA-QSAR: AI inside a platform for predictive toxicology. *Proceedings of the Workshop, PAI*, pp. 21–28 (2013).
48. Shin, W.-H., Lee, G. R., Heo, L., Lee, H. & Seok, C. Prediction of protein structure and interaction by GALAXY protein modeling programs. *Bio. Des.* **2**, 1–11 (2014).
49. Lee, S. *et al.* Measured and predicted affinities of binding and relative potencies to activate the AhR of PAHs and their alkylated analogues. *Chemosphere* **139**, 23–29 (2015).
50. Wiederstein, M. & Sippl, M. J. ProSA-web: interactive web service for the recognition of errors in three-dimensional structures of proteins. *Nucl. Acids Res.* **35**, W407–W410 (2007).

Acknowledgements

This work was supported by the National Research Foundation of Korea (NRF-2017R1E1A1A01075067) and the Natural Sciences and Engineering Research Council of Canada (NSERC) grants.

Author contributions

T. K. performed the data analysis and wrote the original draft under the supervision of J. S. K. and G. S. C., J. S. K., and J. P. G. co-wrote and edited intermediate drafts of the manuscript. J. Z. and R. B. calculated physico-chemical properties for model compounds. J. L., C. L., S. H., B.-O. K., K. H. C., and S. H. helped with data collection and

interpretation. All authors contributed to formulation of the study design, sample preparation, data interpretation, and develop the DR model used to make predictions.

Competing interests

The authors declare no competing interests.

Additional information

Supplementary information is available for this paper at <https://doi.org/10.1038/s41598-020-70704-9>.

Correspondence and requests for materials should be addressed to G.S.C. or J.S.K.

Reprints and permissions information is available at www.nature.com/reprints.

Publisher's note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons licence, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit <http://creativecommons.org/licenses/by/4.0/>.

© The Author(s) 2020

<Scientific Reports>

Supplementary information for

Influence of ligand's directional configuration, chrysenes as model compounds, on the binding activity with aryl hydrocarbon receptor

Taewoo Kim, Juyuan Zhen, Junghyun Lee, Robert Bauer, Changkeun Lee, Bong-Oh Kwon, Keun Hwa Chae, Seongjin Hong, John P. Giesy, Gap Soo Chang*, Jong Seong Khim*

This PDF file includes:

Number of pages: 13

Number of supplementary figures: 7, Figures S1 to S7

Number of supplementary tables: 4, Tables S1 to S4

References

***Corresponding authors:**

E-mail addresses: jskocean@snu.ac.kr, (J.S. Khim); gapsoo.chang@usask.ca (G.S. Chang)

Supplementary Figures

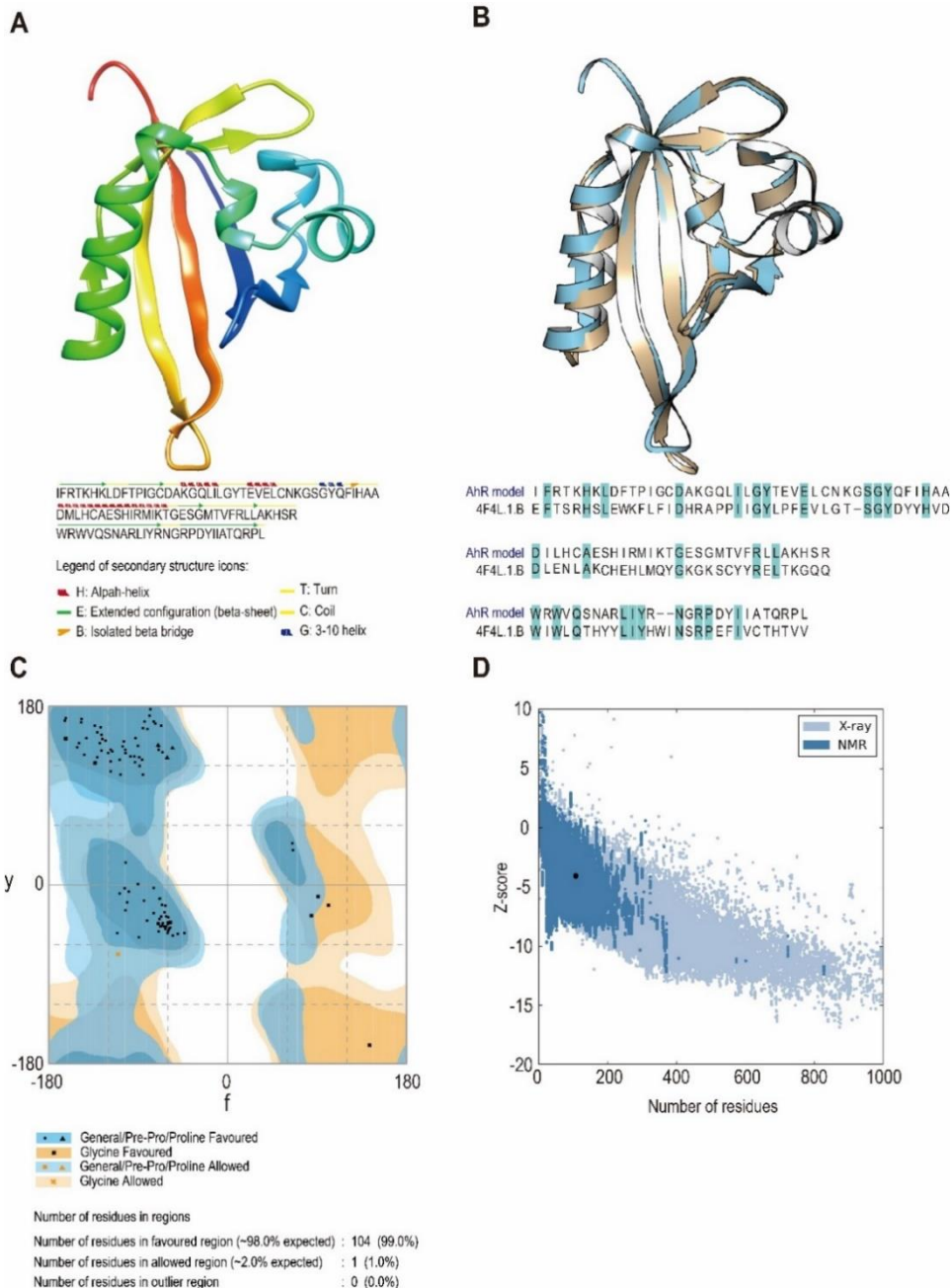


Figure S1. Homology model and quality metrics of AhR LBD. (A) Modelled structure of AhR LBD with helices are shown in red, sheets are shown in yellow and loops are shown in green¹; (B) Superimposition of the template (light blue) with the AhR model (light brown) with an RMSD less than 1.0 Å and a structure overlap of 98.13; (C) Ramachandran plot showing energetically allowed regions for backbone dihedral angles ψ against ϕ of amino acid residues in the AhR structure; (D) The ProSA analysis of the generated AhR structure in this study.² The figure represents the Prosa-web plot of template 4F3L chain A with a z-score value of -4.07 .

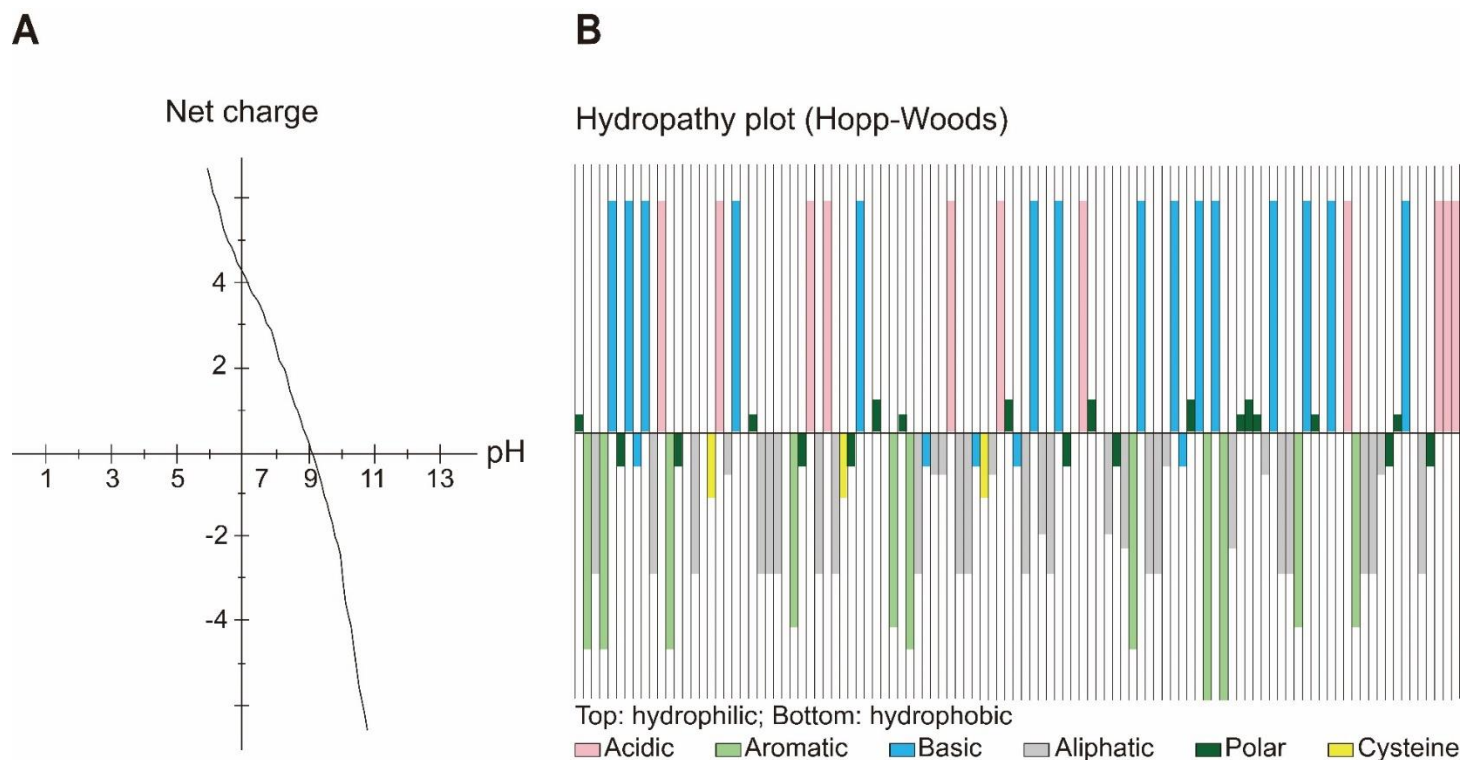


Figure S2. Calculated net charge and hydropathy plot in aryl hydrocarbon homology. **(A)** Expected net charge in AhR homology. AhR homology was expected to have a net charge of 4.1 in pH 7.4³; **(B)** Hydropathy plot of AhR homology indicating its physico-chemical properties. The plot showed the degree of hydrophobicity or hydrophilicity of amino acids of AhR. The hydrophobic amino acids accounted for 42.05 of the total sequence in AhR homology, followed by 28.97 % for the hydrophilic amino acids. Basic and acidic were 18.69 % and 10.28 % respectively.

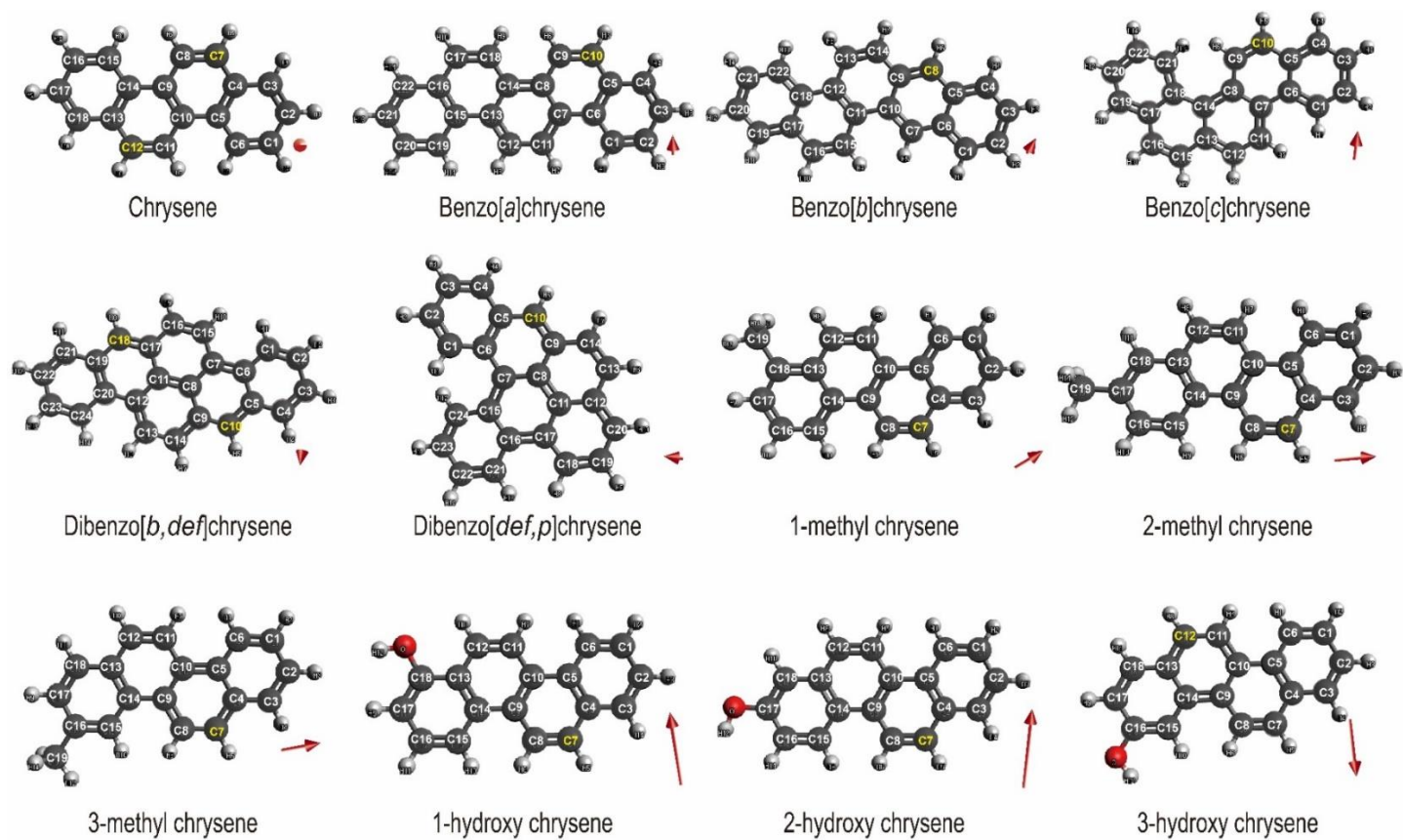


Figure S3. The information of physico-chemical properties in each chrysene homologue. Carbons in structure of chrysene homologues are shown in dark gray, hydrogens are shown in white and oxygen are shown in red color. A carbon with the highest Fukui value of chrysene homologues is shown in yellow, and dipole moments are shown in red arrow. The size of the arrow indicates the quantity of dipole moment.

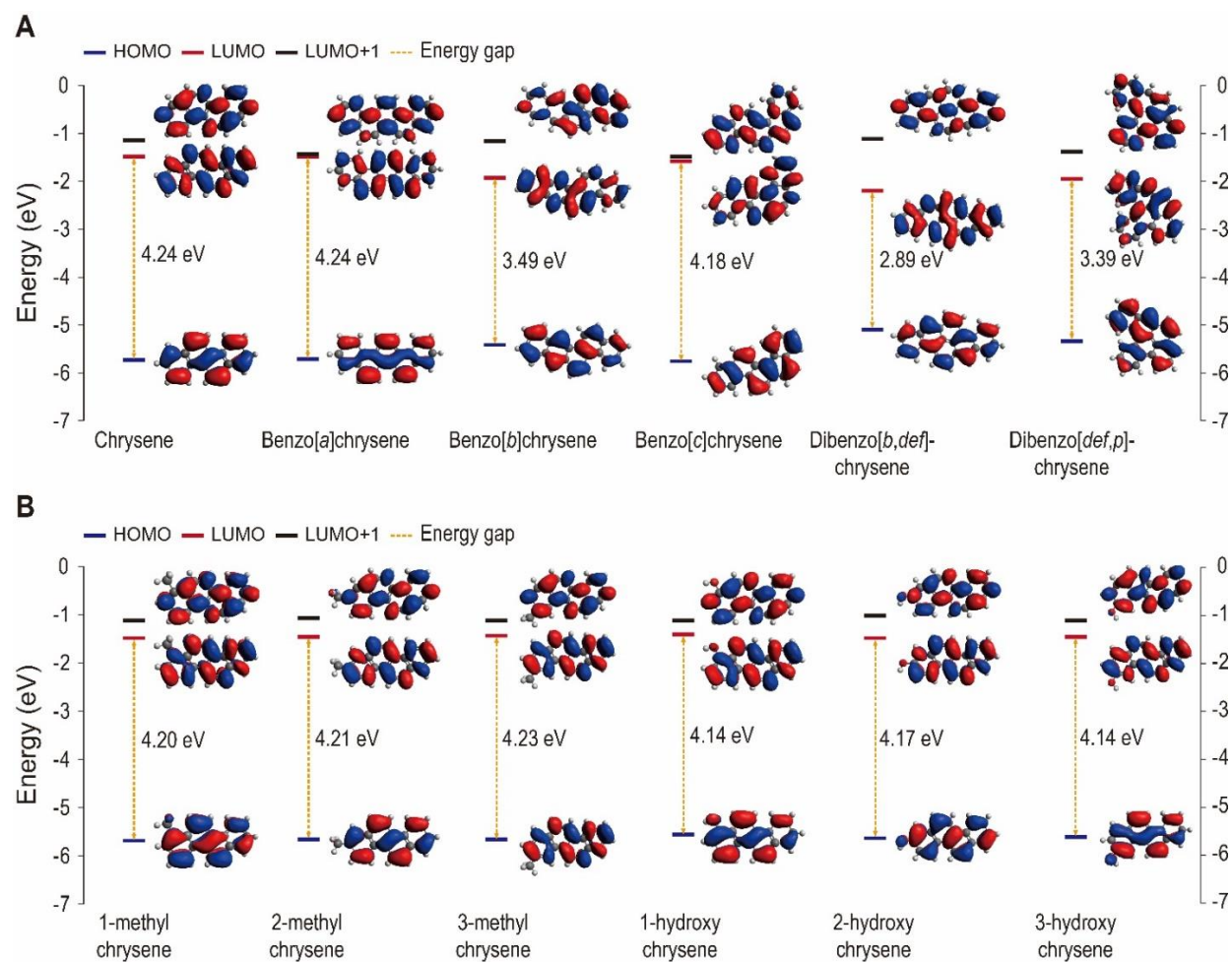


Figure S4. Visualization of the HOMO and LUMO orbital in chrysene homologues. (A and B) Presentation of the energy levels, HOMO-LUMO gap and orbital composition distribution of the HOMO and LUMO for chrysene and benzo-chrysene. HOMO-LUMO gap energy ranged from 2.89 eV to 4.24 eV. The HOMO-LUMO gap is slightly difference in each congener.

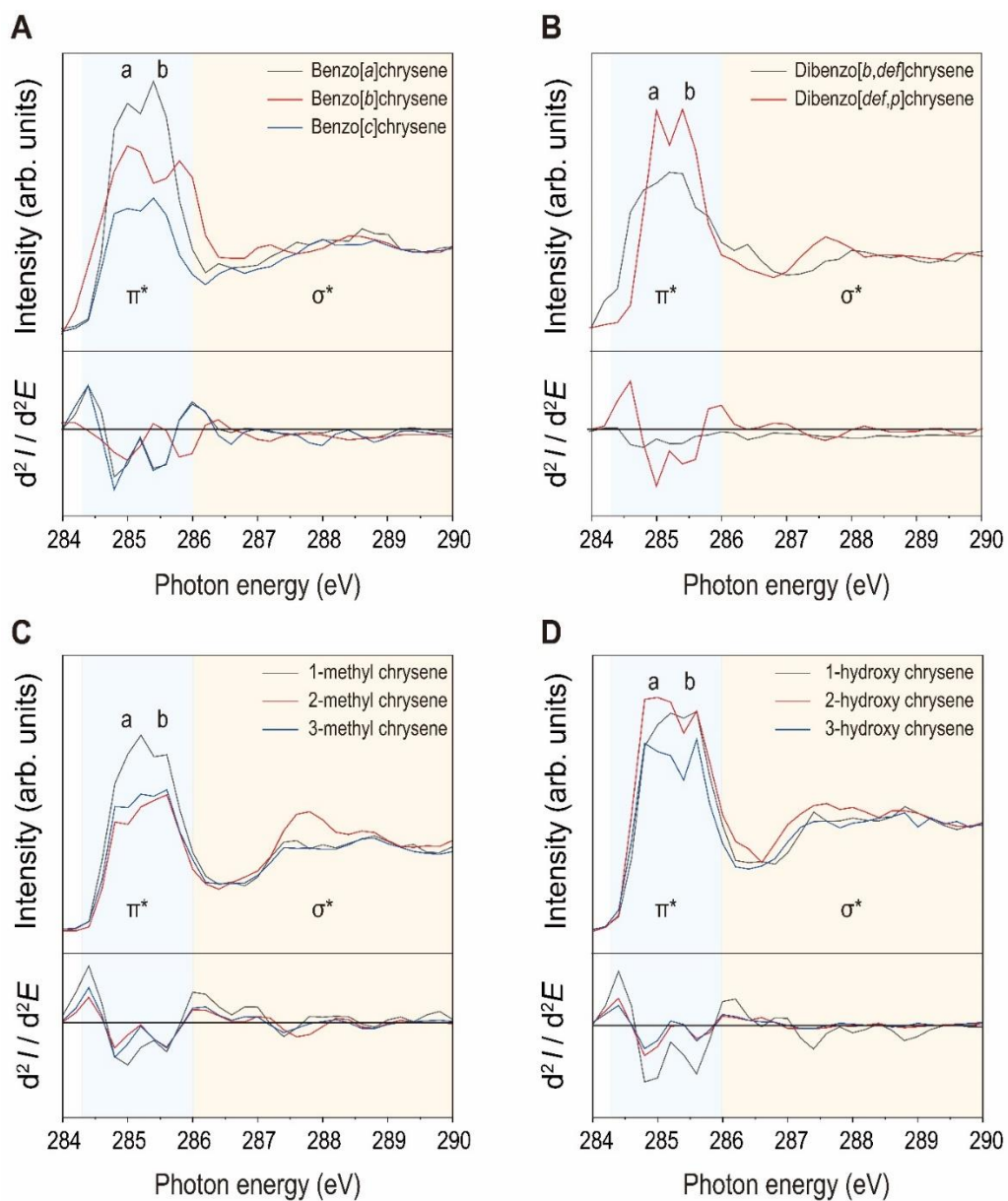


Figure S5. Carbon (1s) NEXAFS spectra of chrysene homologues of chrysene homologues. (A) benzo-chrysene, (B) dibenzo-chrysene, (C) methyl-chrysene, (D) hydroxyl-chrysene. The plotted spectra were collected from powder samples. LUMO orbitals were mainly localized around the carbon with hydrogen or other functional groups (peak a) and with neighboring carbon (peak b)

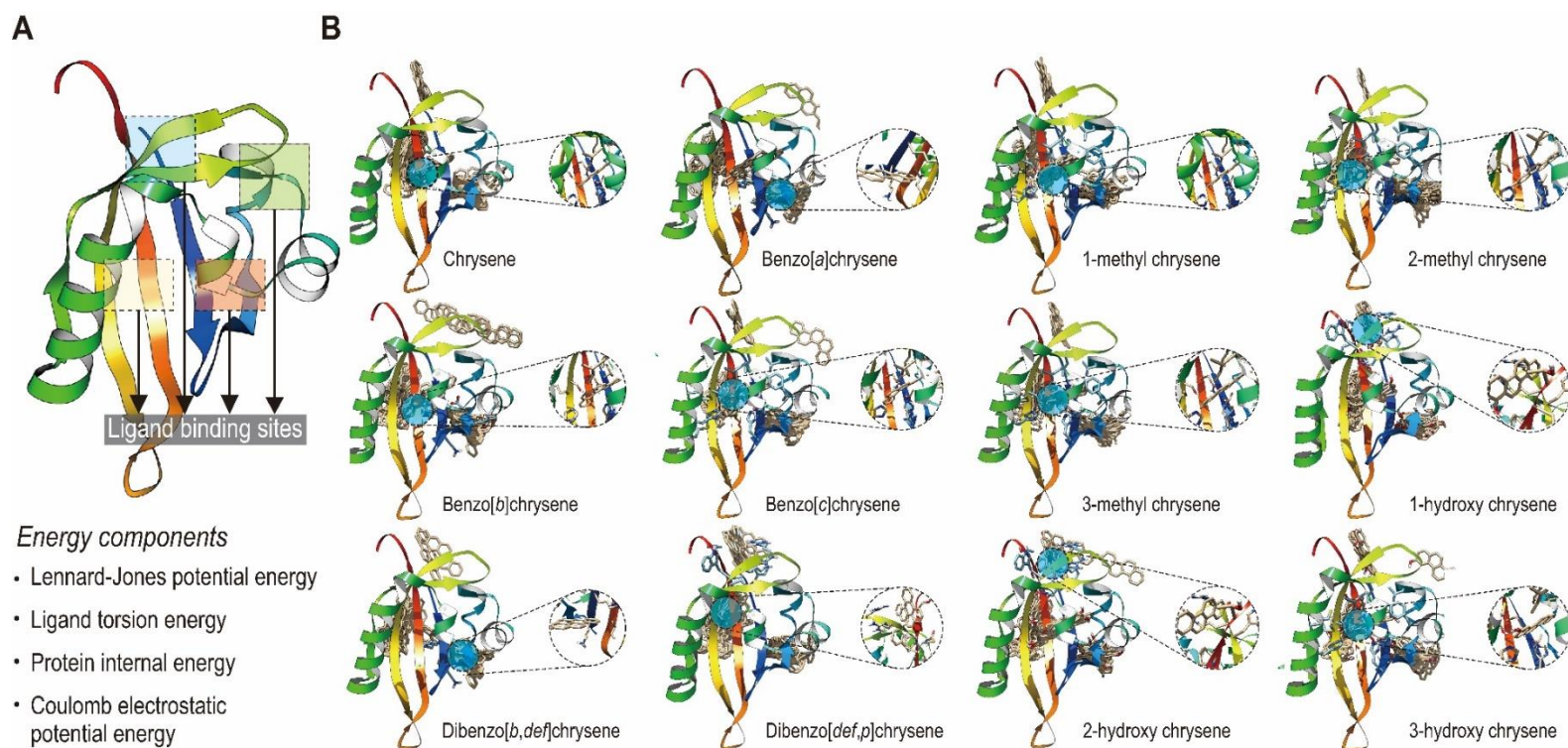


Figure S6. AhR homology and binding poses in the docking model. **(A)** The docking model (Galaxydock) used in this study calculated binding affinity by using various energy components such as Lennard-Jones potential energy, Ligand torsion energy, protein internal energy, and Coulomb electrostatic potential energy⁴; **(B)** The docking results provided fifty possible binding poses between AhR and chrysene homologues and suggests the best optimized binding state which has the lowest binding affinity.

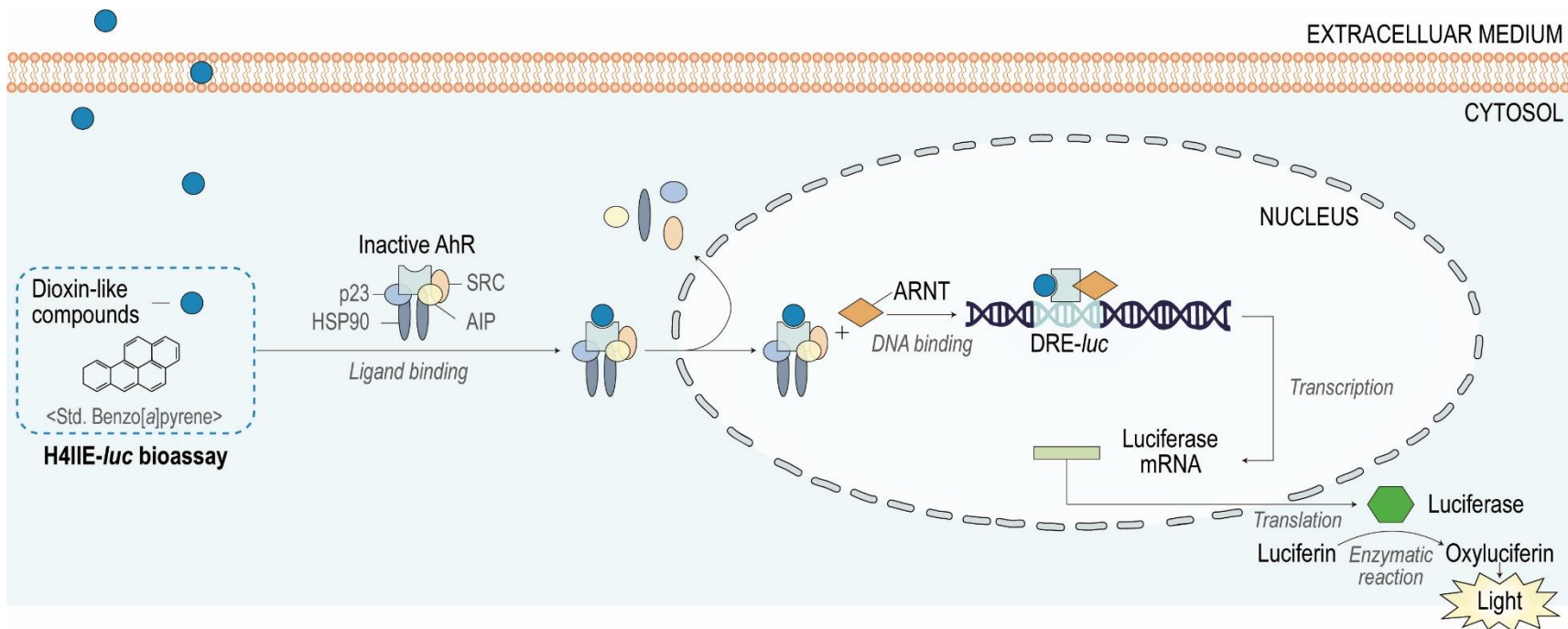


Figure S7. The molecular mechanisms of activation of gene expression by the aryl hydrocarbon receptor (AhR) mediated responses in cell. B[a]P; Benzo[a]pyrene; HSP90; 90 kDa heat shock protein, AIP AhR-interaction protein (also known as XAP2), ARNT; AhR nuclear translocator, DRE; dioxin responsive element.

Supplementary Tables

Table S1. Chemical compounds information of homologues of chrysene used in this study.

Compounds	Abb. ^{a)}	Molecular formula	Number of benzene rings	Molecular weight	Log K_{ow}	CAS RN	Purity (%)	Vendor
Chrysene	Chr	C18H12	4	228.294	5.81	218-01-9	> 98	Sigma-Aldrich (St. Louis, MO, USA)
Benzo[<i>a</i>]chrysene	BaC	C22H14	5	278.354	7.11	213-46-7	> 99	Tokyo Chemical Industry Ltd.
Benzo[<i>b</i>]chrysene	BbC	C22H14	5	278.354	7.11	214-17-5	> 98	Sigma-Aldrich (St. Louis, MO, USA)
Benzo[<i>c</i>]chrysene	BcC	C22H14	5	278.354	7.11	194-69-4	> 98	Sigma-Aldrich (St. Louis, MO, USA)
Dibenzo[<i>b,def</i>]chrysene	DbdC	C24H14	6	302.376	7.28	189-64-0	> 98	Tokyo Chemical Industry Ltd.
Dibenzo[<i>def,p</i>]chrysene	DdpC	C24H14	6	302.376	7.71	191-30-0	> 98	Sigma-Aldrich (St. Louis, MO, USA)
1-methylchrysene	1MC	C19H14	4	242.321	6.07 ^{b)}	3351-28-8	> 98	Sigma-Aldrich (St. Louis, MO, USA)
2-methylchrysene	2MC	C19H14	4	242.321	6.07 ^{b)}	3351-32-4	> 98	Sigma-Aldrich (St. Louis, MO, USA)
3-methylchrysene	3MC	C19H14	4	242.321	6.07 ^{b)}	3351-31-3	> 98	Sigma-Aldrich (St. Louis, MO, USA)
1-hydroxychrysene	1HC	C18H12O	4	244.287	5.04 ^{b)}	63019-38-5	> 98	Chiron
2-hydroxychrysene	2HC	C18H12O	4	244.287	5.04 ^{b)}	65945-06-4	> 98	Chiron
3-hydroxychrysene	3HC	C18H12O	4	244.287	5.04 ^{b)}	63019-39-6	> 98	Chiron

^{a)} Abb.: Abbreviations; ^{b)} Estimated values.

Table S2. Physico-chemical properties of aryl hydrocarbon receptors.³

Single letter code	NFIFRTKHKLDFTPIGCDAKGQLILGYTEVELCTRGSYQFIHAADILHCA ESHIRMIKTGESGMTVFRLLAKHSRWRWVQSNARLIYRNGRPDYIIATQR PLTDEE
Number of residues	107
Molecular weight	12381.1 g mol ⁻¹
Extinction coefficient	16500 M ⁻¹ cm ⁻¹
Iso-electric point	pH 9.14
Net charge at pH 7.4	4.1
Estimated solubility	Good water solubility
Equation for net charge ^{a)}	$\text{Net charge} = \sum_i N_i \frac{10^{pK_{a_i}}}{10^{pH} + 10^{pK_{a_i}}} - \sum_j N_j \frac{10^{pK_{a_j}}}{10^{pH} + 10^{pK_{a_j}}}$

^{a)} N_i are the number, and pK_{a_i} are the pK_a values, of the N-terminus and the side chains of Arginine, Lysine, and Histidine.

Table S3. *In silico* toxicity prediction models estimating the potential toxicity of chrysene homologues.

Compounds	VEGA-QSAR ^{a)}	VirtualToxlab ^{b)}	
	1/LC ₅₀	AhR binding affinity (nM)	ToxPot
Chrysene	0.95	655	0.39
Benzo[<i>a</i>]chrysene	1.13	114	0.46
Benzo[<i>b</i>]chrysene	1.13	63.2	0.48
Benzo[<i>c</i>]chrysene	1.13	133	0.45
Dibenzo[<i>b,def</i>]chrysene	1.12	78.5	0.48
Dibenzo[<i>def,p</i>]chrysene	1.12	221	0.42
1-methylchrysene	0.53	593	0.39
2-methylchrysene	0.53	450	0.41
3-methylchrysene	0.53	501	0.40
1-hydroxychrysene	0.43	352	0.41
2-hydroxychrysene	0.43	466	0.41
3-hydroxychrysene	0.43	250	0.43

^{a)} VEGA-QSAR estimated data (fish); ^{b)} potential toxicity and estimated affinity of binding to AhR.

Table S4. *In vitro* transactivation bioassay conditions for evaluating toxicities of chemical compounds.

Cell line	H4IIE- <i>luc</i>
ATCC#	CRL-1548
Cell type	Recombinant (<i>luc</i> -gene)
Mode of Action	AhR-mediated potency
Endpoint	Luciferase activity
Positive control	Benzo[<i>a</i>]pyrene
Culture condition	37 °C, 5% CO ₂
Exposure time	4 h

References

- (1) Waterhouse, A. *et al.* SWISS-MODEL: homology modelling of protein structures and complexes. *Nucleic Acids Res.* **46**, W296–W303 (2018).
- (2) Wiederstein, M. & Sippl, M. J. ProSA-web: interactive web service for the recognition of errors in three-dimensional structures of proteins. *Nucleic Acids Res.* **35**, W407–W410 (2007)
- (3) Lear, S. & Cobb, S. L. Pep-Calc. com: a set of web utilities for the calculation of peptide and peptoid properties and automatic mass spectral peak assignment. *J. Comput. Aided. Mol. Des.* **30**, 271–277 (2016).
- (4) Shin, W.-H. *et al.* Prediction of protein structure and interaction by GALAXY protein modeling programs. *Bio. Design.* **2**, 1–11 (2014).
- (5) Vedani, A. *et al.* OpenVirtualToxLab—a platform for generating and exchanging in silico toxicity data. *Toxicol. Lett.* **232**, 519–532 (2015).